FINAL

Risk-Based Corrective Action Plan Hazardous Waste Storage Area (Building 560)



Rickenbacker Air National Guard Base Columbus, Ohio

Volume I: Report and Appendices

Prepared For

Air Force Center for Environmental Excellence
Technology Transfer Division
Brooks Air Force Base, Texas
San Antonio, Texas

and

Air Force Base Conversion Agency Operating Location - Rickenbacker Columbus, Ohio

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CORRECTIVE ACTION PLAN HAZARDOUS WASTE STORAGE AREA BUILDING 560

at

RICKENBACKER AIR NATIONAL GUARD BASE COLUMBUS, OHIO

June 1999

Prepared for:

AIR FORCE CENTER FOR ENVIRONMENTAL EXCELLENCE TECHNOLOGY TRANSFER DIVISION BROOKS AIR FORCE BASE SAN ANTONIO, TEXAS

AND

AIR FORCE BASE CONVERSION AGENCY OPERATING LOCATION - RICKENBACKER COLUMBUS, OHIO

Prepared by:

Parsons Engineering Science, Inc. 1700 Broadway, Suite 900 Denver, Colorado 80290

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ACRONYMS AND ABBREVIATIONS

AFBCA Air Force Base Conversion Agency

AFCEE Air Force Center for Environmental Excellence

AMC Air Mobility Command
ANGB Air National Guard Base
bgs below ground surface
BCT Base closure team

BTEX Benzene, toluene, ethylbenzene, and xylenes

BUSTR Ohio Department of Commerce, Division of State Fire Marshal,

Bureau of Underground Storage Tank Regulation

CAH Chlorinated aliphatic hydrocarbon CFR Code of Federal Regulations

cm² Square centimeter

CME Comprehensive Groundwater Monitoring Evaluation

COPC Contaminants of potential concern

CPT Cone penetrometer

CSF Oral carcinogenic slope factor

CSM Conceptual site model

DCA Dichloroethane
DCE Dichloroethene

DERR Ohio EPA, Division of Emergency Remedial Response DHWM Ohio EPA, Division of Hazardous Waste Management

DO Dissolved oxygen
DOD Department of Defense

DPDO Defense Property Disposal Office

DRMO Defense Reutilization and Materials Office

DRO Diesel-range organics
ES Engineering-Science, Inc.

ET Exposure time
°F Degrees Fahrenheit

FAA Federal Aviation Administration

Fe³⁺ ferric iron
Fe²⁺ ferrous iron
FR Federal Register
ft/day Feet per day
ft/ft Foot per foot
ft/vr Feet per year

GRO Gasoline-range organics

HEAST Health Effects Assessment Summary Tables

HI Hazard index

HPLC High-performance liquid chromatrography

HO Hazard quotient

HWSA Hazardous Waste Storage Area

ID Inside diameter in/hr Inches per hour

IRIS Integrated Risk Information Service
IRP Installation Restoration Program
IT International Technology Corporation

IUR Inhalation unit risk factor
LIF Laser-induced fluorescence
LNAPL Light nonaqueous-phase liquid
MCL Maximum contaminant level
MCLG Maximum contaminant level goal

 $\begin{array}{ll} \text{MDL} & \text{Method detection limit} \\ \mu g/kg & \text{micrograms per kilogram} \\ \mu g/L & \text{Micrograms per liter} \\ mg/kg & \text{milligrams per kilogram} \end{array}$

mg/kg-day Milligrams per kilogram per day

mg/L milligrams per liter
MP Monitoring point
msl Mean sea level

NFRAP No Further Response Action Plan NOAEL No-observed-adverse-effect level

NRMRL National Risk Management Research Laboratory

OAC Ohio Administrative Code ORC® Oxygen Release Compound

Ogden Environmental and Energy Services, Inc.

OH hydroxyl

Ohio EPA Ohio Environmental Protection Agency

Parsons ES Parsons Engineering Science, Inc.

PCB Polychlorinated biphenyl
PEF Particulate emission factor
PID photoionization detector
PVC Polyvinyl chloride

RAGS Risk Assessment Guidance for Superfund

RBCA Risk-based closure action

RCRA Resource Conservation and Recovery Act

redox Oxidation/reduction

RfC Inhalation reference concentration

RfD Oral reference dose

RME Reasonable maximum exposure
RNA Remediation by natural attenuation
RDA Rickenbacker Port Authority

RPA Rickenbacker Port Authority
SSTL Site-specific target level

SVOC Semivolatile organic compound

SW Sparge well
TAL Target analyte list
1,1,1-TCA 1,1,1-Trichloroethane
TCE Trichloroethene
TMB Trimethylbenzene

TMB Trimethylbenzene
TOC Top of casing

TPH Total petroleum hydrocarbons

UCL Upper confidence limit

USACE US Army Corps of Engineers

USEPA US Environmental Protection Agency

USGS US Geological Survey
UST Underground storage tank

UWBZ	Upper water-bearing zone
VAP	Voluntary Action Program
VC	Vinyl chloride
VF	volatilization factor
VOC	Volatile organic compound
VW	Vent well

EXECUTIVE SUMMARY

This risk-based corrective action plan includes conservative estimates of risk, conclusions, and recommended corrective actions for the former hazardous waste storage area (HWSA) at Building 560, Rickenbacker Air National Guard Base (ANGB), Ohio. The former HWSA is located at the edge of the shop area in the northern section of the Base. The HWSA was a permitted Resource Conservation and Recovery Act (RCRA) storage facility that received wastes generated during Base activities from 1983 through 1986. Wastes stored at the HWSA consisted primarily of acids, various solvents and cleaning solutions, petroleum products, and spent desiccants. Waste was stored in small containers placed on pallets inside Building 560 and in larger, 55-gallon drums outside the building (within the fenced yard). Additionally, 15 underground storage tanks (USTs) were previously located at or near the HWSA. These USTs were used for the storage of fuel waste, waste petroleum products, and deicing fluid. The Air Force Base Conversion Agency (AFBCA) has been granted No Further Response Action Planned (NFRAP) status for the tanks. Residual contamination in soil and groundwater is addressed as part of the risk-based corrective action presented in this This corrective action plan was prepared in compliance with the Ohio Administrative Code (OAC) Chapter 3745 Part 66 and the Code of Federal Regulations (CFR), Title 40 Part 265, Subpart G.

The risk-based corrective action presented in this plan is intended to provide a risk-based alternative to the closure approach presented in the February 1997 amended closure/post-closure plan (Parsons Engineering Science, Inc. [Parsons ES], 1997a) which was approved with modifications by the Ohio Environmental Protection Agency (Ohio EPA) on May 19, 1997. The February 1997 plan (Parsons ES, 1997a) proposed closure of the HWSA utilizing natural chemical attenuation processes and additional untested source area remedial technologies (i.e., bioventing and air sparging) to minimize contaminant mass, mobility, persistence, and toxicity. Because no remediation standards or risk-based health standards were presented in the February

1997 closure plan, Ohio EPA gave approval for closure of the site as a landfill, with all post-closure care requirements being applicable. The closure period was extended by Ohio EPA to October 2000 to ensure that the final closure strategy is compatible with the planned future use of the HWSA as a paved taxiway.

In compliance with the February 1997 amended closure/post-closure plan, additional site assessment activities were initiated by Parsons ES and International Technology The additional site assessment activities to be performed were Corporation (IT). documented in a June 1997 work plan (Parsons ES, 1997b). The primary purpose of these activities was to collect sufficient site-specific data to develop risk-based health standards for the HWSA through a site-specific risk assessment, and to evaluate the effectiveness of proposed source-reduction technologies (bioventing and air sparging), in the event risk reduction through source area remediation is required under the proposed risk-based closure. The results of the 1997 site assessment activities and the risk assessment were presented in a Draft Final Amended Closure Plan submitted to the Ohio EPA in February, 1998. A meeting was held on February, 1999 to discuss a draft Notice of Deficiency from Ohio EPA concerning the February, 1998 Amended Closure Plan. Based on Ohio EPA comments provided during the February, 1998 meeting and their intent to deliver a notice of deficiency (see Appendix I for meeting minutes and Ohio EPA draft notice of deficiency), the February 1998 Draft Final Amended Closure Plan was withdrawn on March 15, 1999 (see Appendix I for March 15, 1999 withdrawal letter). The primary deficiency Ohio EPA identified in the 2/98 Draft Final Amended Closure Plan was that site-specific target levels cannot exceed groundwater maximum contaminant levels (MCLs) per a 6/16/98 Interoffice Communication from Jeff Parzke to Ed Lim (Appendix I) even if groundwater is not a current or reasonably expected future source of drinking water. This change in Ohio EPA policy was identified to Parsons ES and the Air Force one year after the February 1998 Amended Closure Plan had been submitted.

In light of this development, the February 1997 amended closure/post-closure plan (Parsons Engineering Science, Inc. [Parsons ES], 1997a) will remain as the approved

closure plan (i.e., closure as a landfill). The purpose of this Risk-Based Corrective Action Plan is to provide a conservative risk-based assessment that demonstrates that current contaminant concentrations in soil and groundwater do not exceed carcinogenic and noncarcinogenic target levels consistent with Ohio EPA guidelines for industrial sites where groundwater is not considered a potential drinking water source. This plan provides conclusions and risk-based corrective action recommendations which may serve as an alternative to the closure approach presented in the February 1997 amended closure/post-closure plan (Parsons Engineering Science, Inc. [Parsons ES], 1997a), assuming groundwater is determined to not be a potential drinking water source.

Results from the risk assessment demonstrated that existing concentrations of organic and inorganic compounds in soil and groundwater do not pose unacceptable carcinogenic risks or noncancer hazards consistent with Ohio EPA target goals under current or future receptor exposure scenarios for industrial sites where groundwater is not potable. Considering these findings, a risk assessment-based corrective action plan is proposed for the HWSA. The following activities have been completed or are proposed to be completed as discussed in this risk-based corrective action plan:

- Removal of the remaining four USTs (completed February 1995);
- Decontamination of Building 560 by cleaning the building and the drum wash pad (completed April 1996);
- Installation of ten new monitoring wells in the vicinity of the HWSA and comprehensive groundwater sampling from 37 site monitoring wells (completed July 1997);
- Sampling of soil and soil gas in the vicinity of Building 560 (completed August 1997);
- Treatability testing for determination of the effectiveness of air injection bioventing, in the event residual organic soil contamination required limited in

situ remediation based on the findings of the risk assessment (completed August 1997);

- Treatability testing for determination of the effectiveness of air sparging in the
 event site impacted groundwater required in situ remediation based on the
 findings of the risk assessment (completed December 1997);
- Assessment of risks/hazards associated with chemicals of potential concern (COPCs) at the HWSA (completed within this plan);
- Natural oxidation of residual dissolved fuel hydrocarbons and natural reductive dehalogenation of residual dissolved chlorinated organics (in progress);
- Installation of one additional downgradient monitoring well cluster in order to completely delineate the extent of the chlorinated organics plume (proposed);
- Continued monitoring of the groundwater COPCs identified in the risk assessment (proposed);
- Institution of land-use controls through a deed restriction upon transfer of the HWSA property (proposed); and
- Eventual exposure control by installation of taxiway (proposed as future land use).

With the exception of the taxiway installation, all risk-based corrective activities, as outlined above, will be complete by October 2000. No corrective activities will be necessary based upon the results of the risk assessment presented in this plan.

SECTION 1

FACILITY DESCRIPTION

1.1 INTRODUCTION

This corrective action plan was prepared in accordance with the Ohio Administrative Code (OAC) Chapter 3745 Part 66, and Code of Federal Regulations (CFR) Title 40 Part 265, Subpart G. This corrective action plan provides a risk-based alternative to the closure approach presented in the February 1997 approved amended closure/post-closure plan (Parsons ES, 1997a). As part of the Installation Restoration Program (IRP), AFBCA has identified the former hazardous waste storage area (HWSA) at Building 560, located at Rickenbacker Air National Guard Base (ANGB), for closure. This corrective action plan was prepared by Parsons Engineering Science, Inc. (Parsons ES) under contract with the Air Force Air Mobility Command (AMC) and the Air Force Center for Environmental Excellence (AFCEE). Existing site characterization data and data collected by International Technology Corporation (IT) (under a separate contract with AFCEE) during 1997 assessment activities were used to prepare a risk-based corrective action recommendation for soil and groundwater at the HWSA.

1.1.1 Corrective Action Plan Objective

The purpose of this corrective action plan is to present the results of a site-specific risk assessment conducted for the HSWA, and to provide risk-based recommendations which may serve as an alternative to the closure approach presented in the February 1997 approved amended closure/post-closure plan (Parsons ES, 1997a). Potential impacts to human health and the environment under realistic current and future land use and receptor exposure scenarios have been evaluated in this report. If the results from the risk assessment demonstrate that environmental contaminants in soil and groundwater exceed Ohio EPA target goals for industrial sites where groundwater is

not used as drinking water, then the results of additional assessment activities performed at the HWSA in 1997 will be used to optimize/guide the remedial approach. If it is determined that site contaminant concentrations do not result in an exceedance of Ohio EPA risk-based target goals, then a risk-based closure could be considered for the site following sufficient confirmation monitoring.

This corrective action plan is intended to provide a risk-based alternative to the February 1997, Ohio EPA-approved amended closure/post-closure plan (Parsons ES, 1997a). In lieu of risk-based cleanup goals, the February 1997 amended closure/post-closure plan was approved as a landfill closure with all post-closure care requirements as specified in the state Resource Conservation and Recovery Act (RCRA) facility closure guidance (Ohio EPA 1993a). The risk assessment for affected media at the HWSA was conducted in accordance with Ohio EPA (1993b) risk-based closure action (RBCA) guidance.

1.1.2 Regulatory Framework

1.1.2.1 Site Background

A notification of hazardous waste activity and a RCRA Part A (Interim Status) application was submitted to the US Environmental Protection Agency (USEPA) by Rickenbacker ANGB on March 22, 1983. The site was granted USEPA Interim Status Permit #OH3571924544, but the application was not processed for Ohio, and a Part B permit application was not submitted (Ohio EPA, 1995a). The HWSA was used from 1983 to 1986 for the storage of drummed hazardous wastes generated at the Base. Additionally, 15 underground storage tanks (USTs) used for the storage of fuel, waste petroleum products, and deicing fluid were formerly located at, or within the vicinity of, Building 560/HWSA.

The following information was summarized from the Comprehensive Groundwater Monitoring Evaluation (CME) conducted by Ohio EPA (1995a). A closure plan for Building 560 was submitted to the Ohio EPA on April 15, 1987. The plan was rejected on April 6, 1988, and a revised closure plan was submitted on July 27, 1988. This

plan was approved by Ohio EPA on September 29, 1988. On October 26, 1990, Rickenbacker ANGB informed the Ohio EPA that a "clean closure" in accordance with the closure plan could not be achieved. Ohio EPA advised the Base to continue under the original plan until the vertical and horizontal extent of contamination could be determined.

On May 26, 1992, Rickenbacker ANGB submitted an amended closure/post-closure plan to close the HWSA as a landfill with engineered groundwater remediation to occur during the post-closure care period. Following Ohio EPA's issuance of a Notice of Deficiency on February 9, 1993, a revised closure/post-closure plan (Engineering Science, Inc. [ES], 1993) was submitted on March 15, 1993, and was approved by Ohio EPA on July 13, 1993. This plan proposed closure of the HWSA as a landfill with groundwater remediation via extraction and treatment. The closure plan also proposed completion of closure without a protective cap due to concerns that a cap would interfere with the groundwater extraction and treatment system.

On September 30, 1994, responsibility for closure of the HWSA was transferred from Rickenbacker ANGB to AFBCA. On January 27, 1995, AFBCA informed Ohio EPA that the effectiveness of the proposed groundwater extraction and treatment system was being questioned, and requested an extension of the closure period to allow for an assessment of the effectiveness of natural attenuation processes at remediating dissolved contamination at the HWSA. The extension was approved by Ohio EPA on March 21, 1995.

The findings of the AFCEE-funded remediation by natural attenuation (RNA) demonstration at the HWSA and revised closure recommendations were documented in an October 1995 amended closure/post-closure plan (Parsons ES, 1995a). On August 26, 1996, after review of the October 1995 plan, Ohio EPA informed AFBCA that reliance on natural chemical attenuation processes as the sole remedial action for closure would not be approved. Additionally, Ohio EPA noted that risk-based performance objectives should be established to support any closure recommendations.

In response to this technical direction, the AFBCA submitted a revised amended closure/post-closure plan that proposed various source reduction technologies (bioventing, air sparging/oxygen release compound [ORC®] injection) (Parsons ES, 1996). This plan was submitted to Ohio EPA on October 11, 1996. Following Ohio EPA's issuance of a Notice of Deficiency on January 23, 1997, the plan was modified to correct deficiencies and resubmitted by AFBCA on February 28, 1997 (Parsons ES, 1997a). This version of the amended closure/post-closure was released for public comment the week of March 17, 1997. Notification of formal approval of the plan with modifications by Ohio EPA was received on May 19, 1997. Ohio EPA-required modifications to the approved February 1997 closure plan were submitted to Ohio EPA
Closure plan modifications significant to future site activities include:

- Closure approval of the HWSA as a landfill with all post-closure care requirements being applicable unless it can be demonstrated that risk-based health standards can be achieved at the end of the closure period; and
- Extension of the closure period until October 2000 to allow negotiations with the Rickenbacker Port Authority (RPA) to ensure that the final closure strategy is compatible with the planned future use of the HWSA.

In compliance with the February 1997 amended closure/post-closure plan, additional site assessment activities were initiated by Parsons ES and IT. The additional site assessment activities to be performed were documented in a June 1997 work plan (Parsons ES, 1997b). The primary purpose of these activities was to: 1) collect sufficient site-specific data to conduct a risk assessment; 2) if necessary, develop risk-based health standards for the HWSA; and 3) evaluate the effectiveness of proposed source-reduction technologies (bioventing and air sparging). Source reduction technologies were tested in the event risk reduction through source area remediation was required to more rapidly attain risk-based health standards for the site.

This corrective action plan presents the environmental data collected from the HWSA and surrounding area during the 1997 assessment activities, summarizes the results of bioventing and air sparging treatability testing, presents a site-specific risk assessment for contaminated soil and groundwater at the HWSA, and discusses the need for risk-based health standards in the form of site-specific target levels (SSTLs) for comparison and confirmation. Closure of the HWSA with no post-closure care is proposed as an alternative to the February 1997, Ohio EPA-approved amended closure/post-closure plan (Parsons ES, 1997a) following eight quarters of compliance groundwater monitoring and establishment of land use controls through restrictions on the deed for the property.

1.1.2.2 Regulatory Authority

The 15 USTs formerly located at, or the vicinity of, the HWSA were removed in 1994 and 1995. AFBCA received a "No Further Response Action Planned" (NFRAP) status from the Ohio Department of Commerce, Division of State Fire Marshal, Bureau of Underground Storage Tank Regulation (BUSTR) in March 1996. The NFRAP for four of the USTs was granted for tanks 47 through 50 under the condition that residual contamination in soils and groundwater be addressed in any forthcoming closure plans for the HWSA.

Closure of the HWSA is being conducted in accordance with Ohio EPA standards for RCRA hazardous waste interim facilities (OAC 3745, Chapters 65 through 69). The closure performance standard for interim facilities (OAC 3745-66-11) requires that closure:

- Minimize the need for further maintenance;
- Control, minimize, or eliminate to the extent necessary to protect human health and the environment, post-closure escape of hazardous waste/hazardous constituents to groundwater, surface waters, or to the atmosphere; and
- Comply with various other OAC rules regarding closure of interim facilities.

Rules OAC 3745-66-10 to 3745-66-15 outline closure requirements, and OAC 3745-66-16 through 3745-66-20 specify post-closure care requirements applicable to all interim hazardous waste facilities. After closure plan approval by Ohio EPA, partial and/or final closure activities typically must be completed within 90 days; however, an extension can be requested from Ohio EPA as long as adequate steps are taken to prevent threats to human health and the environment (OAC 3745-66-13). An extended closure period until October 2000 was requested and approved by Ohio EPA for the February 1997 amended closure/post-closure plan (Parsons ES, 1997a). February 1997 plan (Parsons ES, 1997a), post-closure care based on a landfill closure is to begin after completion of closure activities and to continue for 30 years. However, the facility owner/operator can amend the closure plan at any time prior to notifying Ohio EPA of partial or final closure (OAC 3745-66-12), and the post-closure care period can be reduced if human health and the environment are sufficiently protected (OAC 3745-66-17). As with facility closure, the owner/operator can amend the post-closure plan any time during the post closure care period (OAC 3745-66-18). In addition, the owner/operator may petition the Ohio EPA to reduce the post-closure care period based on evidence demonstrating that the secure nature of the hazardous waste facility makes post-closure care requirements unnecessary or supports reduction of the post-closure care period (OAC 3745-66-18(G)). Applicable evidence indicating the secure nature of the facility includes leachate or groundwater monitoring results and application of advanced technologies.

1.1.2.3 Ohio EPA Risk-Based Closure Guidance

Authorization to consider risk assessment, or what constitutes "decontamination" of a site, as a closure option is provided by rule OAC 3745-67-28, which adopts federal regulations 51 FR 16422 and 52 FR 8704, promulgated on May 2, 1986 and March 19, 1987, respectively (Ohio EPA, 1993a). Currently applicable Ohio EPA (1993b) guidance is provided in the *Interim Final Guidance for Reviewing Risk-Based Closure Plans for RCRA Units*. This outlines two options for closing a RCRA unit when no further RCRA obligations are required for managing or monitoring the unit:

- Option 1 Complete removal of waste materials and contaminated soils; or
- Option 2 Complete removal of waste materials and decontamination of environmental media (soil, water, and air) to health-based standards.

For the second option, demonstration of acceptable decontamination is performed through a human health risk assessment for all impacted and potentially impacted environmental media environmental media to be left in place.

Site-specific demonstrations of protection of human health and the environment are required for a risk-based closure (Ohio EPA, 1993b). The owner/operator can propose site-specific, risk-based cleanup and closure strategies that address all routes of exposure and unacceptable risk. Site-specific, risk-based proposals must document that contaminants left in soil will not adversely impact environmental media (groundwater, surface water, or the atmosphere) and demonstrate that direct exposure of receptors through dermal contact, ingestion, or inhalation will not result in a threat to human health or the environment. Also, Ohio EPA (1993b) may consider an industrial exposure scenario if deed restrictions are placed on the property or if a permanent structure is placed directly over the unit.

Ohio EPA is in the process of revising current RBCA policy and procedures on various risk-based guidance issues and developing guidance consistent with DHWM and Division of Emergency Remedial Response (DERR) regulatory objectives. Included within this revision process is the development of cleanup criteria that reflect an acceptable carcinogenic target risk range (10⁻⁴ to 10⁻⁶). Ohio EPA (1995b) has identified a 10⁻⁵ cumulative carcinogenic risk goal as acceptable to ensure protection of human health within the Voluntary Action Program (VAP) for DERR sites. For consistency, this target risk goal also has been adopted by other Ohio EPA programs, including DHWM programs (Ohio EPA, 1997). In addition, Ohio EPA is currently in the process of developing deed restriction guidance for site closures.

1.2 REPORT ORGANIZATION

This corrective action plan consists of seven sections, including this introduction, and 10 appendices. A general facility background, including an overview of the environmental setting, is provided in the remainder of this section. Section 2 presents more detailed information on the HWSA history, a summary of the wastes that were stored at the site, and a discussion of the site's physical setting, and describes the current and future land uses for the site. Section 3 summarizes previous investigations and remedial activities performed at the site. Section 4 presents the results of the 1997 additional assessment activities performed at the HWSA and summarizes the nature and extent of contamination in soil and groundwater at the site. Section 5 represents a comprehensive risk assessment for the site and discusses risk-based health standards for use in compliance monitoring. Section 6 explains the proposed corrective action activities to be implemented and includes a proposed schedule. Section 7 presents references used in preparing this corrective action plan.

Appendix A presents available future land use documentation for the site, including example deed restriction language. Appendix B presents previous site analytical data collected from 1988 through 1996. Appendix C presents analytical data associated with the 1997 additional assessment activities. Appendix D presents available soil boring logs and well installation diagrams. Appendix E provides background data for soils and groundwater and provides text and summary tables related to the statistical analysis of site contaminants. Appendix F provides supporting risk calculations and related risk assessment information. Appendix G presents the sampling and analytical procedures for future compliance monitoring to be conducted at the site, Appendix H presents a health and safety plan for use during future site corrective action activities, and Appendix I contains copies of the March 15, 1999 letter from AFBCA/DB Rickenbacker to Ohio EPA requesting withdrawal of the February 1998 Draft Final Amended Closure Plan, minutes of the February 8, 1999 meeting with Ohio EPA, the Ohio EPA Draft Notice of Deficiency comments on the February 1998 Draft Final Amended Closure Plan, the June 16, 1998 Ohio EPA interoffice communication which

discusses the use of MCLs when risk-based numbers exceed MCLs, and Rickenbacker Port Authority comments on the February 1998 Draft Final Amended Closure Plan.

1.3 ENVIRONMENTAL SETTING OF RICKENBACKER AIR NATIONAL GUARD BASE

Rickenbacker ANGB is located 12 miles southeast of Columbus, and 0.5 mile east of Lockbourne, Ohio (Figure 1.1). The Base covers approximately 2,100 acres in Franklin and Pickaway Counties and is located on a glacial till plain between the Big Walnut and Walnut Creek drainage basins. The area has been used as an air base under the custody of various government branches, including the Army Air Corps and the Air National Guard, since 1942. Access to the Base is restricted through a continuously guarded entrance.

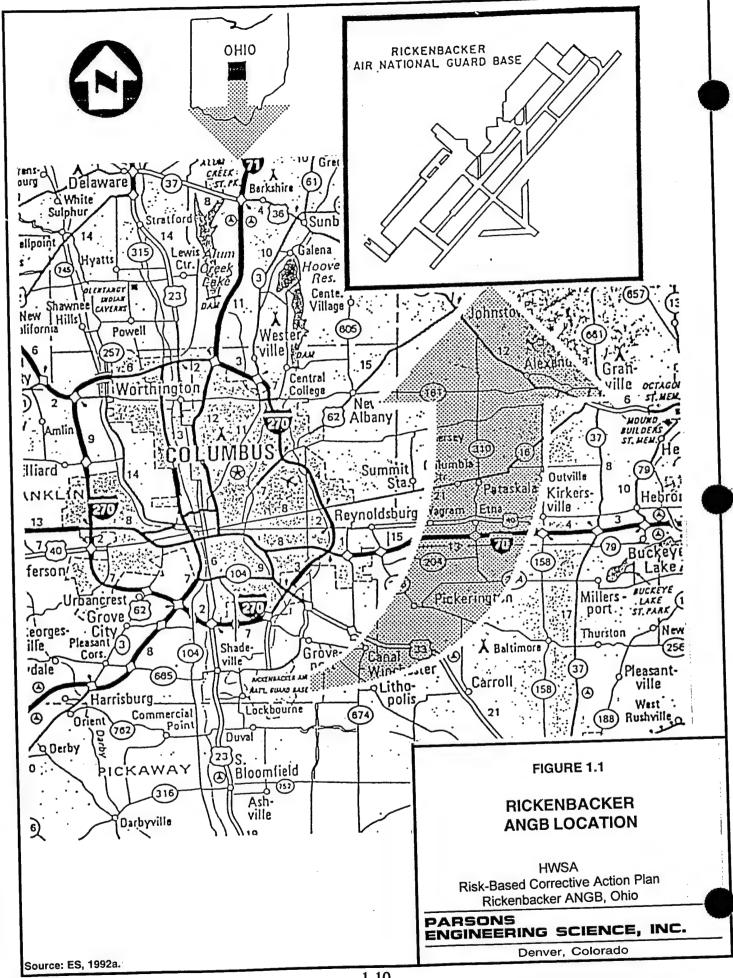
The environmental setting of the Base is described in the following subsections, with an emphasis on the identification of natural features that may influence the migration of hazardous-waste-related contaminants from the HWSA at Building 560.

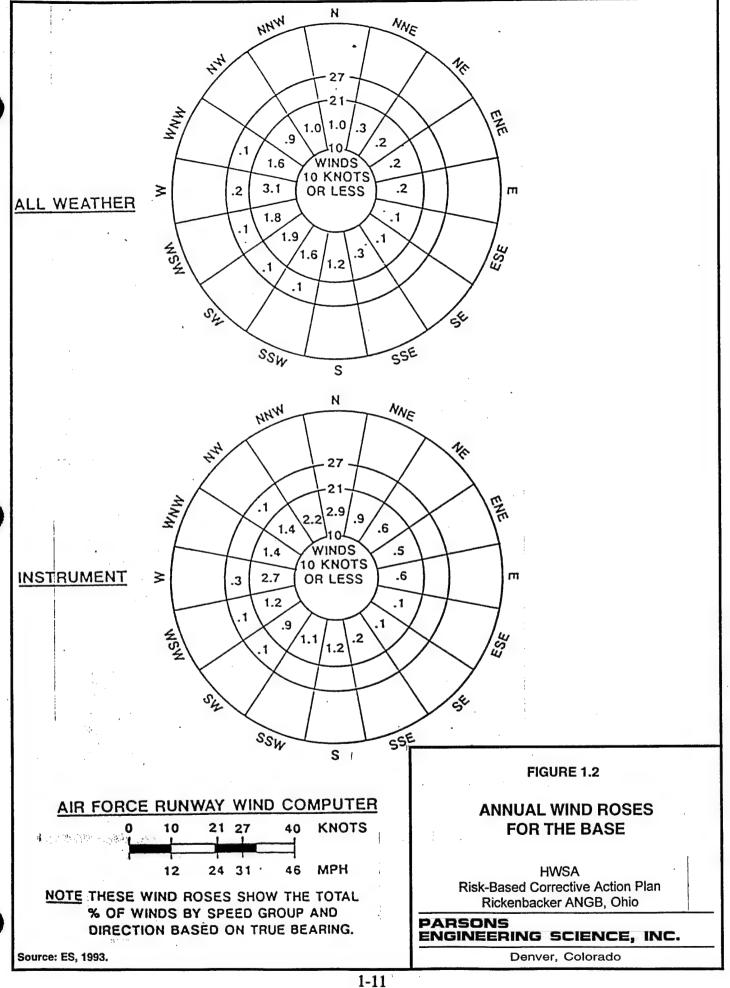
1.3.1 Meteorology

The climate of Columbus, Ohio is continental, characterized by cold winters, hot summers, and moderate rainfall (Pierce, 1959). The mean annual temperature is 52 degrees Fahrenheit (°F). The coldest month is January, with a mean temperature of 30°F; the warmest month is July, with a mean temperature of 74°F. Precipitation at the Base falls primarily during the summer months, with June being the wettest month, and October being the driest month. The mean annual precipitation at the Base is 38 inches. The prevailing wind directions on the Base are from the southwest to northnorthwest, as illustrated on Figure 1.2.

1.3.2 Regional Geology

The Base is located in the Glaciated Central Lowlands Province, just west of the Appalachian Plateau Province. The geology of the area is characterized by up to 200 feet of Pleistocene sandy and gravelly glacial outwash and silty and clayey glacial till filling a preglacial bedrock valley (Schmidt and Goldthwaite, 1958). Bedrock





underlying the valley fill consists of Devonian-age limestones and shales of the Columbus and Delaware Formations.

The surficial geology at Rickenbacker ANGB is composed of two distinct glacial till deposits which overly glacial outwash deposits (IT, 1996a). The uppermost till, a Wisconsin-age ablation till, is composed primarily of a silty clay and a clayey silt with varying amounts of sand and gravel. The upper till unit grades in color from brown to gray with depth. The sand and gravel within this unit occur as isolated lenses that range in thickness from 3 to 30 feet (IT, 1996a). The second underlying till unit, a Wisconsin-age basal till, consists of gray silt with clay and contains varying amounts of fine sand and pebble gravel. This gray till is generally at least 10 feet thick (IT, 1996a). At the contact between the upper, brown to gray till and the lower gray till occur laterally discontinuous sand and gravel lenses.

1.3.3 Regional Soils

Soils mapped at the Base are of the Kokomo and Crosby Series (Soil Conservation Service, 1976). The soils are characterized as deep, very poorly drained, slowly to moderately slowly permeable soils formed in glacial tills on uplands. The Crosby series soils are formed on slopes with up to 6-percent grade, while the Kokomo series soils form on gentler 0- to 2-percent slopes on the higher landscape positions. The Crosby soils exhibit permeabilities of 0.06 to 0.6 inch per hour (in/hr) in unleached horizons. The Kokomo soils have permeabilities of 0.2 to 2.0 in/hr.

1.3.4 Regional Surface Water Hydrology

Rickenbacker ANGB occupies the drainage divide between Big Walnut Creek and Walnut Creek. Surface drainage from the Base is controlled through an extensive storm drain network, which includes corrugated metal and concrete drainage pipes and open, unlined drainage ditches. All of the surface runoff is routed through emergency interceptors before being released into surrounding surface streams, which ultimately discharge into Walnut Creek and Big Walnut Creek. Walnut Creek is the permanent stream nearest the HWSA and is located approximately 1.5 miles east of the site.

1.3.5 Regional Hydrology and Groundwater Use

The regional hydrologic regime is composed of three distinct water-bearing zones: an upper water-bearing zone (UWBZ), an intermediate aquifer, and a deep aquifer (IT, 1996a). The UWBZ occurs primarily in the laterally discontinuous sand and gravel lenses located at the contact between the two till units (Section 1.3.2). The heterogeneity of the UWBZ deposits creates localized unconfined/semiconfined hydraulic regimes with minimal hydraulic interconnection. The gray till, underlying the upper till and associated sand and gravel lenses, acts as an impermeable, lower boundary of the UWBZ. The UWBZ is hydraulically distinct and separate from the intermediate and deep aquifers (IT, 1995a).

Regionally, the intermediate and deep aquifers represent the primary potable water supply. Rickenbacker ANGB and RPA facilities receive water from the City of Columbus. In the nearby village of Lockbourne, most residents are supplied by the City of Columbus or have private wells screened in the intermediate aquifer (well depths range between 48 and 110 feet bgs). Five homeowners in Lockbourne are known to obtain their water supply from shallow, low-yielding wells ranging in depth from 20 to 30 feet bgs and assumed to be in the UWBZ (IT, 1996a). The UWBZ is not likely to be used in the future as a water supply due to the proximity of the much higher yielding intermediate and deep aquifers and historical problems in Lockbourne water, with bacteria and sewage disposal impacting shallow wells in the UWBZ (IT, 1996a).

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SECTION 2

SITE DESCRIPTION

2.1 SITE BACKGROUND

The HWSA is located in the central area of the Base (Figure 2.1). The area where the HWSA was constructed had various other uses in the past. Historical aerial photographs, maps, and drawings indicate that the site was used as a storage yard, probably for drummed lube oils. When the Base was constructed in 1942, individual buildings were heated with coal. The coal storage area for the Base was located west of the HWSA, adjacent to the railroad tracks. The smokestack for a coal-burning furnace is still standing approximately 180 feet from the HWSA.

The HWSA measures 170 feet by 95 feet, and is surrounded by a chain-link fence with a locking gate (Figure 2.2). A majority of the site is unpaved and vegetated with grasses. There is a paved driveway that leads to the now-decontaminated Building 560, a slab-on-grade construction. The area surrounding the site is level and also is vegetated with grasses. To the north and east of the site is a gravel road, and beyond the road are railroad tracks that are no longer in use. The area to the south and west has been used in the recent past as a storage yard for stockpiled telephone poles and drummed nonhazardous material. To the south and east of the site are office buildings and parking lots. Farther east, the active Base runways are used by various military aircraft and private aircraft associated with the RPA.

Building 560 is a 15-foot-square pre-engineered metal structure with an adjacent 4-foot-square concrete drum wash pad. From 1974 to 1983, Building 560 (Figure 2.2) housed water demineralization equipment. Records indicate Buildings 551 and 552,

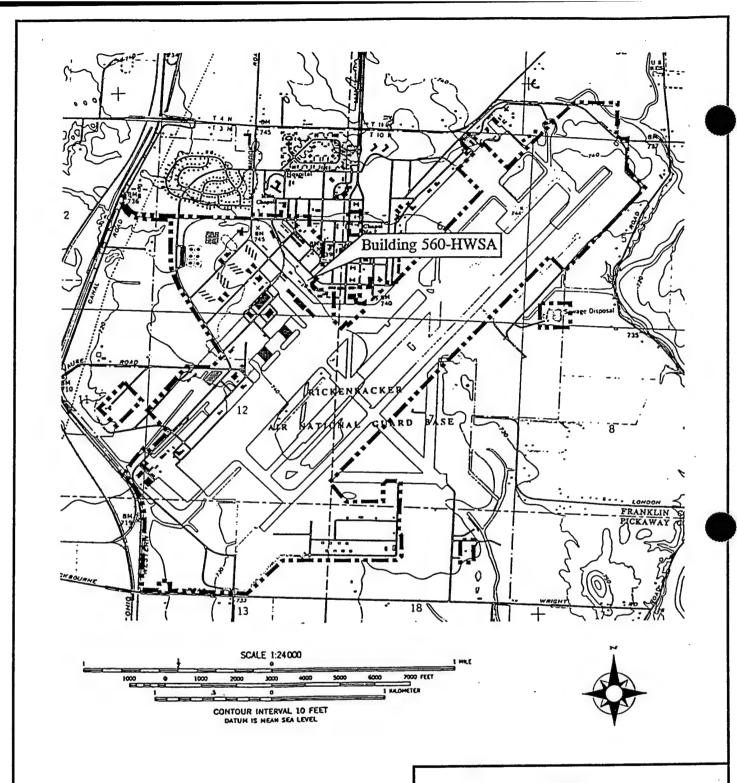


FIGURE 2.1

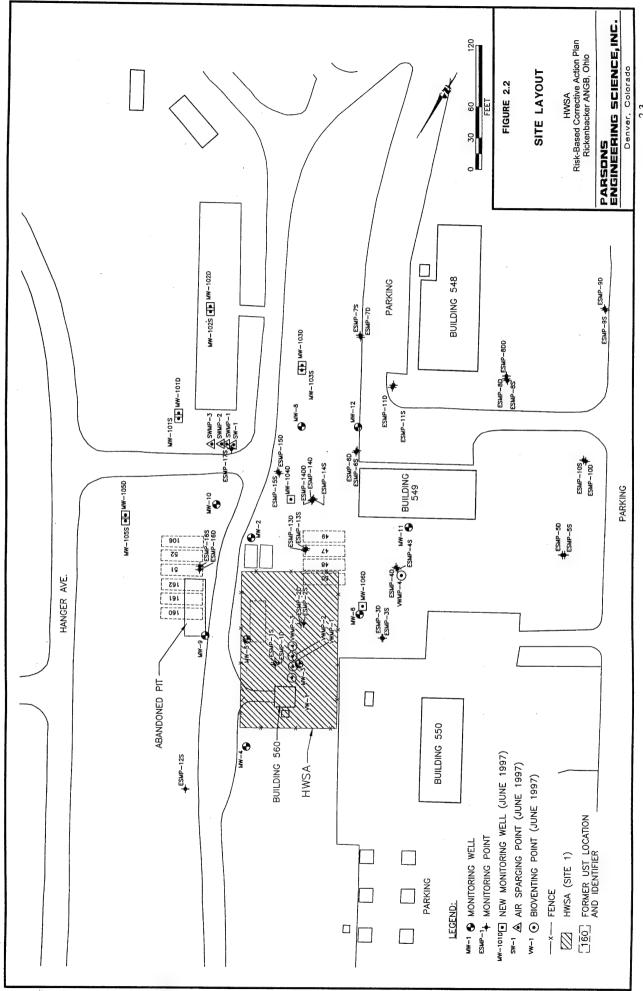
SITE LOCATION

HWSA Risk-Based Corrective Action Plan Rickenbacker ANGB, Ohio

PARSONS ENGINEERING SCIENCE, INC.

Denver, Colorado

Sources: USGS, 1985.



fuel pumping facilities, also were formerly located at this site. One of the buildings housed the valve controls for fuel hydrants used to off-load fuel from rail cars.

In 1983, the HWSA was used as a RCRA-permitted storage facility that received hazardous waste generated during Base activities, and was used for that purpose from 1983 to 1986. Building 560 was converted to a hazardous waste storage facility. The conversion included sealing all floor drains that led to storm sewers, connecting remaining drains to the sanitary sewer, and installing emergency eye-wash and shower fixtures. Wastes were containerized and brought to the site from other areas of the Base. Drum contents were characterized wastes for disposal or reuse and were then turned over to the Defense Property Disposal Office (DPDO) for disposal or recycling. The DPDO is now known as the Defense Reutilization and Materials Office (DRMO).

Building 560 was used to store small (5 gallons or less) containers that usually held acids or spent desiccants. Other materials stored at this site were containerized in 55-gallon drums that were stored outside the building within the fenced area. At any given time, as many as 165 containers were stored on pallets in the grassy area outside Building 560. Table 2.1 summarizes types and quantities of waste stored at the site from 1983 until 1986, when the HWSA was closed. Various chemical constituents likely to be present in some of these wastes have been detected in site soils and or groundwater (see Sections 3 and 4).

Additionally, five 12,000-gallon USTs (Tanks 53 through 57) and ten 25,000-gallon USTs (Tanks 47 through 52, 106, 160, 161, and 162) were previously located at or near the site. The former locations of the ten 25,000-gallon USTs are shown on Figure 2.2. The five former 12,000-gallon USTs were located approximately 700 feet north-northwest of Building 560. The four former USTs at the south corner of the site (Tanks 47 through 50) were used to store waste fuel, waste oil, and deicing fluid. The six former USTs located to the east of the site (Tanks 51, 52, 106, 160, 161, and 162) were used to store jet fuel, but were abandoned in 1964. Four of the former 12,000-gallon USTs were used to store diesel fuel and the fifth UST stored kerosene. These

TABLE 2.1 SUMMARY OF WASTE STORAGE HAZARDOUS WASTE STORAGE AREA RICKENBACKER ANGB, OHIO

		Quantity (by year)			
	USEPA Haz.	1983	1984	1985	1986
Waste Description	Waste No.	(gal)	(gal)	(gal)	(lb)
PD 680 (Stoddard Solvent, Flammable Aliphatic Petroleum Distillate	D001	1,155	1,450	110	2,429
Carbon-Removing Compound (Methylene Chloride, Creosols, Phenols)	F001, F004	590	870		
Highly Aromatic Naptha	D001	500	290		
Methyl Ethyl Ketone	F005	285	525		1,050
Paint Remover	F005	200	540		
Bromochloromethane	_	200	385	-	
Sulfuric Acid	D002	25	25		-
Paint Thinner	D001	20	20		
Ethanolamine and Benzyl Alcohol	D001	220			
Oily Water and Cleaning Solutions Containing Lead, Cadmium, Chromium, and Nickel	D006, D007, D008	495		-	
Hydraulic Fluid		0	360	••	-
Synthetic Oil			440		
Inspection Penetrant	_		150		
Organic Peroxide	D002			1	.
Spent Desiccant (Cobalt chloride)	-	50	40	10	-

Quantities for desiccant for all years are shown in pounds.

five USTs were emptied and abandoned in 1981. All fifteen USTs were removed in 1994 and 1995 (Ogden Environmental and Energy Services, inc. [Ogden], 1995a and 1995b). AFBCA received NFRAP status from the Ohio Department of Commerce, Division of State Fire Marshal, BUSTR in March 1996 for Tanks #51 through 57, 106, 160, 161, and 162.

The NFRAP status for Tanks 47 through 50 excluded soil and groundwater, and was for the tanks only. Residual petroleum contamination in soils and groundwater associated with these four tanks is addressed in this amended closure plan.

2.2 PHYSICAL SETTING

2.2.1 Site Topography and Surface Hydrology

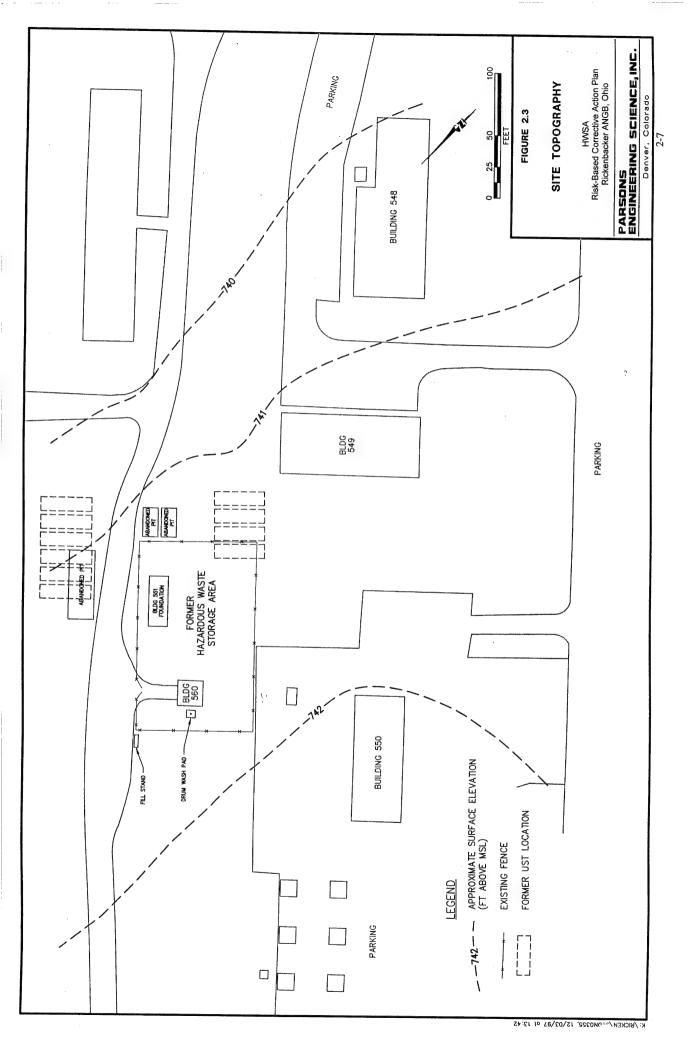
The site topography varies from approximately 740 to 742 feet above mean sea level (msl) (Figure 2.3). The site slopes gently to the east. No significant surface water drainages are present at the site.

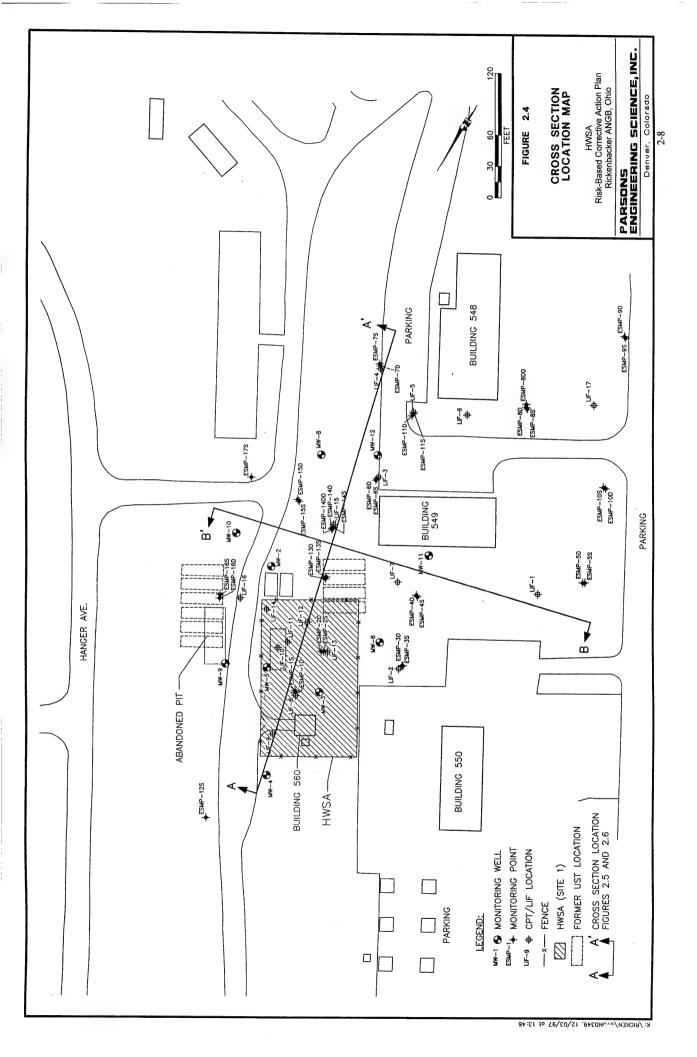
2.2.2 Site Geology and Hydrogeology

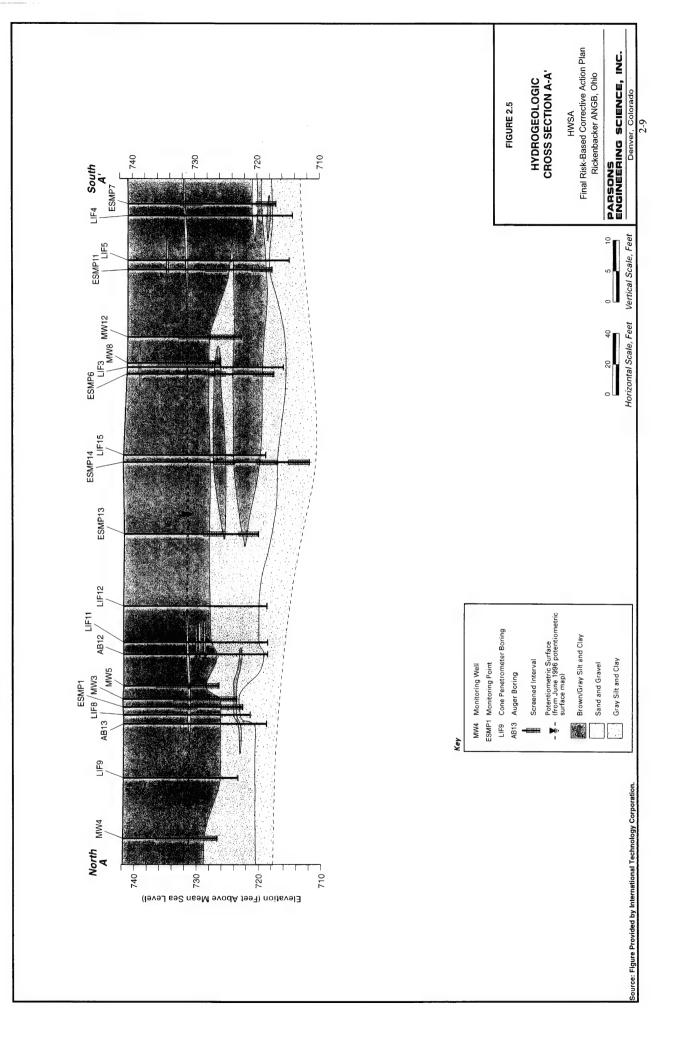
Geologic cross sections were prepared for the HWSA using the lithologic descriptions from the soil borings and monitoring well borings and the results of cone penetrometer (CPT) testing. The cross section locations are shown in Figure 2.4, and the cross sections are in Figures 2.5 and 2.6.

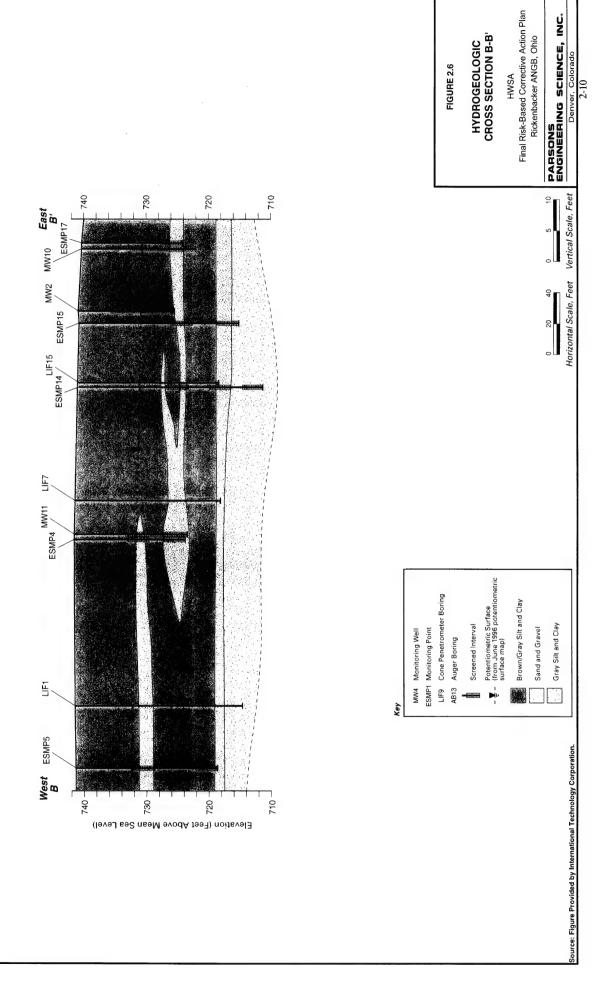
Brown and gray silt and clay are present at HWSA from the surface to a depth of approximately 10 to 14 feet bgs. Within and beneath the uppermost silt and clay, sand and gravel lenses were encountered in most borings. The sand and gravel lenses appear to be locally continuous throughout the HWSA. However, these lenses are not laterally continuous across Rickenbacker ANGB, and represent localized hydraulic regimes (IT, 1996a). The sand and gravel lenses range in thickness from 8 feet in the northern part of the site to less than 5 feet in the central and southern portions of the site. The sand and gravel deposits at the site are underlain by the gray basal till, typically first encountered between 19 and 25 feet bgs. This clay unit acts as a barrier to vertical

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migration of contaminants toward the intermediate and deep aquifers, which are used regionally for water supplies (Section 1.3.5).

The water level measurements and groundwater elevations from quarterly sampling events performed at the HWSA between August 1995 and March 1997 are presented in Table 2.2. These data indicated depths to shallow groundwater ranges between 8 and 12 feet bgs. Review of the potentiometric data indicate that there is generally a considerable difference (often more than 2 feet) in the groundwater elevation in adjacent monitoring point pairs screened at different depths within the UWBZ. The vertical gradient across the site is predominantly downward, though upward gradients The flow direction and horizontal gradient of shallow also have been observed. groundwater was determined using data from monitoring points screened in the same (laterally continuous) stratigraphic unit. Monitoring points ESMP-1D, 4D, 13S, 14S, and 17S are all screened in a sand zone encountered at a depth of approximately 14 feet bgs. Based on measurements from these monitoring points, Figure 2.7 and Figure 2.8 illustrate the potentiometric surface from the June 1996 and December 1996 sampling events, respectively. The groundwater flow direction within this sand lens is generally eastward, with a horizontal hydraulic gradient of approximately 0.0058 foot per foot (ft/ft), or approximately 1 foot change in elevation every 170 feet horizontally. The potentiometric surfaces illustrated on these figures are representative of the groundwater flow directions and gradients noted during other sampling events.

Hydraulic conductivities at the site were estimated from the rising head slug tests conducted in early 1995 (Parsons ES, 1997a); the test results are summarized in Table 2.3. Values for the hydraulic conductivity ranged from 0.13 foot per day (ft/day) to 2.79 ft/day. The average hydraulic conductivity at the HWSA, as determined from these tests, is 1.03 ft/day. Using a hydraulic gradient of 0.0058 ft/ft and assuming an effective porosity of 0.3, the estimated linear advective velocity of groundwater at the HWSA is 7.3 feet per year (ft/yr). IT (1997a) estimated linear groundwater velocities ranging from 4.97 ft/yr to 32.42 ft/yr based on 1996 data from monitoring wells screened in more than one hydrologic unit, and ranging from 5.85 ft/yr to 7.73 ft/yr

TABLE 2.2 GROUNDWATER ELEVATION DATA AUGUST 1995 TO MARCH 1997

HAZARDOUS WASTE STORAGE AREA RICKENBACKER ANGB, OHIO

		8/1:	5/95	12/20/9		3/1	8/96	6/2	25/96
	Datum								Til
Well ID	Elevation	Depth to	Elevation (ft	Depth to	Elevation	Depth to	Elevation (ft	_	Elevation (fi
	(ft msl) ^{2/}	Water (ft)b/	msl)	Water (ft)	(ft msl)	Water (ft)	msl)	Water (ft)	msl)
MW-02	743.36	12.14	731.22	13.23	730.13	11.26	732.10	12.00	731.36
MW-03	743.96	10.46	733.50	13.05	730.91	8.60	735.36	11.49	732.47
MW-04	745.15	11.38	733.77	13.26	731.89	9.86	735.29	11.81	733.34
MW-05	744.97	13.37	731.60	NA (free product)				13.11	731.86
MW-06	745.18	13.75	731.43	14.96	730.22	13.06	732.12	13.70	731.48
MW-08	743.89	10.53	733.36	12.94	730.95	8.43	735.46	11.28	732.61
MW-09	745.25	8.05	737.20	13.55	731.70			7.86	737.39
MW-10	742.64	11.23	731.41	12.30	730.34			11.06	731.58
MW-11	744.15	12.19	731.96	13.55	730.60	11.64	732.51	12.11	732.04
MW-12	743.02	13.10	729.92	13.71	729.31	11.98	731.04	12.28	730.74
ESMP-1S	741.67	7.43	734.24	9.44	732.23	5.65	736.02	8.28	733.39
ESMP-1D	741.72	10.17	731.55	10.56	731.16	8.87	732.85	10.00	731.72
ESMP-2S	741.18	3.57	737.61	5.83	735.35	1.73	739.45	4.38	736.80
ESMP-2D	741.29	6.18	735.11	10.10	731.19	8.55	732.74	9.64	731.65
ESMP-3S	742.23	7.68	734.55	11.26	730.97	7.82	734.41	8.94	733.29
ESMP-3D	742.22	10.69	731.53	11.92	730.30	9.94	732.28	10.55	731.67
ESMP-4S	742.70	5.94	736.76	9.92	732.78	6.92	735.78	10.62	732.08
ESMP-4D	742.69	11.15	731.54	12.23	730.46	10.42	732.27	10.93	731.76
ESMP-5S	741.51	3.83	737.68	6.89	734.62	3.46	738.05	4.87	736.64
ESMP-5D	741.56	6.20	735.36	9.00	732.56	2.07	739.49	7.10	734.4
ESMP-6S	740.98	10.79	730.19	NA				11.92	729.0
ESMP-6D	741.05	9.69	731.36	10.69	730.36	8.87	732.18	9.48	731.57
ESMP-7S	740.85	6.02	734.83	5.33	735.52	3.69	737.16	6.60	734.25
ESMP-7D	740.80	5.96	734.84	8.62	732.18	5.57	735.23	6.75	734.05
ESMP-8S	740.92	3.76	737.16	6.73	734.19	2.87	738.05	4.80	736.12
ESMP-8D	740.89	6.17	734.72	8.85	732.04	10		6.78	734.11
ESMP-8DD	740.83	7.38	733.45	9.58	731.25	7.07	733.76	7.84	732.99
ESMP-9S	741.79	3.38	738.41	5.22	736.57	1.46	740.33	5.14	736.65
ESMP-9D	741.70	6.53	735.17	9.35	732.35	6.27	735.43	7.36	734.34
ESMP-10S	741.56	6.15	735.41	10.03	731.53	7.26	734.30	8.00	733.56
ESMP-10D	741.54	6.41	735.13	9.22	732.32	6.28	735.26	7.22	734.32
ESMP-11S	740.76	6.66	734.10	7.15	733.61	6.42	734.34	7.06	733.70
ESMP-11D	740.80	7.11	733.69	9.04	731.76	6.69	734.11	7.52	733.28
ESMP-12S	742.43	2.57	739.86	6.23	736.20	0.40	742.03	4.21	738.22
ESMP-13S	741.38	10.34	731.04	10.90	730.48	9.34	732.04	10.04	731.34
ESMP-13D	741.38	10.06	731.32	10.56	730.82	9.10	732.28	9.90	731.48
ESMP-14S	741.17	10.13	731.04	10.31	730.86	9.16	732.01	9.84	731.33
ESMP-14D	741.18	9.89	731.29	10.37	730.81	8.92	732.26	9.66	731.52
ESMP-14DD	741.13	9.72	731.41	10.56	730.57	8.93	732.20	9.42	731.71
ESMP-15S	740.37	9.45	730.92	9.68	730.69	8.47	731.90	9.25	731.12
ESMP-15D	740.28	8.98	731.30	9.45	730.83	7.98	732.30	8.80	731.48
ESMP-16S	740.33	8.00	732.33	9.60	730.73	7.78	732.55	8.36	731.97
ESMP-16D	740.33	8.77	731.56	9.18	731.15	7.87	732.46	8.62	731.71
ESMP-17S	739.87	9.19	730.68	9.56	730.31	8.52	731.35	9.19	730.68

TABLE 2.2 (Continued) GROUNDWATER ELEVATION DATA

AUGUST 1995 TO MARCH 1997

HAZARDOUS WASTE STORAGE AREA RICKENBACKER ANGB, OHIO

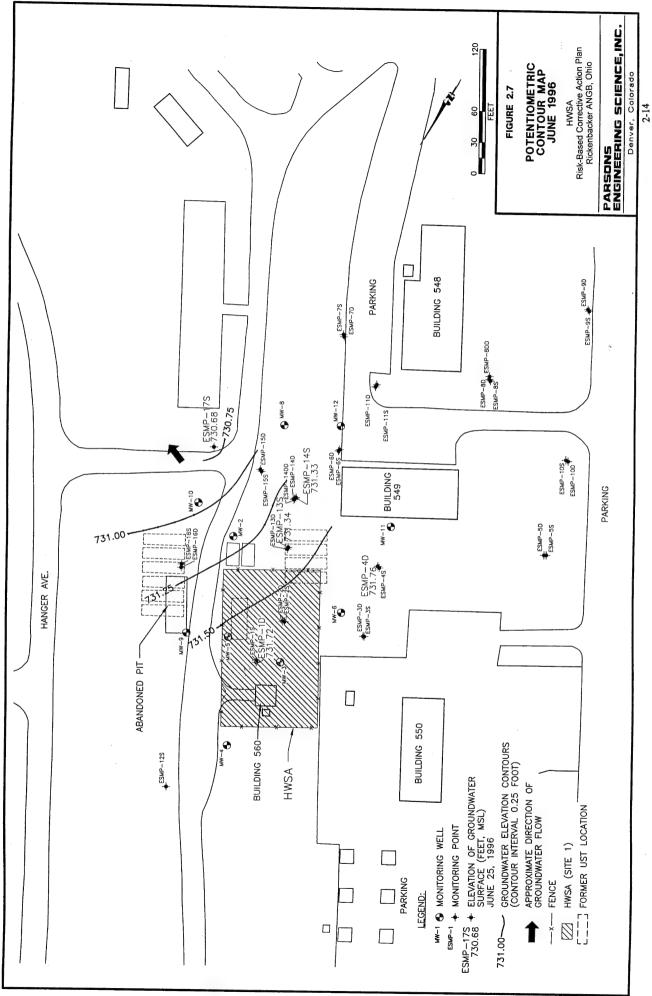
r			3/96	ANGB, OH	2/96	2/19	3/97
	Datum	9/2	3/90	121.	2190	3/10	3/7/
W !! ID	Elevation	D	Tiloundian	Depth to	Elevation	Depth to	Elevation
Well ID		Depth to	Elevation	-	(ft msl)	Water (ft)	(ft msl)
	(ft msl) ^{a/}	Water (ft)	(ft msl)	Water (ft)	(It msi)		(It msi)
MW-02	743.36	15.37	727.99	10.63	732.73	10.39	732.97
MW-03	743.96	15.47	728.49	10.44	733.52	9.18	734.78
MW-04	745.15	16.00	729.15	10.30	734.85	9.69	735.46
MW-05	744.97	13.92	731.05			11.67	733.30
MW-06	745.18	17.12	728.06	12.28	732.90	12.08	733.10
MW-08	743.89	15.78	728.11	8.36	735.53	7.91	735.98
MW-09	745.25						
MW-10	742.64						
MW-11	744.15	15.60	728.55	10.95	733.20	10.78	733.37
MW-12	743.02	15.07	727.95	10.32	732.70	10.39	732.63
ESMP-1S	741.67						
ESMP-1D	741.72						
ESMP-2S	741.18						
ESMP-2D	741.29	10.07	731.22	8.30	732.99	8.00	733.29
ESMP-3S	742.23						
ESMP-3D	742.22	10.96	731.26	9.15	733.07	8.93	733.29
ESMP-4S	742.70	11.08	731.62	9.47	733.23	9.54	733.16
ESMP-4D	742.69	11.33	731.36	9.52	733.17	9.44	733.25
ESMP-5S	741.51						
ESMP-5D	741.56						
ESMP-6S	740.98						
ESMP-6D	741.05	9.86	731.19	8.06	732.99	7.83	733.22
ESMP-7S	740.85						
ESMP-7D	740.80						
ESMP-8S	740.92	6.80	734.12	4.18	736.74	3.33	737.59
ESMP-8D	740.89						
ESMP-8DD	740.83					•	
ESMP-9S	741.79						
ESMP-9D	741.70						
ESMP-10S	741.56	9.96	731.60	8.29	733.27	7.38	734.18
ESMP-10D	741.54						
ESMP-11S	740.76						
ESMP-11D	740.80						
ESMP-12S	742.43						
ESMP-13S	741.38	10.36	731.02	8.64	732.74	8.38	733.00
ESMP-13D	741.38						
ESMP-14S	741.17						
ESMP-14D	741.18	10.13	731.05	8.29	732.89	8.01	733.17
ESMP-14DD	741.13						
ESMP-15S	740.37						
ESMP-15D	740.28			7			
ESMP-16S	740.33	9.11	731.22	7.18	733.15	6.51	733.82
ESMP-16D	740.33	9.07	731.26	7.27	733.06	6.94	733.39
ESMP-17S	739.87	9.41	730.46	7.87	732.00	7.81	732.06

Note: Blank spaces indicate water level not measured during this sampling event.

of ft msl = Feet above mean sea level.

Feet below top of well casing.

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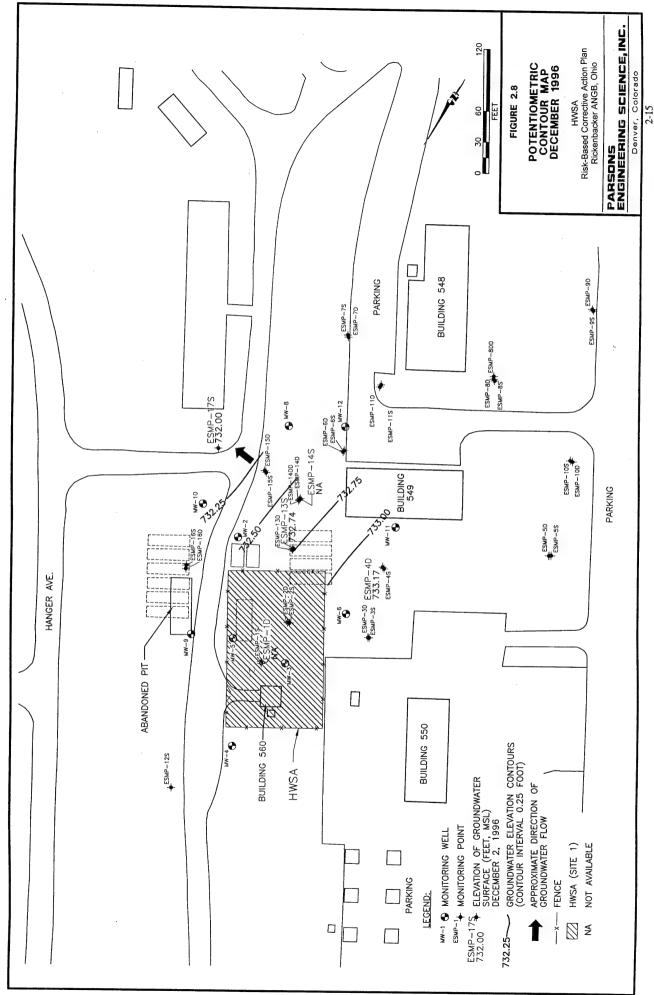


TABLE 2.3
HYDRAULIC CONDUCTIVITY (K) VALUES FROM SLUG TESTS, MARCH 1995
HAZARDOUS WASTE STORAGE AREA
RICKENBACKER ANGB, OHIO

Well	K (ft/min)	K (ft/day)
MW-4	5.23E-04	0.75
MW-4	4.59E-04	0.66
MW-6	1.94E-03	2.79
MW-6	9.01E-04	1.30
MW-9	8.77E-04	1.26
MW-9	6.56E-04	0.94
MW-12	9.88E-05	0.14
MW-12	8.98E-05	0.13

Average K	7.17E-04	1.03
Minimum K	8.98E-05	0.13
Maximum K	1.94E-03	2.79

based on monitoring points screened in the sand seam. Due to the heterogeneity of the unconsolidated deposits at the site, the advective flow velocity calculated for the sand seam is assumed to represent the most reliable estimate of advective velocity in the vicinity of the HWSA (IT, 1997a). Advective velocities for the site based on additional assessment activities conducted in 1997 are reviewed in Section 4.1.2.1.

2.3 LAND USE

2.3.1 Site Access

The HWSA is centrally located within Rickenbacker ANGB (Figure 2.1). Access to Rickenbacker ANGB is restricted by a chain-link fence surrounding the Base and by security personnel stationed at the Base entrance 24 hours a day. Active flying missions by military aircraft, as well as other ongoing and proposed private/commercial flying activities, ensure that access to the facility will continue to be restricted. The inactive HWSA is surrounded by a 6-foot-high chain-link fence with a locked gate (Figure 2.2).

2.3.2 Current Land Use

Currently, the HWSA is an inactive, interim-status RCRA facility. The site area and the surrounding area within a 150-foot radius, have not been used for any purpose since the 1995 removal of the four USTs in the southern corner of the site (Figure 2.2). Use of surrounding Rickenbacker ANGB facilities also has been limited because of Base closure/conversion activities initiated by the Department of Defense (DOD) in the mid-1980s. Buildings 548, 549, and 550, to the south and west of the HWSA, are unoccupied. Hangars used for airfield support, located approximately 500 feet to the south and west of the site, represent the closest, routinely used Rickenbacker ANGB facilities.

Off-Base land usage closest to the site is commercial/industrial in nature. The RPA owns property to the north and east of the HWSA across Hangar Avenue (Figure 2.2). RPA land adjacent to Hangar Avenue is undeveloped. The closest occupied RPA-

owned facilities are located more than 500 feet from the HWSA, and are leased for commercial use. Active base runways are located approximately 1,000 feet east and south of the HWSA, and the eastern and southern Base boundaries are located approximately 1.3 miles and 2 miles from the site, respectively.

RPA and Rickenbacker ANGB obtain their water supply from the city of Columbus (see Section 1.3.5). None of the former water supply wells at the Base are used. The active water supply wells closest to the HWSA and screened in the UWBZ are located in the village of Lockbourne approximately 1.5 miles to the southwest.

2.3.3 Proposed Land Use

The proposed future use of the HWSA and surrounding property is industrial/commercial in nature. In July 1997, the RPA submitted a concept plan for redevelopment of the HWSA and surrounding area to the Federal Aviation Administration (FAA). The concept plan was approved by the FAA in September 1997. Under the plan, the HWSA property and all property within 700 to 800 feet of the site will be used for airfield support. The HWSA and areas to the northwest and southeast, including the aircraft runways are included in Parcel D1, which is planned for development under Phase I of the concept plan (see Appendix A). A concrete aircraft and aircraft support equipment access taxiway is planned for installation over the site. Buildings in the immediate vicinity of the HWSA, including Building 560, will be demolished as part of the Phase I development. The existing aircraft hangars south and west of the HWSA are expected to remain in place.

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SECTION 3

PREVIOUS SITE INVESTIGATIONS AND REMEDIATION ACTIVITIES

Preliminary environmental investigations were conducted at the HWSA in 1988 (ES, 1990), 1990 (ES, 1992a), and 1991 (ES, 1992b). Additional fieldwork was conducted in February and March 1995 to evaluate the potential for naturally occurring chemical attenuation processes to effect *in situ* remediation of dissolved contamination at the HWSA (Parsons ES, 1995a and 1995b, and 1997a). Supplemental data were collected during subsequent groundwater sampling events between August 1995 and December 1996 (IT, 1995b, 1996b, 1997a, and 1997b) and remediation activities pursuant to site closure which were conducted at the HWSA in 1995 and 1996 (Ogden, 1995a and 1995b; Parsons ES, 1996). The activities conducted during these investigations are summarized in this section, and supporting data are presented in Appendices B and D. Analytical results are reviewed in Section 4.

3.1 PREVIOUS INVESTIGATIONS

3.1.1 Pre-Closure Sampling

During the investigations conducted from 1988 through 1991, surface soil, subsurface soil, soil gas, and groundwater samples were collected and analyzed for volatile organic compounds (VOCs), semivolatile organic compounds (SVOCs), and metals. The analytes targeted in these sampling efforts are listed in Table 3.1. A total of 15 boreholes were completed, 12 monitoring wells were installed, and numerous soil, soil gas, and groundwater samples were collected during pre-closure sampling activities (ES, 1990, 1992a, and 1992b). The borings (including wells) ranged in depth from 10 to 27 feet bgs (Table 3.2).

TABLE 3.1

LIST OF TARGET ANALYTES FOR PREVIOUS INVESTIGATIONS (1988-1991) HAZARDOUS WASTE STORAGE AREA RICKENBACKER ANGB, OHIO

Base/Neutral Extractable Sem	nivolatile Organic Compounds (USEPA Method SW 8270)
Acenaphthene	Fluoranthene
Acenaphthylene	Fluorene
Anthracene	
	Hexachlorobenzene
Benzo(b)fluoranthene	Hexachlorobutadiene
Benzo(k)fluoranthene	Hexachloroethane
Benzo(a)pyrene	Hexachlorocyclopentadiene
Benzo(a)anthracene	
Benzo(ghi)perylene	Indeno(1,2,3-cd)pyrene
Benyl Alcohol *	Isophorone
Bis(2-chloroethyl)ether	
Bis(2-chloroethoxy)methane	Naphthalene
Bis(2-ethylhexyl)phthalate	Nitrobenzene
Bis(2-chloroisopropyl)ether	N-Nitrosodiphenylamine
4-Bromophenyl phenyl ether	2-Nitroaniline
Butylbenzlphthalate	3-Nitroaniline
	4-Nitroaniline
2-Chloronaphthalene	N-Nitroso-Dimethylamine *
4-Chloroaniline	N-Nitroso-di-n-dipropylamine
4-Chlorophenyl phenyl ether	
Chrysene	2-Methylnaphthalene
Dibenzo(a,h)anthracene	Phenanthrene
Dibenzofuran	Pyrene
Di-n-octylphthalate	
1,3-Dichlorobenzene	1,2,4-Trichlorobenzene
1,2-Dichlorobenzene	
1,4-Dichlorobenzene	
3,3'-Dichlorobenzidine	
Diethyl phthalate	
Dimethyl phthalate	
2,4-Dinitrotoluene	
2,6-Dinitrotoluene	
Di-n-octylphthalate	

TABLE 3.1 (CONTINUED)

LIST OF TARGET ANALYTES FOR PREVIOUS INVESTIGATIONS (1988-1991) HAZARDOUS WASTE STORAGE AREA RICKENBACKER ANGB, OHIO

Volatile Organic	Compounds (USEPA Methods SW 8240/8260)
Acrolein *	1,2-Dichloroethane
Acetone	trans-1,2-Dichloroethene
Acrylonitrile *	trans-1,3-Dichloropropene
Benzene	Ethylbenzene
Bromomethane	
Bromodichloromethane	2-Hexanone
Bromoform	
2-Butanone	Methylene Chloride
	4-Methyl-2-pentanone
Carbon disulfide	
Carbon tetrachloride	Styrene
Chlorobenzene	
Chloroethane	1,1,2,2-Tetrachloroethane
Chloroform	Tetrachloroethene
2-Chloroethyl vinyl ether *	Toluene
Chloromethane	1,1,1-Trichloroethane
	1,1,2-Trichloroethane
Dibromochloromethane	Trichloroethene
1,2-Dichloropropane	Trichlorofluoromethane *
1,3-Dichlorobenzene *	
cis-1,3-Dichloropropene	Vinyl chloride
1,2-Dichlorobenzene *	Vinyl Acetate *
1,4-Dichlorobenzene *	
1,1-Dichloroethene	Xylenes

Me	als (USEPA Methods SW6010 and SW7470/7471)	
Antimony	Mercury	
Arsenic	Nickel	
Beryllium	Selenium	
Cadmium	Silver	
Chromium	Thallium	
Copper	Zinc	
Lead		

^{*} These compounds were not on the Target Compound List (TCL) for the method, but were included in the laboratory report.

TABLE 3.2 MONITORING POINT AND WELL COMPLETION DETAILS HAZARDOUS WASTE STORAGE AREA RICKENBACKER ANGB, OHIO

Installation	\vdash	⊢	Datum	Ground	PVC	Screen	Total	Depth to Top	Depth to Base
Date	Easting	Northing	Elevation (ft msl)	Elevation (ft msl)	Casing ID (inches)	Length (feet)	Depth _b / (ft btoc) ^b /	of Screen (ft btoc)	of Screen (ft btoc)
2/23/95	5 1845016	662614	741.67	741.60	0.50	3.28	11.70	8.42	11.70
2/23/95	5 1845015	662615	741.72	741.60	0.50	3.28	18.80	15.52	18.80
2/23/95	5 1845023	662568	741.18	741.20	0.50	3.28	11.42	8.14	11.42
2/23/95	5 1845025	662569	741.29	741.20	0.50	3.28	22.50	19.22	22.50
2/23/95	5 . 1844958	662524	742.23	741.80	0.50	3.28	12.55	9.27	12.55
2/23/95	5 1844959	662526	742.22	741.80	0.50	3.28	22.19	18.91	22.19
2/23/95	5 1844996	662467	742.70	742.60	0.50	3.28	12.58	9.30	12.58
2/23/95	5 1844997	662467	742.69	742.60	0.50	3.28	18.44	15.16	18.44
2/23/95	5 1844891	662345	741.51	741.50	0.50	3.28	12.51	9.23	12.51
2/23/95	5 1844893	662346	741.56	741.50	0.50	3.28	22.54	19.26	22.54
2/23/95	5 1845105	662411	740.98	741.00	0.50	3.28	15.80	12.52	15.80
2/23/95	5 1845105	662412	741.05	741.00	0.50	3.28	23.51	20.23	23.51
2/23/95	5 1845181	662330	740.85	740.80	0.50	3.28	11.75	8.47	11.75
2/23/95	1845179	662331	740.80	740.80	0.50	3.28	23.59	20.31	23.59
2/23/95	5 1845049	662261	740.92	740.80	0.50	3.28	10.72	7.44	10.72
2/23/95	5 1845050	662261	740.89	740.90	0.50	3.28	22.81	19.53	22.81
2/23/95	1845052	662259	740.83	740.80	0.50	3.28	29.74	26.46	29.74
2/23/95	35 1845034	662143	741.79	741.60	0.50	3.28	11.81	8.53	11.81
2/23/95	35 1845036	662143	741.70	741.60	0.50	3.28	21.85	18.57	21.85
2/23/95	35 1844945	662265	741.56	741.50	0.50	3.28	15.84	12.56	15.84
2/23/95	95 1844942	662265	741.54	741.50	0.50	3.28	22.07	18.79	22.07
2/23/95	95 1845124	662341	740.76	740.80	0.50	3.28	15.78	12.50	15.78
2/23/95	95 1845123	662343	740.80	740.80	0.50	3.28	22.82	19.54	22.82
2/24/95	95 1844991	1 662762	741.43	741.30	0.50	3.28	15.73	12.45	15.73



TABLE 3.2 (Continued)
MONITORING POINT AND WELL COMPLETION DETAILS
HAZARDOUS WASTE STORAGE AREA
RICKENBACKER ANGB, OHIO

				_		_			_		_				\neg	\neg				
Depth to Base of Screen (ft btoc)	16.18	21.48	17.70	24.58	29.71	17.52	24.90	15.57	22.65	15.62	16.91	20.10	18.30	18.04	17.99	18.44	18.27	20.16	19.77	20.06
Depth to Top of Screen (ft btoc)	12.90	18.20	14.42	21.30	26.43	14.24	21.62	12.29	19.37	12.34	6.91	10.10	8.30	8.04	7.99	8.44	8.27	10.16	<i>LL</i> .6	10.06
Total Depth b/ (ft btoc) b/	16.18	21.48	17.70	24.58	29.71	17.52	24.90	15.57	22.65	15.62	16.91	20.10	18.30	18.04	17.99	18.44	18.27	20.16	19.77	20.06
Screen Length (feet)	3.28	3.28	3.28	3.28	3.28	3.28	3.28	3.28	3.28	3.28	10	10	10	10	10	10	10	10	10	10
PVC Casing ID (inches)	0.50	0.50	0.50	0.50	0.50	0.50	0.50	0.50	0.50	0.50	2	2	2	2	2	2	2	2	2	2
Ground Elevation (ft msl)	741.40	741.40	741.40	741.00	741.00	740.20	740.20	740.30	740.30	739.90	741.10	741.60	741.80	741.60	741.70	740.40	741.60	740.30	741.40	740.80
Datum Elevation (ft msl)	741.38	741.38	741.17	741.18	741.13	740.37	740.28	740.33	740.33	739.87	743.36	743.96	745.15	744.97	745.18	743.89	745.25	742.64	744.15	743.02
Northing	662515	662515	662475	662474	662476	662479	662478	662598	662601	662495	662544	662299	662691	662617	662524	662431	662643	662544	662429	662394
Easting	1845072	1845071	1845102	1845103	1845103	1845143	1845145	1845132	1845130	1845192	1845118	1844999	1844979	1845051	1844991	1845159	1845083	1845165	1845017	1845120
Installation Date	2/24/95	2/24/95	2/24/95	2/24/95	2/24/95	2/24/95	2/24/95	2/24/95	2/24/95	2/24/95	7/29/88	8/10/88	1/29/90	1/31/90	1/30/90	1/30/90	2/9/90	10/14/91	10/12/91	10/15/91
Location	ESMP-13S	ESMP-13D	ESMP-14S	ESMP-14D	ESMP-14DD	ESMP-15S	ESMP-15D	ESMP-16S	ESMP-16D	ESMP-17S	MW-2	MW-3	MW-4	MW-5	9-MM	MW-8	6-WM	MW-10	MW-11	MW-12

" ft ms! = feet above mean sea level.

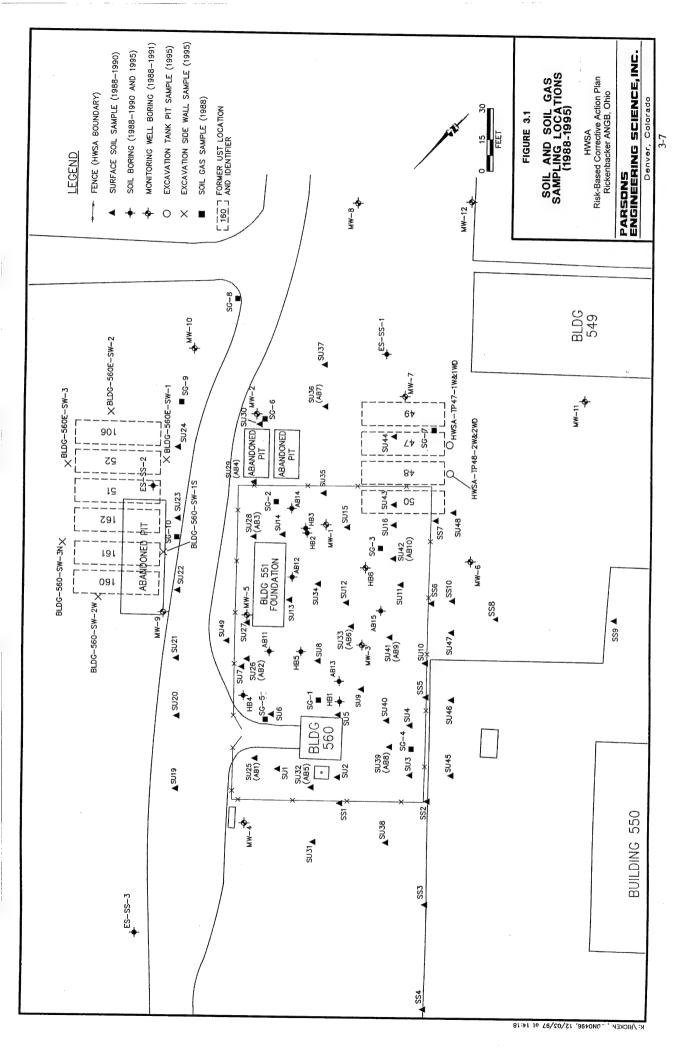
022/73196/CP/1.XLS/Table 3.2

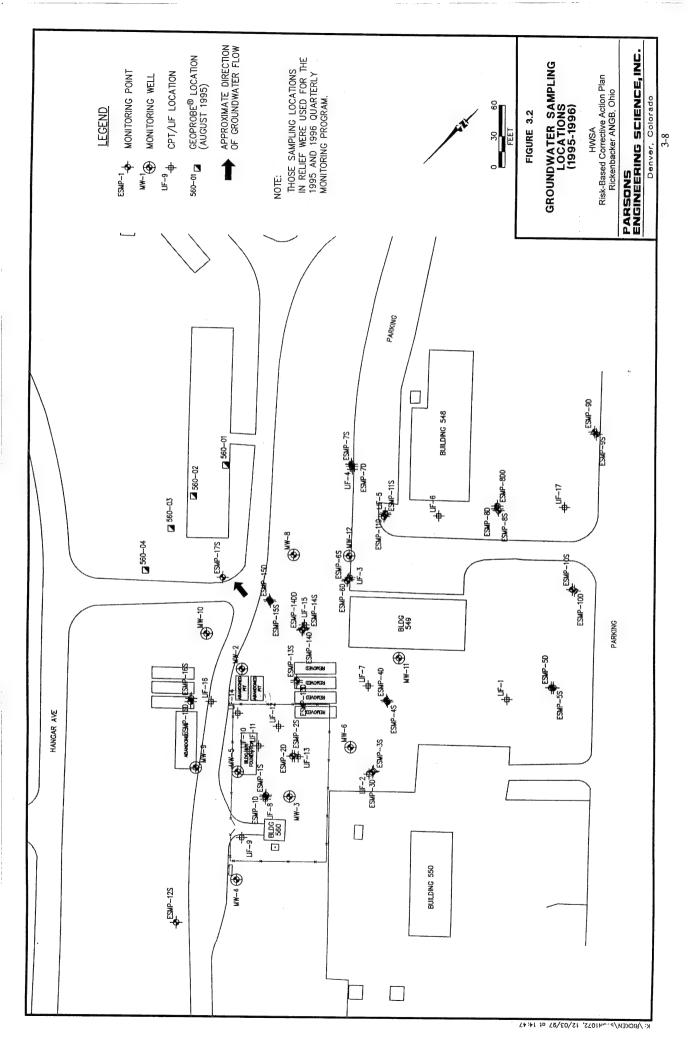
b' ft btoc = feet below top of casing.

Soil and soil gas sampling locations from these investigations as well as from the 1995 site investigations are shown on Figure 3.1. The sampling results indicated residual VOC, SVOC, and metals contaminants in site soils, and dissolved in groundwater. The analytical results from the investigations conducted prior to 1995 are compiled in tables included in Appendix B. These data, combined with data collected during 1997 additional assessment activities at the HWSA, are used in Section 4.2 to describe the nature and extent of contamination at the site.

3.1.2 Parsons ES Natural Chemical Attenuation Investigation

In February/March 1995, personnel from Parsons ES, US Army Corps of Engineers (USACE), and the USEPA National Risk Management Research Laboratory (NRMRL, formerly the Robert S. Kerr Environmental Research Laboratory) performed RNA sampling to determine if natural chemical attenuation processes were occurring at the HWSA, and if so, to evaluate if these processes can play a significant role in groundwater remediation. Additional data were collected to evaluate near-surface geology, aquifer properties, and the nature and extent of soil and groundwater contamination at the site. Site characterization activities included performing CPT with laser-induced fluorescence (LIF); sampling and analyzing soils from CPT pushes; installing groundwater monitoring points; sampling and analyzing groundwater from the monitoring points and previously installed monitoring wells; and measuring and estimating hydrogeologic parameters of the UWBZ (static groundwater levels, groundwater gradient, groundwater flow direction, and hydraulic conductivity). The field methods for all site activities are described in the Draft Work Plan for a Treatability Study in Support of the Intrinsic Remediation (Natural Attenuation) Option at Site 1 (Hazardous Waste Storage Area), Rickenbacker ANGB (Parsons ES, 1995b). Groundwater sampling locations for the 1995 RNA study, and for additional groundwater monitoring activities performed in 1995 and 1996, are shown on Figure 3.2.





The following data were collected during this investigation:

- Depth from measurement datum to the water table or potentiometric surface in monitoring wells;
- Slug test water level recovery rates (used to estimate hydraulic conductivity);
- Location of potential groundwater recharge and discharge areas;
- Stratigraphy of subsurface media;
- Nature and extent of residual petroleum and chlorinated aliphatic hydrocarbon
 (CAH) contamination in soils;
- Total organic carbon (TOC) in select soil samples.
- Nature and extent of benzene, toluene, ethylbenzene, and xylenes (BTEX), trimethylbenzene (TMB), total petroleum hydrocarbon (TPH), and CAHs in groundwater;
- Concentrations of dissolved oxygen (DO), nitrate, ferrous iron, sulfate, methane, chloride, ammonia, and TOC in groundwater; and
- Temperature, specific conductance, reduction/oxidation (redox) potential, total alkalinity, and pH of groundwater.

An overview of these early 1995 site activities is presented in the following paragraphs. A more detailed discussion of field methods and study results is provided in the work plan (Parsons ES, 1995b). Study results are reviewed in Section 4 and detailed in the February 1997 closure plan (Parsons ES, 1997a).

Subsurface conditions at the site were characterized by the USACE and Parsons ES from February 21 through February 24, 1995, using a CPT/LIF device. Seventeen CPT pushes were performed to determine subsurface stratigraphy at the locations LIF-1 through LIF-17 (Figure 3.2). LIF was performed simultaneously at these locations to

evaluate the presence of residual or mobile (free-phase) hydrocarbons in the soil and groundwater. Final CPT logs are presented in Appendix D.

The CPT apparatus also was used to collect undisturbed soil samples at three monitoring point locations. Sample ES-SS-1 was collected from 10 to 10.7 feet bgs during the placement of monitoring point ESMP-14. Sample ES-SS-2 was collected from 14 to 14.7 feet bgs during the placement of ESMP-16. Sample ES-SS-3 was collected from 10 to 10.7 feet bgs adjacent to ESMP-12, following the placement of the monitoring point (Figure 3.1).

Thirty-four 0.5-inch inside-diameter (ID) groundwater monitoring points were installed at 17 locations in February to March 1995. Clusters of three monitoring points were installed at locations ESMP-8 and ESMP-14. Clusters of two monitoring points were installed at locations ESMP-1, ESMP-2, ESMP-3, ESMP-4, ESMP-5, ESMP-6, ESMP-7, ESMP-9, ESMP-10, ESMP-11, ESMP-13, ESMP-15, and ESMP-Single monitoring points were installed at locations ESMP-12 and ESMP-17. Where monitoring points were installed in clusters, the shallowest screen was placed across or just below the observed water table. Screens for deeper monitoring points within the same cluster were placed approximately 7 to 10 feet below the next shallowest point in the cluster. At the locations with paired monitoring point clusters, the point with the shallowest screened interval was designated with the suffix "S", while the point with the deeper screened interval was designated with the suffix "D". At the locations with three monitoring points, the point with the shallowest screened interval was designated with the suffix "S", the point with the intermediate screened interval was designated with the suffix "D", and the deepest point in the cluster was labeled "DD". All monitoring point locations are shown on Figure 3.2. Completion details for the monitoring points and the monitoring wells installed prior to 1997 are summarized in Table 3.2.

During the 1995 RNA investigation, groundwater samples were collected from 10 existing monitoring wells and from 31 of the 34 newly installed monitoring points.

Samples were not collected from wells MW-1 and MW-7, which were destroyed prior to the investigation during the 1995 UST removal (Ogden, 1995a), and from points ESMP-3S, ESMP-8D, and ESMP-11S, which produced an insufficient volume of groundwater. Groundwater was monitored for temperature and DO during purging.

Groundwater samples were analyzed in the field by USEPA personnel for pH, conductivity, redox potential, total alkalinity, hydrogen sulfide, ferrous iron, chloride, sulfate, nitrogen, carbon dioxide, and ammonia. Analyses for methane, ethene, fuel hydrocarbon compounds, and VOCs were performed at NRMRL in Ada, Oklahoma. Mobile, light non-aqueous-phase liquid (LNAPL) was encountered in monitoring well MW-5 near the former Building 551, and a sample of the product was collected for analysis of the mass fraction of BTEX. Analytical data collected during the RNA investigation are reviewed in Section 4 and presented in Appendix B.

In addition to the groundwater sampling, two rising head slug tests were performed on monitoring wells MW-4, MW-6, MW-9, and MW-12, all of which are screened in the UWBZ sand lenses (Figure 3.2 and Table 3.2). Hydraulic conductivities for the HWSA based on these slug tests are presented in Section 2.2.2.

3.1.3 IT Corporation 1995/1996 Groundwater Monitoring Events

Quarterly groundwater monitoring at the site commenced in August 1995 and has been performed since August 1995 by IT. Results from the August 1995 through December 1996 monitoring events are presented in the 1995 and 1996 Annual Groundwater Monitoring Reports (IT, 1997b and 1997a) and analytical data from these sampling events are provided in Appendix B. In addition, IT (1995b) collected groundwater samples from four temporary Geoprobe® points (560-01 through 560-04) during the August 1995 sampling event to evaluate the downgradient extent of CAH contamination near ESMP-17 (Figure 3.2). Data from this delineation sampling event also are presented in Appendix B. The nature and extent of VOC contamination in groundwater based on these and more recent sampling events are presented in Section 4.

3.2 PREVIOUS REMEDIATION ACTIVITIES

This section summarizes the actions that have been implemented to date at the HWSA to facilitate closure of the unit. As prescribed in the previous closure plan (Parsons ES, 1997a), to attain closure of contaminated soils and groundwater the following activities have been completed or are pending.

- Decontamination of Building 560 by cleaning the building and the drum wash pad (completed April 1996);
- Removal of the remaining four USTs (completed February 1995);
- Limited in situ remediation of organic soil contamination via passive or air injection bioventing (if necessary);
- Natural oxidation of residual dissolved BTEX and natural reductive dehalogenation of residual dissolved chlorinated VOCs (in progress);
- In situ remediation of residual dissolved chlorinated VOCs via groundwater amendment (passive or active oxygenation) (if necessary);
- Continued monitoring and site access controls as part of closure commitments (ongoing); and
- Exposure control by installation of taxiway (proposed future land use).

Additional assessment activities have been completed to optimize the final design of the closure approach based on the activities outlined above and on risk-based health standards developed for the HWSA. Decontamination of Building 560 and removal of the remaining four USTs are discussed below. The results of the additional assessment activities are presented in Section 4, and the development of site-specific health standards based on results of a risk assessment is presented in Section 5.

3.2.1 Summary of Decontamination Of Building 560

3.2.1.1 Building and Pad Decontamination Activities

Acids and spent desiccants were the most common types of waste stored in Building 560. The decontamination of Building 560 was completed as part of the IRP in April 1996. Details of the decontamination activities, including analytical results from post-decontamination sampling, are presented in a recent technical report (AFCEE, 1996a).

The decontamination plans for the building included:

- · Removing all items inside the building to facilitate decontamination activities;
- Vacuuming the building to remove dust, dirt, and debris;
- Washing the floor, walls, shelving of the building and the drum wash pad with a
 hot-water pressure washer, scrub brushes, and an all-purpose, household
 detergent; and
- · Triple rinsing all surfaces with hot water.

All wash and rinse water generated during decontamination activities were collected using a wet/dry vacuum and transferred to a skid-mounted storage tank for testing and disposal. After rinsing with the hot water, rinseate samples to be tested analytically to determine the effectiveness of the decontamination activities were collected. These rinseate samples were collected by pouring high-performance liquid chromatography (HPLC) water over the target media. The HPLC water was then collected for analysis. Until analytical results were received, the storage tank for the decontamination liquids was managed in compliance with all applicable hazardous waste tank requirements of the Ohio Administrative Code (OAC) 3745-66-90 through 3745-66-991.

3.2.1.2 Documenting Complete Decontamination

HPLC rinseate samples were submitted to a laboratory and analyzed using the methods listed in Table 3.3. Note that only representative analytes for each method

TABLE 3.3 HPLC RINSEATE ANALYTE LIST HAZARDOUS WASTE STORAGE AREA RICKENBACKER ANGB, OHIO

Compound	Reporting Limit (mg/L)
SW8240 - GC/M	S Volatile Organics
Acetone	0.8
Benzene	0.3
Chlorobenzene	0.1
Chloromethane	0.7
Chloroform	0.5
1,1-Dichloroethane	0.5
cis-1,2-Dichloroethene	0.2
trans-1,2-Dichloroethene	0.4
Dichloromethane	0.8
Ethylbenzene	0.3
Methyl ethyl ketone (2-butanone)	2
1,1,2,2-Tetrachloroethane	0.6
Tetrachloroethene	0.5
1,1,1-Trichloroethane	0.5
1,1,2-Trichloroethane	0.5
Trichloroethene	0.2
Toluene	0.6
Vinyl Chloride	0.6
	0.4
m/p-Xylenes	0.3
o-Xylene	Semivolatile Organics
	T 10
Acenaphthalene	10
Anthracene m, p, and o-Dichlorobenzene	10
	10
Flourene Fluoranthene	10
Hexachlorobenzene	10
2-Methylnaphthalene	10
Naphthalene	10
Phenanthrene	10
	10
Phenol	10
Pyrene	10
1,2,4-Trichlorobenzene	
	rine Pesticides and PCBs 0.02
4,4'-DDE	0.1
4,4'-DDT	0.02
Dieldrin	0.02
Endrin	
Heptachlor	0.02
PCB-1221	0.5
PCB-1232	1
Toxaphene	
	- Inorganics
Cadmium	0.5
Chromium	1 20
Cobalt	20
Lead	2
Nickel	20

Source: AFCEE, 1996a.

with their respective reporting limits (laboratory method detection limits [MDLs]) are provided. The building floor, walls, shelving, and drum wash pad were to be considered clean if the HPLC rinseate from the cleaning operation met the following standards:

- The federal public drinking water maximum contaminant level (MCL) as promulgated in 40 CFR 141.11 and OAC 3745-81-11 for inorganics and in 40 CFR 141.12 and OAC 3745-81-12 for organics;
- If an MCL was not available for a particular contaminant, then 15 times the federal MCL goal (MCLG), as promulgated in 40 CFR 141.50 was used as the clean standard; or
- If the product of 15 times the MCL or MCLG exceeds 1 milligram per liter (mg/L), or if neither an MCL nor an MCLG is available for a particular contaminant, 1 mg/L was used as the clean standard.
- If the MCL or MCLG was less than the contaminant's analytical detection limit using methods found in the USEPA (1986) SW846 document (*Test Methods for Evaluation Solid Waste: Physical/Chemical Methods*), 15 times the SW846 analytical detection limit was used as the clean standard.

Only a few of the analytes listed in Table 3.3 were detected in the HPLC rinseate samples at concentrations slightly above the laboratory MDL. No compounds were detected at concentrations above the "clean" standards listed above. Consequently, the decontamination of Building 560 and the concrete pad is considered complete. Analytical results and supporting documentation are provided elsewhere (AFCEE, 1996a).

3.2.1.3 Wastewater Management

All wash and rinse water generated during the Building 560/pad decontamination procedures were pumped into a temporary holding tank. A representative sample of

VOCs. SVOCs. analyzed for collected and containerized water was pesticides/polychlorinated biphenyls (PCBs), metals, and pH. Because none of the analytes were detected above the "clean" standards listed above, the containerized waste did not require management as a listed hazardous waste. On May 13, 1996, the AFBCA requested the City of Columbus to approve a 300-gallon wastewater discharge to the Columbus sewer system. After receiving approval from the City of Columbus on May 30, 1996, the AFBCA discharged the wastewater into the sanitary sewer located near Building 560 on June 4, 1996.

3.2.2 Underground Storage Tank Removal

The four former USTs (Tanks #47 through #50) located at the southwestern corner of the site (Figure 3.1) were removed by Ogden (1995a) on February 16 and 17, 1995. The following summary describes tasks performed during the tank removal process. All applicable BUSTR requirements were followed.

On February 16, 1995, the contents of the four 25,000-gallon USTs were inspected and sampled. Following sampling and prior to tank removal, the contents of the tanks were pumped out, transported offsite, and recycled. During the excavation process, a water line located near the western end of the USTs was damaged, resulting in flooding of the excavation. Two monitoring wells (MW-1 and MW-7) were also destroyed during the excavation process.

The water that had entered the UST excavation was treated onsite by Petro's mobile water treatment unit before it was discharged to the closest sanitary sewer inlet. The water pipe was disposed offsite on February 28, 1995, at the Athens-Hocking Reclamation Center. Soils surrounding the USTs were excavated and stockpiled to allow for tank removal. Excavated soils were visually examined and screened for VOC contamination using a photoionization detector (PID). Contaminated soil was segregated and placed in roll-off containers for analysis and proper disposal. Soil that was not contaminated (based on field screening) was returned to the excavation following UST removal. Concrete encountered in the excavation was placed in rolloff

containers for analysis and disposal. Following removal from the excavation, the USTs were decontaminated and transported offsite for recycling.

Two soil samples and one water sample were collected from the excavation following the removal of the USTs. One soil sample also was collected from each of the three rolloff containers in which the contaminated soil from the excavation was placed. After characterization of the soil, the contents of the rolloff containers were disposed of in an appropriate manner (Ogden, 1995a).

Following UST removal activities, the damaged water line was repaired, and the excavation was backfilled. Prior to backfilling, a layer of fine gravel was placed in the bottom of the excavation. Backfilling of the excavation was completed with the stockpiled soil from the excavation. The site was then compacted, graded, seeded, and covered with straw mulch. Complete details of the UST removal activities and analytical results for all samples collected during UST removal are presented by Ogden (1995a and 1995b).

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SECTION 4

NATURE AND EXTENT OF CONTAMINATION

The purpose of this section is to describe additional assessment activities required by the February 1997 amended closure/post-closure plan (Parsons ES, 1997a) and to define the nature and extent of site contamination based on characterization data collected to date. Preliminary results of source area treatability testing also are reviewed, and chemical fate in groundwater is evaluated and discussed.

To comply with the approved closure plan (Parsons ES, 1997a), additional groundwater monitoring wells were installed in an attempt to fully delineate the CAH plume east of the HWSA. Also, pilot-scale bioventing and air sparging systems were installed to evaluate the potential effectiveness of these technologies at remediating source area contamination in soil and groundwater. As part of these activities, additional environmental sampling/monitoring was completed, and these data, combined with previously collected site characterization data, were used to perform a comprehensive risk assessment and to discuss the development of the risk-based health standards for the site (Section 5). Additional details regarding monitoring well and bioventing/air sparging system placement and installation, sampling methodologies, and laboratory analytical requirements and procedures, are presented in the February 1997 amended closure/post-closure plan (Parsons ES, 1997a) and the June 1997 additional assessment activities work plan (Parsons ES, 1997b).

4.1 ENVIRONMENTAL SAMPLING AND MONITORING

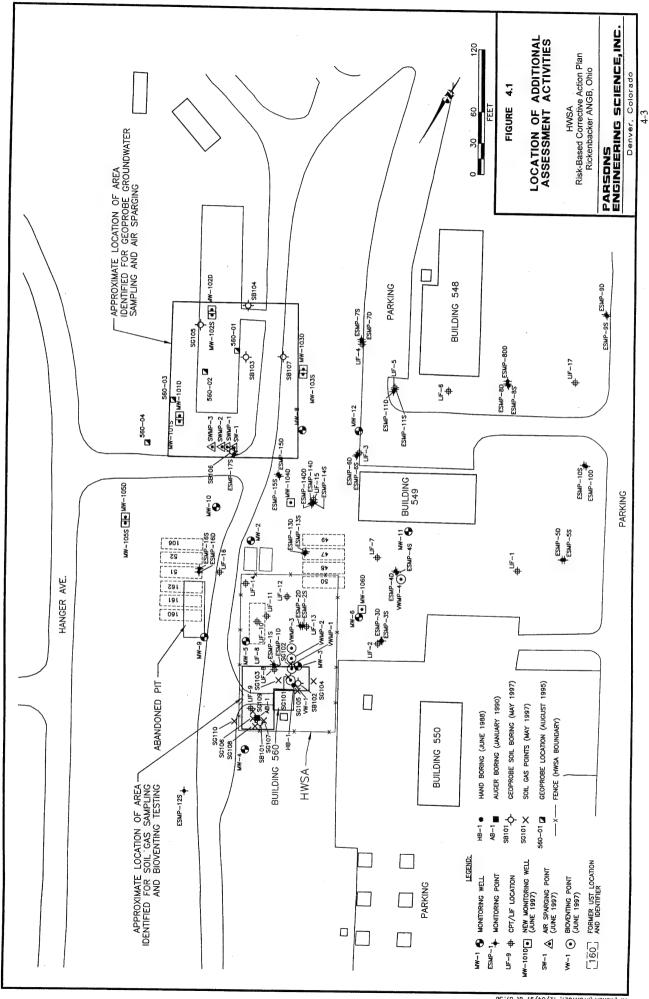
Environmental sampling and well installation activities were performed by IT to further clarify the nature and extent of soil and groundwater contamination at the HWSA and to provide the necessary wells and monitoring points for treatability tests (i.e., soil bioventing and groundwater oxygenation pilot tests). The sampling and well installation was conducted in two phases. The first phase, which was performed in May 1997, involved additional delineation of the extent of soil and groundwater contamination using Geoprobe® sampling techniques. The second phase was completed in June 1997, and involved soil sampling, monitoring well installation, and installation of treatability test system wells and monitoring points using rotosonic drilling methods.

4.1.1 Geoprobe® Investigations

Field assessment activities using Geoprobe® sampling techniques were performed by IT between 19 and 22 May 1997 at two locations. Soil and soil gas samples were collected near Building 560 to further delineate the extent of VOC soil contamination, and groundwater samples were collected approximately 150 feet east/southeast of the HWSA, near ESMP-17S, to further delineate groundwater contamination in this area. The locations of these sampling points are shown on Figure 4.1. Soil gas results from this sampling are not presented because relatively impermeable soils encountered during sampling caused excessive sample extraction vacuums and the likely dilution of vapor samples by atmospheric air leaking into the sampling train. This conclusion is confirmed by relatively high VOC concentrations in soil samples, and non-detect VOC concentrations in soil gas samples, which were collected from the same location and depth.

4.1.1.1 Source Area Soil Sampling

Soil samples were collected using the Geoprobe® at two locations which demonstrated significant petroleum hydrocarbon contamination during previous site investigations. In a June 1988 hand-boring sample (HB-1), collected approximately 10 feet southeast of Building 560 at 3 to 5 feet bgs, o-xylene was detected at 1,900 milligrams per kilogram (mg/kg) and ethylbenzene was detected at 120 mg/kg (Figure 4.1). Another soil sample collected in January 1990 from augered soil boring AB-1 (located approximately 20 feet northeast of Building 560) indicated lesser but still significant petroleum contamination at 8 to 10 feet bgs. At this location, total xylenes



were detected at 18 mg/kg and ethylbenzene was detected at 6.7 mg/kg. Based on these somewhat dated results, it was determined that additional sampling was warranted in this area. During the May 1997 sampling event, soil boring SB101 was installed near former soil boring AB-1, and boring SB102 was installed near former soil boring HB-1. The boring logs for SB101 and SB102 are provided in Appendix D. The soil samples were collected from 8 to 10 feet bgs in boring SB101 and from 3 to 5 feet bgs in boring SB102 using a Geoprobe® Macro-Core® sampler. The samples were shipped to Pace Analytical Services for analysis of VOCs by USEPA Method SW8260 and gasoline-range organics (GRO) by Method 8015M. The soil sample results are presented in Table 4.1.

Soil sample results from these two locations indicate that petroleum hydrocarbon contamination is still present in the vicinity of Building 560, especially near SB102. As compared to the non-detect VOC soil gas sample results from the vicinity of HB-1 and SB102 (SG101 through SG105) (Figure 4.1), the SB102 soil sample results indicate substantial hydrocarbon contamination in these vadose zone soils. A total BTEX concentration of 1,112 mg/kg was detected in this sample; the majority being xylenes. Comparatively, a total BTEX concentration of 4.2 mg/kg was detected in the 8 to 10 foot bgs sample from SB101. The sample results from SB101 are indicative of smear zone contamination near the water table in the vicinity of Building 560. The soil sample results from SB102 indicate that weathered petroleum contamination is still present in the vadose zone soils southeast of Building 560. These soil results suggest that oxygen should be utilized by fuel-degrading bacteria present in vadose zone soils and substantiated pilot-scale bioventing system installation and testing.

4.1.1.2 Groundwater Sampling Near ESMP-17S

Geoprobe® groundwater samples were collected east and southeast of ESMP-17S to delineate the downgradient extent of groundwater contamination. During the August 1995 sampling event, the only VOCs detected were 26 micrograms per liter (μ g/L) of vinyl chloride (VC), 27 μ g/L of cis-1,2-dicloroethene (DCE), and 4 μ g/L of trans-1,2-DCE at Geoprobe® location 560-01 (Figure 4.1). Therefore, the May 1997 Geoprobe®

TABLE 4.1 SOIL ANALYTICAL RESULTS **MAY 1997**

HAZARDOUS WASTE STORAGE AREA RICKENBACKER ANGB, OHIO

	Sample Locat	ion/Depth
Analyte	SB101 (8 to 10 ft bgs)	SB102 (3 to 5 ft bgs)
Volatile Organics (μg/kg) ^{a/}	4,200	_b/
Benzene	4,200	2,000
Toluene	690	170,000
Ethylbenzene	1,800	940,000
Xylenes	9,200	6,900
Sopropylbenzene	1,700	1,300
n-Propylbenzene	2,000	1,600
1,3,5-Trimethylbenzene	1,700	4,300
1,2,4-Trimethylbenzene	750	
sec-Butylbenzene	640	640
n-Butylbenzene	2,400	2,600
Naphthalene	460	_
1,2,3-Trichlorobenzene	400	
Gasoline Range Organics (mg/kg)c/	610	820

Note: Samples analyzed by USEPA Methods SW8260A and SW8015M.

μg/kg = micrograms per kilogram.
 - = not detected.

o' mg/kg = milligrams per kilogram.

sampling was focused in the vicinity of this sampling location. To facilitate groundwater sample collection, five borings were installed. Four borings (SB103, SB104, SB105, and SB107) were installed in the vicinity of 560-01, and one boring (SB106) was installed immediately adjacent to ESMP-17S. The soil boring/groundwater sampling locations are shown on Figure 4.1, and the boring logs are provided in Appendix D.

The groundwater samples were collected from 2-foot-long, stainless steel temporary well points. Groundwater samples were collected from two sand zones in four of the borings. Only one sand zone was encountered in boring SB104. The upper sand zone was encountered between depths of approximately 11 to 17 feet bgs. The second sand zone was encountered just above the gray silt and clay basal till, at depths of approximately 18 to 25 feet bgs. The temporary wells were sampled using a peristaltic pump and polyethylene tubing. The samples were analyzed for VOCs by USEPA Method SW8260 at the mobile laboratory.

Results from the May 1997 Geoprobe® groundwater sampling are summarized on Table 4.2. As expected, the VC and *cis*-1,2-DCE concentrations detected in the sample from the 14- to 16-foot-bgs sample at SB106 were similar to those recently measured at ESMP-17S, screened from approximately 13 to 16 feet bgs (IT, 1997a and 1997b). VOC contamination was not evident in the lower sand zone at any of the sampling locations. VOCs were detected in the upper sand zone sample collected at SB103, at generally lower concentrations than were detected in the May 1995 sample from this location (560-01). No other VOC contamination was evident in this area. These data were used to help determine the placement of the new downgradient monitoring wells required by the February 1997 amended closure/post-closure plan (Parsons ES, 1997a).

4.1.2 Well and Monitoring Point Installation and Sampling

In June 1997, the additional monitoring wells, the vent well (VW) and vapor monitoring points (MPs) for the bioventing system, and the sparge well (SW) and MPs

TABLE 4.2
GROUNDWATER ANALYTICAL RESULTS
MAY 1997
HAZARDOUS WASTE STORAGE AREA
RICKENBACKER ANGB, OHO

				Š	Sample ID/Depth	a /			
	SB103G001	SB103GO02	SB104GO01	SB105G001	SB105GO02	B103G002 SB104G001 SB105G001 SB105G002 SB106G001 SB106G002 SB107G001 SB107G000	SB106GO02	SB107GO01	SB107G002
	11.5 - 13.5	18 - 20	15 - 17	14 - 16	20 - 22	14 - 16	20 - 22	12 - 14	22 - 24
Volatile Organic Compounds (µg/L)b/									
Chloromethane	1.3	P	1	:	1	6.1	I	•	•
Vinyl Chloride	ı	ı	1	:	:	680E	1	1	1
1,1-Dichloroethene	1	;	ı	ı	1	 8	:	I	1
Methylene Chloride	1	;	1	:	ı	3.3	:	ı	:
trans-1,2-Dichloroethene	ì	1	:	:	ŀ	80	1	;	1
cis-1,2-Dichloroethene	5.3	1	1	:	:	1,900E	ı	:	ŀ
Trichloroethene	1.3	ı	1	1	•	:	:	1	ŀ
Naphthalene	4.0	1		1	;		•	1	•

Note: Samples analyzed by USEPA Method SW8260 at the mobile laboratory.

a/ Sample depth given in feet below ground surface.

b/ μg/L = Micrograms per liter.

c/ -- = Not detected.

d/ E = Concentration exceeded upper calibration standard; value shown represents an estimated concentration.

for the air sparging system were installed by IT. Soil samples were collected from the bioventing and air sparging system boreholes during drilling activities. Rotosonic drilling techniques were used to obtain a continuous core sample from each borehole and to minimize the generation of soil cuttings requiring disposal.

4.1.2.1 Installation of Additional Monitoring Wells

Four additional groundwater monitoring well clusters (MW-101S and D, MW-102S and D, MW-103S and D, and MW-105S and D) were installed downgradient from the HWSA. Two additional crossgradient monitoring wells MW-104D and MW-106D were installed south and southeast of the HWSA as part of the post-closure monitoring network (Figure 4.1). The well locations were determined based on the results of the May 1997 Geoprobe® groundwater sampling and mobile laboratory analysis.

The wells were constructed of 2-inch-diameter polyvinyl chloride (PVC) casing and screen. The lengths and depths of the screened sections was determined by the depths and thicknesses of the sand seams encountered. Wells were not screened across more than one sand zone. Well completion details for the newly installed groundwater monitoring wells are presented in Table 4.3. Following installation of the new wells in June 1997, IT measured groundwater elevations at all site monitoring wells and performed groundwater sampling. Table 4.4 presents the groundwater elevation data for this monitoring event, and Figure 4.2 illustrates the June 1997 potentiometric surface for the site. Groundwater sampling results from June/July 1997 are presented in Section 4.2.2.

The potentiometric surface shown on Figure 4.2 is based on groundwater elevations at ESMP-1D, -4D, -13S, and -17S, and MW-2. The hydraulic gradient between the HWSA and ESMP-17S appears to be consistent with previous site measurements presented in Section 2.2.2 (approximately 0.005 ft/ft). Considering the well installation depths listed in Table 4.3, monitoring wells MW-101S, MW-102S, MW-103S, and MW-105S appear to have been installed in the same sand zone as ESMP-17S. However, the high groundwater elevations at MW-102S and MW-103S

S/CP/1.xls/Table 4.3

JUNE 1997 MONITORING WELL COMPLETION DETAILS
HAZARDOUS WASTE STORAGE AREA
RICKENBACKER ANGB, OHIO

			Datum	Ground	Depth to Top	Depth to Bottom	Screen	Screen	Riser
			Elevation	Elevation	of Screen	of Screen	Length	Diameter	Diameter
Well ID	Easting	Northing	(ft msl) ^{a/}	(ft msl)	(feet bgs) ^{b/}	(feet bgs)	(feet)	(inches)	(inches)
		X							
MW-101S	1845251	662509	739.34	739.47	8.5	13.5	5	7	2
MW-101D	1845254	662506	739.41	739.52	16.5	21.5	5	2	2
MW-107S	1845299	662415	739.50	739.61	8.5	13.5	5	2	2
MW-102D	1845302	662413	739.52	739.74	16.75	21.75	5	2	2
MW-103S	1845197	662394	740.15	740.32	10	15	2	2	7
MW-103D	1845199	662391	740.30	740.36	18	23	2	2	7
MW-104D	1845118	662492	740.88	741.04	19	24	2	2	7
MW-105S	1845218	662617	739.10	739.16	6	14	5	7	5
MW-105D	1845222	662613	738.99	739.18	18	23	5	7	5
MW-106D	1844997	662516	741.37	741.52	17.5	22.5	5	2	2

 a^{\prime} ft msl = feet above mean sea level.

b' bgs = below ground surface.

TABLE 4.4 GROUNDWATER ELEVATION DATA JUNE 17, 1997 HAZARDOUS WASTE STORAGE AREA RICKENBACKER ANGB, OHIO

	Datum	Depth to	
	Elevation	Water	Elevation
Well ID	(ft msl)a/	(ft) ^{b/}	(ft msl)
MW-02	743.36	10.70	732.66
MW-03	743.96	10.02	733.94
MW-04	745.15	10.57	734.58
MW-05	744.97	12.15	732.82
MW-06	745.18	12.37	732.81
MW-08	743.89	7.82	736.07
MW-09	745.25	7.93	737.32
MW-10	742.64	9.75	732.89
MW-11	744.15	10.92	733.23
MW-12	743.02	9.42	733.60
MW-101S	739.34	8.92	730.42
MW-101D	739.41	6.39	733.02
MW-102S	739.50	6.04	733.46
MW-102D	739.52	1.72	737.80
MW-103S	740.15	5.98	734.17
MW-103D	740.30	7.80	732.50
MW-104D	740.88	8.51	732.37
MW-105S	739.10	7.83	731.27
MW-105D	738.99	3.18	735.81
MW-106D	741.37	8.87	732.50
ESMP-1S	741.67	6.99	734.68
ESMP-1D	741.72	8.68	733.04
ESMP-2S	741.18	3.35	737.83
ESMP-2D	741.29	8.29	733.00
ESMP-3S	742.23	8.25	733.98
ESMP-3D	742.22	9.23	732.99
ESMP-4S	742.70	9.34	733.36
ESMP-4D	742.69	9.55	733.14
ESMP-5S	741.51	4.17	737.34
ESMP-5D	741.56	6.99	734.57
ESMP-6S	740.98	8.30	732.68
ESMP-6D	741.05	8.15	732.90
ESMP-7S	740.85	4.76	736.09
ESMP-7D	740.80	8.76	732.04

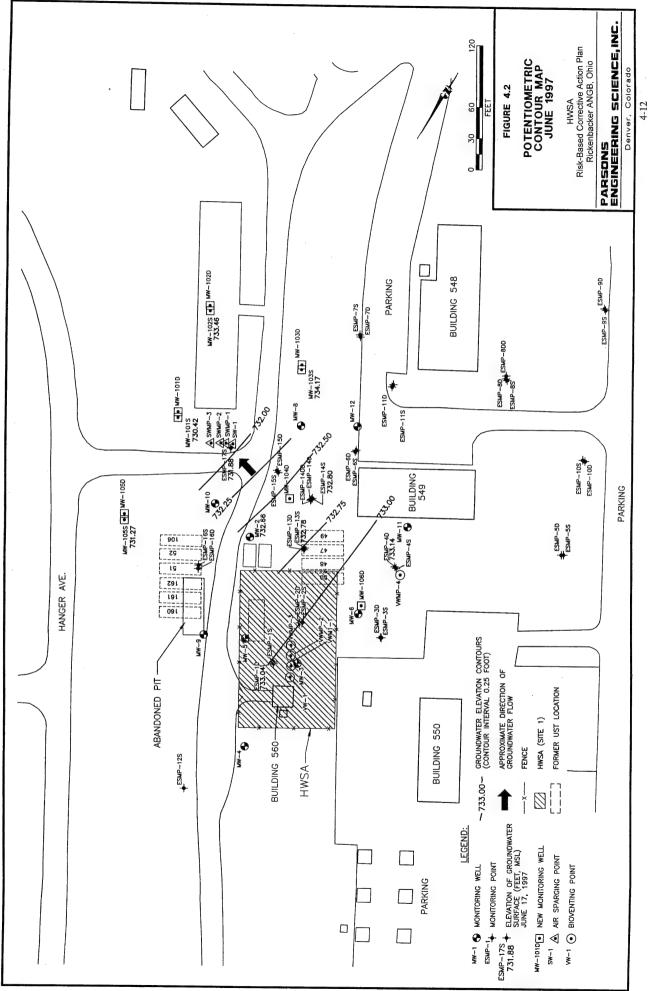
TABLE 4.4 (Continued) GROUNDWATER ELEVATION DATA JUNE 17, 1997

HAZARDOUS WASTE STORAGE AREA RICKENBACKER ANGB, OHIO

	Determ	Donth to	
	Datum	Depth to	Elevation
	Elevation	Water	
Well ID	(ft msl) ^{a/}	(ft) ^{b/}	(ft msl)
ESMP-8S	740.92	2.81	738.11
ESMP-8D	740.89	6.86	734.03
ESMP-8DD	740.83	7.74	733.09
ESMP-9S	741.79	3.65	738.14
ESMP-9D	741.70	7.37	734.33
ESMP-10S	741.56	7.20	734.36
ESMP-10D	741.54	7.98	733.56
ESMP-11S	740.76	6.88	733.88
ESMP-11D	740.80	8.70	732.10
ESMP-12S	742.43	1.68	740.75
ESMP-13S	741.38	8.60	732.78
ESMP-13D	741.38	8.58	732.80
ESMP-14S	741.17	8.37	732.80
ESMP-14D	741.18	8.34	732.84
ESMP-14DD	741.13	8.22	732.91
ESMP-15S	740.37	7.77	732.60
ESMP-15D	740.28	7.44	732.84
ESMP-16S	740.33	7.47	732.86
ESMP-16D	740.33	7.31	733.02
ESMP-17S	739.87	7.99	731.88

a' ft msl = feet above mean sea level.

b/ Feet below top of casing.



and the relatively low groundwater elevations at MW-101S and MW-105S (Table 4.4 and Figure 4.2) suggest that the groundwater levels in these newly installed wells may not have sufficiently stabilized prior to initial measurement of the groundwater elevations and may not be representative of the actual potentiometric surface elevation in this sand zone. Subsequent quarterly monitoring events will be necessary to establish the potentiometric surface in the vicinity of these new monitoring wells and verify the hydraulic gradient.

4.1.2.2 Installation of Bioventing Wells and Monitoring Points

Based on the elevated hydrocarbon concentrations detected in soil sample SB102 in May 1997 (Section 4.1.1.2), a bioventing VW (VW-1) and three multi-depth MPs (VWMP-1, VWMP-2, and VWMP-3) were installed on the southeast site of Building 560 for treatability testing. A fourth background MP (VWMP-4) was installed near ESMP-4S and -4D, south of Building 560. The VW and MP locations and completion details are presented in Figure 4.1 and Table 4.5, respectively. A cross section of the bioventing system configuration is shown on Figure 4.3.

The VW screen was placed from approximately 3.4 to 8.4 feet bgs for air injection treatment of vadose zone soils. The casing and screen were constructed with 2-inch-diameter Schedule 40 PVC. A 0.04-inch slotted screen was installed. The MPs were constructed with 6-inch-long sections of 1-inch-diameter Schedule 40 PVC, 0.020-inch slotted well screen, and 0.25-inch Schedule 80 PVC risers. Additional details on the bioventing system are presented in the June 1997 work plan (Parsons ES, 1997b) and the treatability testing results report (IT, 1997c).

During rotosonic drilling of the bioventing system boreholes, soil samples were collected by IT and sent to Pace Analytical Services for laboratory analysis. Soil samples were analyzed for VOCs, SVOCs, and TPH (gasoline- and diesel-range organics) by USEPA Methods SW8260, SW8270B, SW8015M, respectively. In addition, soil samples were analyzed for target-analyte-list (TAL) metals, ammonia, ortho-phosphate, pH, and percent moisture. Soil analytical data are presented in

6/CP/1.xls/Table 4.5

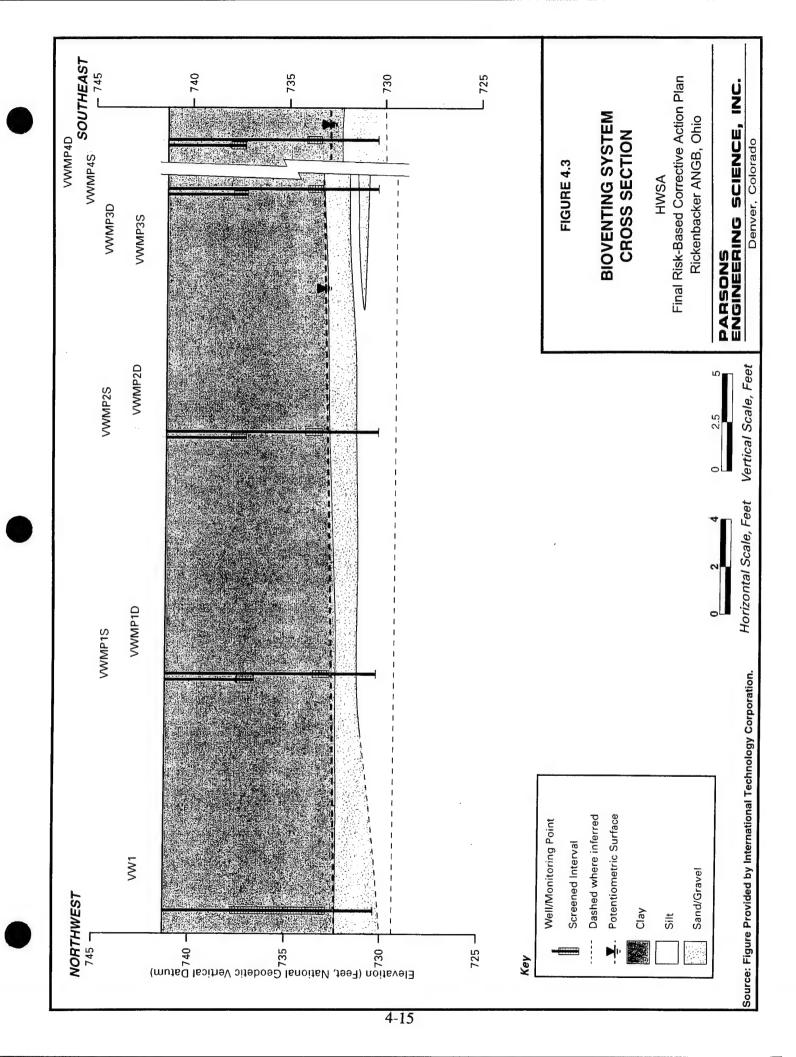
TABLE 4.5
BIOVENTING SYSTEM COMPLETION DETAILS
HAZARDOUS WASTE STORAGE AREA
RICKENBACKER ANGB, OHIO

			Ground	Denth to Top	Depth to Bottom	Screen	Screen	Riser
			Flevation	of Screen		Length	Diameter	Diameter
91	Doction	Northing	(fr msl) ^{a/}	(feet bgs) ^{b/}		(feet)	(inches)	(inches)
well ID	Casting	1101 min	(10.00)					
3	0007707	613633	741 65	3.4	4.8	2	2	2
1-M^	1844998	510700	711.03	31.0	V .	0.75	_	0.25
VWMP-1S	1845005	909799	/41.33	5.73			٠ +	30.0
WWWP-1D	1845005	909299	741.53	7.75	8.5	0.75	-	0.23
OF GIVEN	1045013	005699	740 95	3.25	4	0.75		0.25
VWMF-2S	1842012	00700	20.011	30.0	0	0.75	_	0.25
VWMP-2D	1845012	662299	740.95	7.72	٥	C	٠,	0.05
OC CLANA	1945018	662699	740.91	3.25	4	0.75		0.73
V WIVIE-33	1043010	20200	140.01	7 25	«	0.75	_	0.25
VWMP-3D	1845018	760700	140.91	67.1) 1	31.0	-	20 0
VWMP-4S	1844994	662444	742.43	3.25	4	0.73	,	6.00
Ch Charach	1944004	AK2444	742.43	7.25	∞	0.75	_	0.72
V W MIT-4D	1044224	11700						

٠,

a' ft msl = feet above mean sea level.

b' bgs = below ground surface.



Appendix C. The inorganic and organic soil sample results are summarized in Tables 4.6 and 4.7, respectively. Soil results from this sampling event and previous sampling events at the HWSA are discussed in Section 4.2.

4.1.2.3 Installation of Air Sparging Wells

Based on historical elevated concentrations of VOCs in groundwater near ESMP-17S, one air sparging well (SW-1) and three monitoring wells (SWMP-1, SWMP-2, and SWMP-3) were installed in this area (Figure 4.1). Each of the wells was drilled using rotosonic drilling techniques. At each well location, a vadose zone screened interval (VZMP-1, VZMP-2, VZMP-3, and VZMP-4), similar to the bioventing system MPs, was installed to evaluate diffusion of oxygen from the saturated sand zone treatment area into the overlying clay. The vadose zone MPs were installed approximately 3 to 4 feet bgs. One or two screened intervals were installed within the saturated sand zone between 10 and 16 feet bgs. These intervals were completed with 6-inch to 2-foot lengths of 2-inch-diameter, 0.02-inch slotted PVC screen.

The air sparging system completion details are presented in Table 4.8, and a cross section of the system configuration is shown in Figure 4.4. Additional details on the air sparging system are presented in the June 1997 work plan (Parsons ES, 1997b) and the treatability testing results report (IT, 1997c). Soil samples also were collected during installation of the air sparging wells, and the analytical results are provided in Appendix C.

4.2 NATURE AND EXTENT OF CONTAMINATION

This section briefly summarizes the nature and extent of VOCs, SVOCs, and metals in site soils and groundwater. Soil contaminant data are summarized from site investigations conducted between June 1988 and June 1997. Groundwater contaminant data are primarily based on results obtained during the June/July 1997 comprehensive groundwater sampling performed by IT; however, results from previous sampling events also are provided for comparison. Emphasis is given to describing the extent of dissolved VOC contamination in groundwater, as VOCs generally are more mobile than

JUNE 1997 HAZARDOUS WASTE STORAGE AREA RICKENBACKER ANGB, OHIO TABLE 4.6 METALS IN SOIL

					SAMPI	SAMPLE ID/SAMPLE D	DEPTH				
ANALYTE	VW1S001	VWMP1S001	VWMP1S002	VWMP1S052	VWMP2S001	VWMP2S002	VWMP3S001	VWMP3S002	VWMP45001	VWMP4S051	VW/MP4S002
(me/ke)"	A6-8	3.5-4.5	7.5-8.5	7.5-8.5	34	7-8	3-4	7-8	3-4	3-4	7-8
Aluminum	5,800	10,000	7,400	5,900	10,000	7,000	8,000	2,900	14,000	18,000	6,300
Antimony	3	:	:	:	:	:	;	:	:	:	:
Amenic	6.6	9.5	10	8.2	9.1	6.6	4.0	10	==	9.5	90 90
Barium	57	*	65	89	82	49	98	57	130	061	Z
Beryllium	0.49	0.71	0.52	0.46	0.65	0.52	0.50	0.44	6.9	1.2	0.45
Cadmium	:	:	:	:	1.1	:	0.88	:	:	:	:
Calcium	63,000	32,000	73,000	78,000	18,000	53,000	31,000	71,000	5,200	6,300	86,000
Chromium	. ac	7		9.4	13	10	=======================================	9.1	61	22	10
Cobalt	11	13	12	12	11	11	12	13	17	16	13
Copper	24	27	56	22	24	56	21	56	35	30	23
Iron	25,000	27,000	23,000	21,000	26,000	24,000	22,000	24,000	35,000	37,000	22,000
Lead	6.4	8.1	8.9	6.1	8.0	6.4	8.7	7.0	7.8	7.9	6.2
Magnesium	21,000	11,000	22,000	24,000	7,400	23,000	9,400	24,000	3,800	4,500	24,000
Mangancac	280	400	290	390	320	230	390	310	260	640	340
Mercury	0.033	0.027	1	0.025	0.032	0.026	0.027	0.030	0.043	0.033	0.026
Nickel	31	31	31	28	26	32	21	36	48	41	30
Potassium	1,200	1,200	1,900	1,400	950	1,500	1,000	1,300	1,700	1,400	1,600
Selenium	0.97	29.0	0.82	:	1.9	0.63	0.88	:	1.1	0.87	1.2
Thellium	2.5	2.6	1.7	1.7	2.6	2.1	2.3	2.8	2.9	3.3	2.4
Silver	:	;	:	•	•	:	:	:	:	:	:
Sodium	95	95	130	120	83	110	83	120	71	69	120
Vanadium	61	56	23	19	25	22	77	18	37	38	19
Zinc	110	68	92	98	76	8	99	110	110	110	81

" mg/kg = milligram per kilogram.

V Sample depth in feet below ground surface.

"Represents a sample duplicate for the sample depth interval shown.

" = not detected.

022/731 296/CP/1.XLS/Table 4.7

ORGANICS DETECTED IN SOIL JUNE 1997 HAZARDOUS WASTE STORAGE AREA RICKENBACKER ANGB, OHIO TABLE 4.7

					SAMP	SAMPLE ID/SAMPLE DEPTH	EPTH				
ANALYTE	VW15001 8-9a/	3.5-4.5	VWMP15002 7.5-8.5	VWMP1S052b/ VWMP2S001 7.5-8.5 3-4	VWMP25001 3-4	VWMP2S002 7-8	3-4	VWMP3S002 7-8	VWMP45001 3-4	VWMP4S002 7-8	VWMP4S002 VWMP4S051b/ 7-8 3-4
Semivolatiles (mg/kg) c/											
USEPA Method SW8270B											
Anthracene	- q	t	1	ı	1	ı	0.042 Jo/	ı	ı	ı	ı
Benzo (a) anthracene	1	ı	ı	1	ı	ı	0.092 J	ı	1	ı	1
Benzo (b) fluoranthene	1	ı	ı	ı	1	1	0.11 J	1	ı	1	ı
Benzo (k) fluoranthene	ı	ı	ı	1	1	:	0.058 J	1	ı	:	ı
Benzo (a) pyrene	ı	1	ı	ı	ı	1	0.092 J	1	ı	ı	1
Benzo (g,h,i) pyrene	1	ı	ı	1	ı	1	0.047 J	1	ı	1	:
Bis (2-ethylbexyl) phthalate	ı	1	1.1	1	1	1	1	1		ı	1
Свлужеве	ı	1	ŧ	ı	ı	1	0.11.9	ı	1	1	1
Fluoranthene	ı	1	ı	ı	ı	1	0.33 J	1	1	1	1
Indeno (1,2,3-cd) perylene	ı	ı	1	ı	t	ı	0.060 J	1	1	ı	ı
2-Methylraphthalene	3.5 El/	ı	1	t	1	1	1	ı	ı	ı	ı
Naphthalene	1.6	1	1	1	ı	ı	ı	ı	ı	1	ι
Phenanthreno	ı	ı	ı	1	ı	ı	0.24 J	ı	1	1	1
Pyrene	ı	ı	ı		ı	1	0.22 J	ı	1	1	ı
Volatiles (mg/kg)											
Acetone	0.043 Bg/	7.1 B	7.6 B	0.042Db/B	0.021 DJB	0.036 DB	6.0 B	0.039 Df/B	0.0086 B	ı	0.0081 B
2-Butanone	0.011 B	63 EB	1	1	1	0.0039 J	0.26 JB	0.0020 J	0.011	1	0.0044 J
1,2-Dichloroethano		ı	ì	0.032 D	1	0.0026 JB	ı	1	1	1	0.0027 JB
Ethylbenzene	0.86 E	0.54 J	1	0.0025 J	0.0018 J	1	2.4	1	0.0024 J	ı	0.0018 J
4-Methyl-2-pentanone	ı		ı	1	ı	ı	ı	1	0.0000	1	0.0020 J
Methylene Chloride	1	2.7 B	2.5 B	1	1	1	0.78 B	1	1	0.0022 J	ı
Styrene	0.036	1	ı	1	ı	1		1	:	ı	1
Tetrachloroethens	0.0026 J	:	ı	1	,	1	:	1	1	ı	ı
Tolucne	0.13	1	ı	0.011 DJ	ı	0.015 DJ	1	1	1	ı	ı
Trichloroethene	0.0024 J	ı	1	1	1	ı	1	1	1	1	1
Xylenes	3.4 EB	1.1	ı	0.0082 B	0.0032 JB	0.0048 JB	12	0.0040 JB	0.0087 B	0.0061 B	0.0068 B
USEPA Method SW 8015	98	1	1	1	1	1	1	ļ	1	1	1
Catomic Names Organica (mg/12)	300		•		•	1	•	ı	1	!	
Diesel Range Organics (mg/kg)	230	ı	ı	ı,	ı	ı	ı	ı	ı	ı	1

s/ Sample depth in feet below ground surface.
b/ Represents a sample duplicate for the sample depth interval shown.
c/ mg/kg = milligram per kilogram.
d/ = not detected.
d/ = indicates an estimated value. The compound was detected, but was below the laboratory reporting limit.
f/ E = concentration exceeded the upper calibration standard; value shown represents an estimated concentration.
g/ B = analyte found in sasociated blank as well as in sample.
b/ D = compound analyzed at a secondary dilution factor.

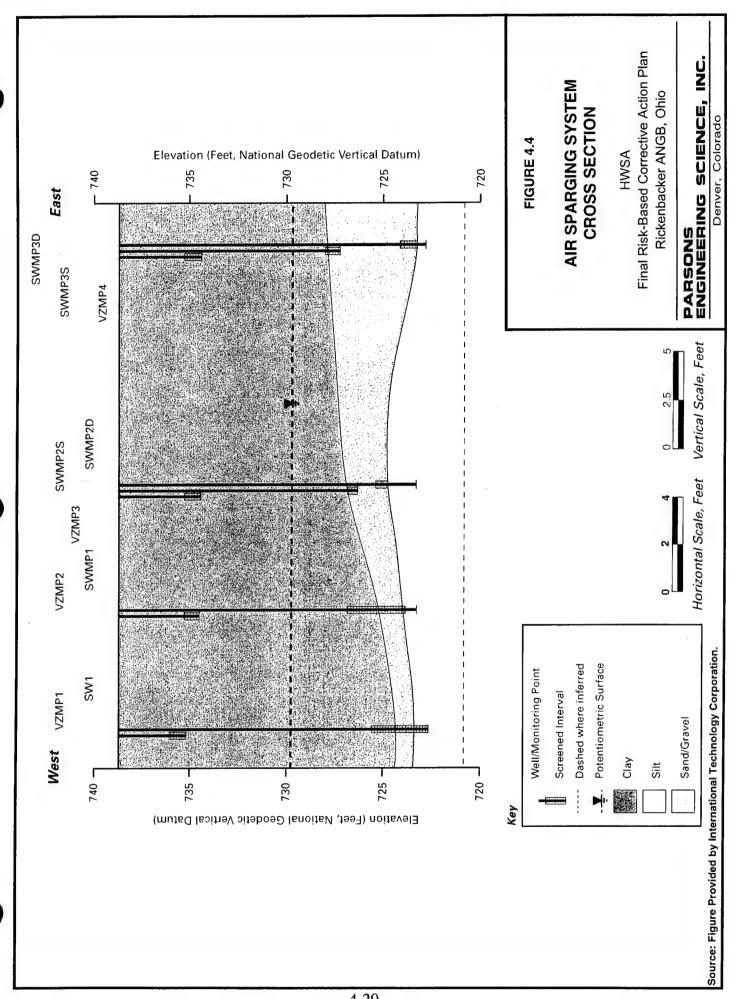
AIR SPARGING SYSTEM COMPLETION DETAILS
HAZARDOUS WASTE STORAGE AREA
RICKENBACKER ANGB, OHIO

			Datum	Ground	Depth to Top	Depth to Bottom	Screen	Screen	Riser
Well ID	Easting	Northing	(ft msl) ^{2/}	(ft msl)	(feet bgs) ^{b/}	(feet bgs)	(feet)	(inches)	(inches)
SW-1	1845196	662492	730 50	730 81	13.7	15.3	c	,	,
SWVZ-1	1845196	267200	NAC	730 81	2.C1	2.5	2 0	7 -	7
SWMP-1	1845200	662495	739 31	739.72	12.73	5.5	ر. ر	- c	0.23
SWVZ-2	1845200	662495	NA	739.72	3.5	4 25	0.75	۷ -	200
SWMP-2S	1845203	662498	739.53	739.80	12.08	12.58	5.5	, ,	ر2:0
SWMP-2D	1845203	662498	739.38	739.80	13.5	14	0.5	1 6	1 0
SWVZ-3	1845203	662498	NA	739.80	3.5	4.25	0.75	ı	0.25
SWMP-3S	1845210	662505	739.59	739.74	10.83	11.33	0.5	2	2
SWMP-3D	1845210	662505	739.41	739.74	14.75	15.25	0.5	2	2
SWVZ-4	1845210	662505	NA	739.74	3.5	4.25	0.75	1	0.25

a' ft msl = feet above mean sea level.

^{b'} bgs = below ground surface.

c' NA = not applicable.



the other contaminants (i.e., metals and SVOCs), and better represent the areal extent of groundwater contamination in the vicinity of the HWSA.

4.2.1 Soil Quality

Other than the May and June 1997 soil sampling performed during the bioventing system installation (Figure 4.1), the majority of soil sampling at the HWSA primarily was performed during early site investigations (June 1988 through October 1991). Soil samples collected during these investigations typically were analyzed for VOCs, SVOCs, and metals, although not all soil samples were analyzed for all parameters. The surface soil samples collected from the sampling grid (designated "SU" and "SS") were analyzed for SVOCs and metals. The auger boring samples, hand boring samples, and the soil samples from monitoring well borings (designated "AB", "HB", and "MW") were analyzed for VOCs, SVOCs, and metals. Soil sampling locations and analytical results for sampling conducted prior to 1997 are presented on Figure 3.1 and in Appendix B, respectively. The 1997 sampling locations and analytical results are presented on Figure 4.1 and in Appendix C, respectively. Soil analytical results are summarized in the following sections for three depth intervals: 0 to 2 feet, 3 to 5 feet, and 7 to 10 feet bgs. Significant soil sampling results from UST removal and closure activities also are presented. Maximum soil contaminant concentrations for soils between 0 and 10 feet bgs, based on the 1997 and all previous sampling events, are provided in Appendix E. Soil results from depths greater than 10 feet bgs are included in the analytical data provided in Appendix B.

4.2.1.1 Depth Interval 0 to 2 Feet

Near-surface soil samples that have been analyzed for VOCs at the HWSA include samples from borings HB-1 through HB-6 (collected in June 1988) and from MW-2 (collected in August 1988). The highest VOC detections were o-xylene in HB-1 at 440 mg/kg, o-xylene in HB-2 at 43 micrograms per kilogram (μ g/kg), and methylene chloride in the sample from MW-2 at 5 μ g/kg.

Sixty-one near-surface soil samples (0 to 2 feet bgs) have been collected at, or in the vicinity of, the HWSA and analyzed for metals since June 1988. Based on these sample results, cadmium is the metal most often detected at, or above, the 95-percent upper confidence limit (UCL) of the mean background concentration of 0.77 mg/kg (IT, 1997d) (Appendix E). Results from 22 surface soil samples collected in and around the HWSA indicate cadmium concentrations greater than 0.77 mg/kg, and the maximum cadmium concentration (9.1 mg/kg) was detected in a composite soil sample from SU-9 and SU-10 (Figure 3.1). Cadmium concentrations appear to be greatest in soil samples collected in the southwestern two-thirds of the site (SU-5 through SU-16 and HB-1 and HB-2). Chromium was detected in 10 samples at concentrations greater than its background level (18.8 mg/kg), with a maximum concentration of 25.6 mg/kg (SU-5 and SU-6 composite sample). Other metals detected above background levels in the 0-to 2-foot interval include arsenic (42 mg/kg at HB-4), beryllium (1.8 mg/kg at SU-20), and silver (7.2 mg/kg at SU-26, 3 mg/kg at SS-1, and 1.8 mg/kg at SU-46).

Several SVOCs were detected in surface soil samples collected within and near the HWSA. Based on the initial comparison criteria presented in the June 1997 work plan (Parsons ES, 1997b), benzo(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, and indeno(1,2,3-cd)pyrene were the four SVOCs most frequently detected at potentially significant concentrations. The highest concentrations of SVOC contamination within the HWSA boundary appear to be on the southern side of Building 560 and its driveway near SU-5 and SU-6 (Figure 3.1). In a composite sample collected from these two locations in June 1988, benzo(a)anthracene was detected at 2.1 mg/kg, benzo(a)pyrene was detected at 2.6 mg/kg, benzo(b)fluoranthene was detected at 3.2 mg/kg, and indeno(1,2,3-cd)pyrene was detected at 1.7 mg/kg. It should be noted, however, that SVOC contamination has been detected outside the fence (outside the HWSA) near the western corner of the site at levels far exceeding concentrations detected within the HWSA. In a January 1990 sample collected from SU-45 (approximately 15 feet outside the HWSA), benzo(a)anthracene was detected at 15 mg/kg, benzo(a)pyrene was detected at 15 mg/kg, benzo(b)fluoranthene was detected at 14 mg/kg, and

indeno(1,2,3-cd)pyrene was detected at 10 mg/kg. Benzo(k)fluoranthene, chrysene, and dibenz(a,h)anthracene also have been detected at elevated concentrations both within and outside the HWSA, and naphthalene was detected at 5.4 mg/kg from the soil sample collected at HB-2 (Figure 3.1).

4.2.1.2 Depth Interval 3 to 5 feet

Twenty-four soil samples have been collected from the 3- to 5-foot depth interval within and around the HWSA. VOCs have been detected in 10 of these 24 samples. The VOCs detected at the highest concentrations in samples collected prior to 1997 were from HB-1, where ethylbenzene was detected at 120 mg/kg and o-xylene was detected at 1,900 mg/kg. The VOCs detected in the May 1997 and June 1997 sampling events are listed in Tables 4.1 and 4.7, respectively. The highest VOC detections during the June 1997 sampling event were from VWMP-3, where xylenes were detected at 12 mg/kg and ethylbenzene was detected at 2.4 mg/kg. At VWMP-1, xylenes were detected at 1.1 mg/kg, and 2-butanone was detected at 63 mg/kg; however, 2-butanone also was detected in the associated sample blank.

Twenty-three soil samples collected from the 3- to 5-foot depth interval have been analyzed for metals. From samples collected within the HWSA prior to 1997, arsenic was detected at 29 mg/kg at HB-4, cadmium was detected at 3.3 mg/kg at AB-6, and thallium was detected at 1.1 mg/kg and 1.2 mg/kg at AB-1 and AB-2, respectively. Elevated metals detections also have occurred in samples collected outside of the HWSA. Arsenic was detected at 20.7 mg/kg in the MW-12 borehole located approximately 140 feet south of the HWSA. Beryllium and chromium were detected at 1 mg/kg and 28.6 mg/kg, respectively, at MW-10, which is located approximately 70 feet southeast of the HWSA (Figure 3.1). For soil samples collected in June 1997, metal detections were most significant at VWMP-4 in the 3- to 4-foot depth interval (Table 4.6). At VWMP-4, barium was detected at 190 mg/kg, beryllium was detected at 1.2 mg/kg, manganese was detected at 640 mg/kg, and vanadium was detected at 38 mg/kg.

Since June 1988, 23 soil samples collected within and around the HWSA from the 3-to 5-foot depth interval have been analyzed for SVOCs. SVOCs were not detected in 18 of these soil samples. At HB-1, HB-2, and AB-2, naphthalene concentrations of 1,600 μ g/kg, 570 μ g/kg, and 1,200 μ g/kg, respectively, were detected. Also, benzo(b)fluoranthene was detected at 170 μ g/kg in the soil sample collected at AB-4.

4.2.1.3 Depth Interval 7 to 10 Feet

VOCs have been detected in 13 of the 25 soil samples collected from the 7- to 10-foot depth interval since June 1988. Pre-1997 sample results from within the HWSA are as follow: AB-1 contained ethylbenzene at 6.7 mg/kg, *m/p*-xylene at 6.0 mg/kg, and *o*-xylene at 12 mg/kg; AB-3 contained benzene at 39 μg/kg; AB-4 contained acetone at 250 μg/kg, *o*-xylene at 51 μg/kg, *m/p*-xylene at 36 μg/kg, and ethylbenzene at 20 μg/kg; AB-6 contained benzene at 1 μg/kg; AB-14 contained benzene at 15 mg/kg, ethylbenzene at 15 mg/kg, *m/p*-xylene at 15 mg/kg, and *o*-xylene at 27 mg/kg. From the MW-7 borehole, formerly located adjacent to the four USTs at the southern corner of the site (Figure 3.1), benzene was detected at 2.1 mg/kg, ethylbenzene was detected at 980 μg/kg, m/p-xylene at 1.2 mg/kg, *o*-xylene was detected at 1.2 mg/kg, and 1,1,1-trichloroethane (1,1,1-TCA) was detected at 86 μg/kg. In the June 1997 soil samples, xylenes were detected at 3.4 mg/kg, GRO were detected at 300 mg/kg, and diesel-range organics (DRO) were detected at 270 mg/kg at the VW-1 borehole (Figure 4.1).

Twenty-three soil samples collected at the 8- to 10-foot depth interval in the HWSA vicinity have been analyzed for metals and SVOCs. Elevated metal detections from within the HWSA include arsenic (26 mg/kg) from the MW-5 borehole sample and thallium (10.5 mg/kg) from the MW-3 borehole sample. SVOCs were not detected in 17 of the 23 samples. Naphthalene was detected at 1.8 mg/kg at AB-2 and 1.6 mg/kg at VW-1. 2-Methylnaphthalene and bis(2-ethylhexyl)phthalate also were detected in the June 1997 samples from VW-1 and VWMP-1, respectively (Table 4.7).

4.2.1.4 UST Closure Soil Sampling

Soil samples were collected as part of the UST closure activities conducted in 1994 through 1996 for the 15 former USTs (Tanks #47 through #57, #106, #160, #161, and #162) located near the HWSA (Ogden, 1995a and 1995b). Results for cis-1,2-DCE (5.8 mg/kg), trans-1,2-DCE (570 μ g/kg), 1,1,2,2-tetrachloroethene (6.3 μ g/kg), trichloroethene (TCE) (2.6 mg/kg), and VC (1.3 μ g/kg) all represent the maximum detected site soil concentrations for these compounds. The two soil samples represented by these values (samples HWSA-TP48-2W and HWSA-TP48-2WD) were collected from the western end of the excavation for former Tank #48, at the bottom of the pit (Figure 3.1).

4.2.1.5 Soil Contamination Summary

VOCs, SVOCs, and metals all have been detected in vadose zone soils within the HWSA. SVOCs and metals also have been detected in soils at elevated concentrations outside the HWSA. VOC contamination, while present to some degree throughout the HWSA, appears to be most significant at 3 to 5 feet bgs near borings HB-1 and SB102 (Figure 4.1 and Table 4.1). At this location, xylenes were detected at high levels both in 1988 (1,900 mg/kg at HB-1) and in 1997 (940 mg/kg at SB102). However, VOC concentrations at this same depth interval are significantly less at the bioventing MP boreholes installed approximately 10 to 30 feet from this area (Figure 4.1 and Table 4.7). At VW-1, which was installed adjacent to HB-1 and SB102, VOC contaminant levels at 8 to 9 feet bgs are generally several orders of magnitude less than those measured at SB102.

The highest SVOC concentrations in soils have generally been measured in surface soil samples collected at SS-3, SS-4, SU-44, and SU-45 outside the site fence near the western corner of the HWSA (Figure 3.1 and Appendix C). Because the shallow SVOC soil contamination extends beyond the HWSA boundaries, this contamination probably originated offsite. Similarly, significant metals concentrations in soil have been measured both on and offsite. The likely sources of these nonsite-related

contaminants are the former coal-fired power plant and coal storage pile west of the site. The HWSA is approximately 180 feet downwind of the coal-fired boiler smokestack and storage pile (Figure 1.2) and the association of SVOCs and metals is common in contamination resulting from the outfall of coal burning.

4.2.2 Groundwater Quality

In June/July 1997, following the installation of the additional groundwater monitoring wells (Section 4.1.2.1), comprehensive groundwater sampling was performed by IT (1997b) at 37 site area wells. Sampling was performed in accordance with sampling and analytical procedures prescribed in the February 1997 amended closure/post-closure plan (Parsons ES, 1997a). Samples were sent to Pace Analytical Services for analysis of VOCs, SVOCs, and metals. In addition, samples were analyzed both in the field and at the laboratory for various geochemical RNA indicators. VOC, SVOC, metals contaminant results are presented in this section, primarily based on the June/July sampling event; previous VOC groundwater contaminant data also are presented for comparison. Groundwater analytical data from the 1997 sampling events are provided in Appendix C, and maximum and 95-percent upper confidence limit (UCL) concentrations are presented in Appendix E. Geochemical data pertinent to the fate of groundwater contaminants is discussed in Section 4.4 and presented in Appendix C.

4.2.2.1 Metals in Groundwater

During the June/July 1997 sampling event, groundwater samples were collected from 18 site monitoring wells for analysis of total and dissolved inorganics. Groundwater samples were collected from the 10 newly installed wells (MW-101S, MW-101D, MW-102S, MW-102D, MW-103S, MW-103D, MW-104D, MW-105S, MW-105D, and MW-106D) and eight existing wells (MW-3, MW-4, MW-5, MW-6, MW-8, MW-9, MW-11, and MW-12). Inorganic concentrations in groundwater from the June/July 1997 sampling event are presented in Table 4.9, and elevated inorganic concentrations also are shown on Figure 4.5.

TABLE 4.9
METALS IN SHALLOW GROUNDWATER
JUNE/JULY 1997
HAZARDOUS WASTE STORAGE AREA
RICKENBACKER ANGB, OHO

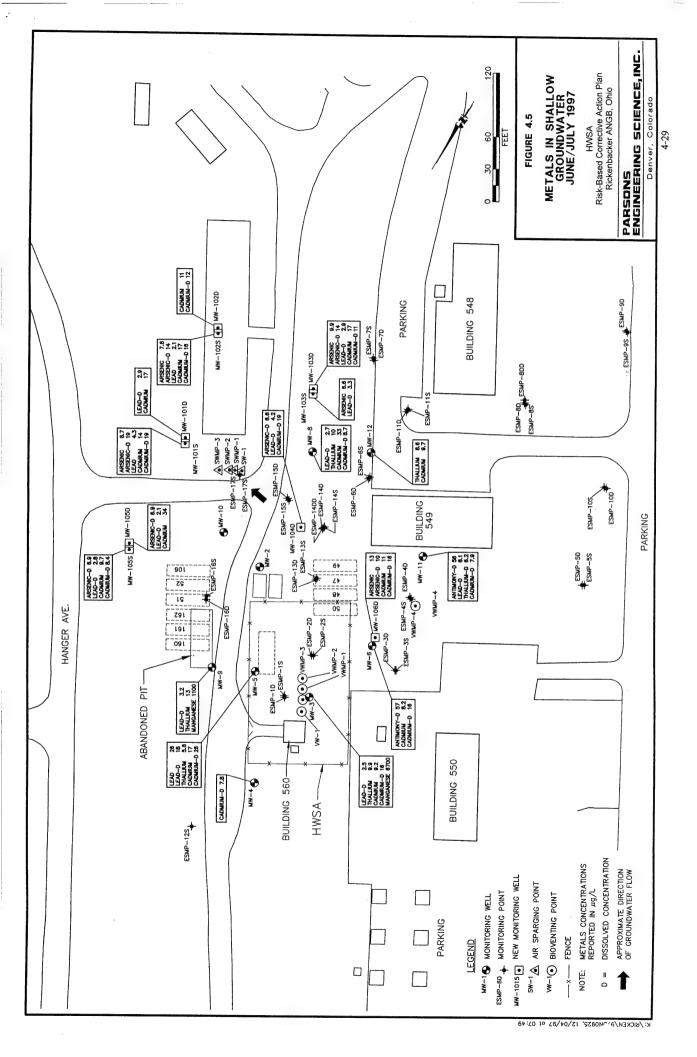
Q- <u>#</u> (4	5.0 U	19	91	12	=	S.0 U	61	S.0 U	*	16	6.7	3.0 U	91	7.8	92	16	.7	5.0 U	e	L)	6.	9	2.0 U	2.0 U	2.9	1,3	1.2	=	œ.	2.0 U	5.1	2.0 U	2.5	2.0 U		œ.	18 2.0 U
Cadmium-D	(#B/L)	S	_						S	90		•			•	•••		90		Lead-D	(µg/L)	1	4	7				4	-	_								
Cadmium	(#B/L)	17	14	17	11	17	5.0 U	5.0 U	¥	7.6	11	5.0 U	7.6	9.2	5.0 U	17	8.2	33	5.0 U	Lead	(4g/L)	2.0 U	2.0 U	2.1	2.0 U	76	07	2.0 U										
Beryllium-D	(#g/L)	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	Iron-D	(μg/L)	2,100	4,300	5,700	3,000	1,900	1,400	1,200	2,500	2,700	1,000	28	¥	25 U	48	4 300	2001	D 22									
Beryllium	(µg/L)	2.0 U	2.0 U	2.0 U	2.0 U	. 2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	Iron	(µ8/L)	2,300	7,400	5,700	2,900	1,900	1,500	1,400	2,600	2,600	1,300	49	4	750	83	2.000		4
Barium-D	(#B/L)	160	981	260	180	200	ま	140	180	360	66	8	28	140	110	230	62	100	43	Copper-D	(µg/L)	6.4	6.3	7.7	3.4	5.3	4.8	3.0	6.2	5.4	6.4	4.3	3.9	4.7	3.4	4.3		3.4
Barium	· (µg/L)	160	081	270	81	200	\$	140	170	360	100	88	82	170	110	220	19	110	4	Copper	(µg/L)	6.7	9.6	5.3	4.4	4.6	5.3	5.3	3.9	4.6	5.8	5.6	3.0 U	3.0 U	3,3	3.0 U		3.0 U
Arsenio-D	(µg/L)	5.0 U	19	14	5.0 U	14	5.0 U	9.9	6.9	6.9	10	5.0 U	Cobalt-D	(µg/L)	10 U		10 U																					
Arsenic	(µg/L)	5.0 U	8.7	7.8	5.0 U	6.6	9.9	5.0 U	5.0 U	5.0 U	13	5.0 U	Cobalt	(µg/L)	10 U		10 U																					
Antimony-D	(µg/L)	40 U	26	40 U	40 U	40 U	40 U	57	40 U	40 U	Chromium-D	(µg/L)	5.0 U	S.0 U	5.0 U		5.0 U																					
Antimony	(µg/L)	40 U	D 04	40 U	40 U	40 U	40 U	40 U	40 U	40 U	40 U	40 U	40 U	40 U	40 U	Chromium	(4g/L)	5.0 U	S.0 U	5.0 U		5.0 U																
Aluminum-D	(#g/L)	52	4	110	47	&	41	\$	9	51	4	78	76	25 U	36	41	76	49	8	Calcium-D	(µg/L)	110,000	97,000	110,000	120,000	110,000	000'06	110,000	110,000	100,000	110,000	110,000	130,000	130,000	120,000	110,000		130,000
١.	(µg/L)	150	2200	150	150	72	81	130	8	8	150	25 U	46	450	25 U	76	25 U	360	33	Calcium	(µg/L)	110,000	100,000	110,000	110,000	120,000	000'06	110,000	110,000	100,000	110,000	120,000	130,000	140,000	130,000	110,000		140,000
Sample	Location	MW-101D	MW-101S	MW-102S	MW-102D	MW-103D	MW-103S	MW-104D	MW-105D	MW-105S	MW-106D	MW-11	MW-12	MW-3	MW-4	MW-5	MW-6	MW-8	6-WM			MW-101D	MW-101S	MW-102S	MW-102D	MW-103D	MW-103S	MW-104D	MW-105D	MW-105S	MW-106D	MW-11	MW-12	MW-3	MW4	MW-5		MW-6

METALS IN SHALLOW GROUNDWATER HAZARDOUS WASTE STORAGE AREA RICKENBACKER ANGB, OHIO TABLE 4.9 (Continued) JUNE/JULY 1997

Torotion	TATE OF THE PERSON	Magnesium-D	Manganese	Manganese-D	Mercury	Meraury-D	Nickel	Nickel-D	Potassium	romstan-D	SCICILIE II	-mamara
TOTAL	(µg/L)	(#g/L)	(µg/L)	(μg/L)	(ug/L)	(µg/L)	(µg/L)	(µg/L)	(μg/L)	(μg/L)	(µg/L)	$(\mu g \Lambda L)$
MW-101D	36,000	36,000	150	150	0.20 U	0.20 U	20 U	20 U	1,700	1,800	6.5	5.0 U
MW-101S	42,000	42,000	230	140	0.20 U	0.20 U	20 U	20 U	1,200	086	S.0 U	5.0 U
MW-102S	33,000	34,000	270	540	0.20 U	0.20 U	D 02	20 U	D 009	1,200	5.0 U	5.0 U
MW-102D	39,000	40,000	140	130	0.20 U	0.20 U	20 U	20 U	1,100	1,800	5.0 U	5.0 U
MW-103D	38,000	37,000	130	130	0.20 U	0.20 U	20 U	20 U	1,500	1,500	6.4	S.0 U
MW-103S	30,000	30,000	170	170	0.20 U	0.20 U	20 U	20 U	1,100	1,300	5.0 U	5.0 U
MW-104D	36,000	36,000	190	190	0.20 U	0.77	20 U	20 U	1,500	2,300	5.2	S.0 U
MW-105D	36,000	36,000	130	130	0.32	0.20 U	20 U	20 U	1,500	2,100	5.0 U	5.0 U
MW-105S	42,000	42,000	230	220	0.20 U	0.20 U	20 U	20 U	1,300	1,500	5.0 U	5.4
MW-106D	36,000	36,000	160	150	0.20 U	0.20 U	20 U	20 U	1,100	1,500	5.0 U	5.0 U
MW-11	36,000	36,000	10	13	0.20 U	0.20 U	20 U	20 U	1,400	1,100	5.0 U	5.0 U
MW-12	43,000	42,000	17	21	0.20 U	0.20 U	20 U	20 U	1,300	1,000	S.0 U	5.0 U
MW-3	29,000	28,000	6,700	200	0.24	0.20 U	20 U	20 U	006	840	5.0 U	5.0 U
MW4	46,000	46,000	8	83	0.20 U	0.20 U	20 U	20 U	620	1,000	5.0 U	5.0 U
MW-5	32,000	32,000	430	400	0.20 U	0.20 U	20 U	20 U	O 009	720	8.6	5.0 U
9-MW	40,000	40,000	4.6	2.5	0.20 U	0.20 U	20 U	20 U	069	099	5.0 U	5.0 U
MW-8	32,000	31,000	130	3.4	0.26	0.20 U	20 U	20 U	U 009	O 009	5.0 U	5.0 U
MW-9	76,000	76,000	1,100	1,000	0.20 U	0.20 U	20 U	20 U	1,100	1,200	5.0 U	5.0 U
	Silver	Silver-D	Sodium	Sodium-D	Thallium	Thallium-D	Vanadium	Vanadium-D	Zinc	Zino-D		
	(µg/L)	(µg/L)	(µg/L)	(µg/L)	(µ8/L)	(µg/L)	(µg/L)	(µg/L)	(µg/L)	(µg/L)		
MW-101D	S.0 U	5.0 U	006'6	008'6	5.0 U	5.0 U	5.0 U	S.0 U	5.4	13		
MW-101S	5.0 U	5.0 U	23,000	21,000	5.0 U	5.0 U	5.0 U	5.0 U	23	13		
. MW-102S	5.0 U	19	10,000	10,000	5.0 U	5.0 U	5.0 U	5.0 U	16	18		
MW-102D	5.0 U	5.0 U	9,300	9,400	5.0 U	5.0 U	5.0 U	5.0 U	15	12		
MW-103D	5.0 U	S.0 U	7,800	7,900	5.0 U	5.0 U	5.0 U	5.0 U	14	13		
MW-103S	5.0 U	5.0 U	2,000	2,000	5.0 U	5.0 U	5.0 U	5.0 U	13	12		
MW-104D	5.0 U	5.0 U	000'6	6,500	5.0 U	5.0 U	5.0 U	5.0 U	. 15	22		
MW-105D	. 5.0 U	5.0 U	000'6	9,200	5.0 U	5.0 U	5.0 U	5.0 U	14	13		
MW-105S	5.0 U	5.0 U	006'9	7,000	5.0 U	5.0 U	5.0 U		15	12		
MW-106D	5.0 U	5.0 U	35,000	36,000	5.0 U		5.0 U		12	13		
MW-11	5.0 U	5.0 U	2,600	2,500	5.0 U		5.0 U		4.0 U	20		
MW-12	5.0 U	5.0 U	12,000	12,000	8.6	5.0 U	5.0 U	5.0 U	4.0 U	13		
MW-3	5.0 U	5.0 U	11,000	11,000	6.6	5.0 U	5.0 U		8	11		
MW4	5.0 U	5.0 U	6,100	6,100	5.0 U	5.0 U	5.0 U	5.0 U	4.0 U	22		
MW-5	5.0 U	5.0 U	13,000	12,000	5.8	5.0 U	5.0 U		4.0 U	19		
9-MW	5.0 U	5.0 U	5,400	2,500	5.0 U	5.0 U	5.0 U	5.0 U	4.0 U	12		
MW-8	5.0 U	5.0 U	4,000	4,000	10	5.0 U	5.0 U	5.0 U		20		
MW-9	5.0 U	5.0 U	7,900	8,100	13	5.0 U	5.0 U	5.0 U	4.0 U	15		

. 2

 $[\]mu g/L=$ micrograms per liter. b D= dissolved concentration of analyte. b U= analyte not detected; number shown represents the laboratory method detection limit.



Based on the June 1997 work plan (Parsons ES, 1997b) and previous groundwater sampling results, antimony, arsenic, beryllium, cadmium, lead, and thallium, are the most significant inorganic contaminants in site groundwater considering preliminary risk screening analysis. During the June/July 1997 sampling event, antimony was detected in the dissolved sample at two wells (57 μ g/L at MW-6 and 56 μ g/L at MW-11), but was not detected in any of the non-filtered samples. Arsenic was detected in five of the non-filtered samples and seven of the dissolved samples. The maximum total arsenic concentration was detected at MW-106D (13 µg/L) and the maximum dissolved arsenic concentration was detected at MW-101S (19 µg/L). Cadmium, both total and dissolved, was detected in 13 of the 18 wells. The two highest total cadmium concentrations were detected at MW-105D (34 μ g/L) and MW-8 (33 μ g/L). maximum dissolved cadmium concentration was detected at MW-5 (26 μ g/L). Total lead was detected in two site monitoring wells (26 μ g/L at MW-5 and 2.1 μ g/L at MW-102S), and dissolved lead was detected in 12 monitoring wells. As with total lead, the maximum dissolved lead concentration was detected at MW-5 (18 µg/L). thallium was detected in five site monitoring wells (the maximum concentration was 13 μ g/L at MW-9), and dissolved thallium was detected only at MW-11 (6.2 μ g/L). Beryllium, total and dissolved, was not detected in any of the 18 samples. Manganese, while not previously identified as a significant groundwater contaminant (Parsons ES, 1997b), was detected above the 95-percent UCL background concentration of 1.34 mg/L (IT, 1997d) during the June/July 1997 sampling event. The maximum total manganese concentration (6.7 mg/L) was detected at MW-3. Dissolved manganese was not detected above background concentrations for any of the 18 samples.

4.2.2.2 SVOCs in Groundwater

Groundwater samples were collected in June/July 1997 from 32 site monitoring wells and analyzed for SVOCs by USEPA Method SW8270B. bis(2-Ethylhexyl)phthalate was detected in samples collected from six monitoring wells. The maximum detected concentration of bis(2-ethylhexyl)phthalate (22 μ g/L) was detected in the sample from MW-105D, but this analyte also was detected in the associated QC

blank. Lower bis(2-ethylhexyl)phthalate concentrations also were detected in samples collected from MW-5, MW-12, ESMP-4S, ESMP-6D, and ESMP-13S. No other SVOCs were detected at more than one location, and most of these were present at near non-detect levels in the sample collected from ESMP-13S. In this sample, acenaphthene (5 μ g/L), carbazole (6 μ g/L), dibenzofuran (2 μ g/L), fluoranthene (1 μ g/L), fluorene (4 μ g/L), and phenanthrene (2 μ g/L) were detected. Other SVOC detections in groundwater include naphthalene (28 μ g/L) and 2-methylnaphthalene (70 μ g/L) at MW-5, and di-n-octylphthalate (4 μ g/L) at MW-105D.

4.2.2.3 VOCs in Groundwater

During the June/July 1997 sampling event, samples were collected from 37 site monitoring wells and analyzed for VOCs by USEPA Method SW8260. Aromatic hydrocarbons (i.e., BTEX) and CAHs historically have been detected in groundwater samples collected from the site. The CAHs that historically have been detected are *cis*-1,2-DCE, *trans*-1,2-DCE, 1,1-DCE, 1,2-dichloroethane (DCA), TCE, and VC. The analytical results for these VOCs are presented in Table 4.10, and include groundwater sampling events conducted at the site between September 1988 and June/July 1997.

The approximate horizontal extent of BTEX in groundwater has been estimated based on various sampling events at the HWSA since February 1995. Figure 4.6 shows the estimated extent of total BTEX concentrations exceeding 5 μ g/L. BTEX generally were detected in groundwater samples collected from the shallow sand and gravel interval (i.e., ESMP-13S, ESMP-16S, and MW-5). The highest total BTEX concentrations historically occur at ESMP-13S, which is located adjacent to one of the former UST cavities at the site. During the June/July 1997 sampling event, the total BTEX concentration at this location was 872 μ g/L, and the benzene concentration was 670 μ g/L. The vertical extent of contamination at this location was defined by the sample collected from ESMP-13D where total BTEX was not measured above detection limits. The ESMP-13S cluster is located downgradient from and adjacent to the four former 25,000-gallon USTs (Figure 4.2). The second highest concentration of BTEX

TABLE 4.10
BTEX AND CHLORINATED VOCS DETECTED IN SHALLOW GROUNDWATER
SEPTEMBER 1988 - JUNE/JULY 1997
HAZARDOUS WASTE STORAGE AREA
RICKENBACKER ANGB, OHIO

Sema-le	1	P	T.	Total Language	Total	Total	Vinyl	1	F F		1 2 P.C E 44-1-11	150.0	i c
Location	Date	(µg/L)	(ug/L)	(µg/L)	λy reale (μg/L)	(LR/L)	(µg/L)	(ug/L)	(ug/L)	(ug/L)		(ug/L)	(ug/L)
ESMP-1S	Feb-95		BLO		-	BLQ	,						
ESMP-1D	Feb-95	:	BLQ	BIQ	:	BLQ	:	:	:	:	1	:	:
ESMP-2S	Feb-95	1	3.6	1	,	3.6	,	:		:		:	:
	Jun/Jul-97	ŧ	:	:	:	:	1	:	:	:		:	1
ESMP-2D	Feb-95	:	1.05	:	1	1.05	1	:	1	BCL	1	:	92.6
	Aug-95	ı	:	ı	;	:	i	:	:	1		;	20
	Dec-95	:	1	1	ı	:	i	:	ı	ı	:	:	14
	Mar-96	1	ı	:	;	ı	1	;	ı	:	ı		٥
	Jun-96	ŀ	:	:	:	ŀ	:	:	1	;	:	;	6.5
	Sep-96	:	:	:	1	:	ı	ŧ	;	ł	:	:	7.7
	Dec-96	i	1	ı	1.51	1.53	:	;	:	:	:	:	31
	Mar-97	:	ŧ	;	ı	ı	:	ŀ		:	:	:	9
	Jun/Jul-97	:		:	:	:	1	:	:	;		:	4.4
ESMP-3S	Feb-95	:	:	:	:	1	ı	:	ì	BCL	BCL	:	-
	Jun/Jul-97	:	1				1	:	:	:		:	:
ESMP-3D	Feb-95	1	BLQ	:	:	BLQ	:	:	1	1.3	1.3	:	6.5
	Aug-95	:	1		:	:	1	:	:	;		ł	:
	Dec-95	;	t	:	1	1	:	:	1	:	:	:	:
	Mar-96	:	:	;	;	1	;	1	:	:	ï	:	:
	Jun-96	:	ı	:	:	ı	:	:	:	;	:	:	:
	Sep-96	:	ı	;	:	;	ı	:	:	:	:	:	:
	Dec-96	:	t	:	4.51	4.51	;	:	1	:	:	ì	ı
	Jun/Jul-97	:	:	:	:	:	:	-	:	;	:	:	ł
ESMP-4S	Feb-95	:	BLQ	:	:	BLQ	1	:	:	:	•	·	:
	Var.95	:	:	i	:	1	:	;	:	1	t	:	:
	Dec-95	ł	:	:	;	ŧ	:	:	:	:	:	1	ł
	Mar-96	:	ı	:	:	:	:	:	:	:	i	;	:
	Jun-96		:	:	1	ŧ	:	:	•	:	:	;	:
	Sep-96	ı	:	ı	:	;	:	:	:	:	ŧ	:	:
	Dec-96	ı	:	:	1	ł	22	1	1	:	:	:	1
	Jun/Jul-97	1	:	:		:	:	:	:	i	:	:	:
ESMP-4D	Feb-95	:	2.48	:	:	2.48	91	1	:	ı	:	:	:
	Aug-95	:	ŧ	:	:	1	15	1	ı	ı	1	:	:
	Dec-95	1	ı	:	ŧ	:	15	:	1	:	:	;	:
	Mar-96	:	i	:	:	:	19	:		1	:	•	:
	Jun-96	:	:		:	1	17	ı	1	1	;	:	1
	Sep-96	:	:	:	1	;	17	1	:	:	:	:	:
	Dec-96	:	:	1	;	:	22	:	1	:	:	:	:
	Jun/Jul-97	1	:	:	:	1	55E	t	:	:	:	ŧ	;
ESMP-5S	Feb-95	1	:	;	-	:		:	:	:	ı	1	:
ESMP-5D	Feb-95	:	:	:	:	:	:	;	:	;	:	:	1
													п

TABLE 4.10 (Continued) BTEX AND CHLORINATED VOCS DETECTED IN SHALLOW GROUNDWATER SEPTEMBER 1988 - JUNE/JULY 1997 HAZARDOUS WASTE STORAGE AREA RICKENBACKER ANGB, OHO

					Total	Total	Vinyl			ı		1040	202
Sample	Sample	Benzene	Toluene	Ethylbenzene	Xylence	BTEX	Chloride	1,1-DCE	T-1,2-DCE	C-1,2-DCE	1,2-DCE (total)	1,2-DCA (ue/L)	ICE (##/L)
Location	Date	(mg/L)	(#8/L)	(HB/L)	(48/17)	3 13	NA U	NA NA	NAN		NA	VX.	Ϋ́
ESMP-6S	reb-93	ı	5.13	t	:	3.13	V.	4	4		1		:
	/A-Jnf/unf	٠		:	•								
ESMP-6D	Feb-95	:	ı	:	:	:	1.7	:	:	1	ı	:	:
	Aug-95	:	:	ı	:	:	:		:	1	ı	:	:
	Dec-95	ı	:	1	:	:	7	:	1	:	:	:	:
	Mar-96	:	:	ı	1	1	1.9 J	:	:	1	:	:	ı
	Jun-96	1	1	:	ŀ	:	1.71	:	:	1	:	:	:
	Sen-96	:	:	:	:	:	;	!	1	1	1	:	:
	Dec-96	1	:	;	:	:	;	1	:	ı	:	:	:
	Tum/ful.07	1	1	:	:	;	2.8	:	;	:	ì	:	:
26 MD 76	Eath OK					:				:	1	:	ŧ
CENT TO	Ech 05		PIO	:		BLO	;	:	:		:	;	:
ESME-1D	Ech 06		=	:		1.3			:		:	:	BCL
MF-63	Aug. 05	: 1	} :	: :	· •	1	1	:	:	ŧ	1	:	:
	70-04 70-04		:	ı	:	;		t	1	ţ	. 1	ı	:
	Mar-96	: :	1	:	:	:	1	;	1	ı	ı	ŀ	•
	Jun-96	:	:	:	:	1	:	ı	:	ŧ	:	:	:
	Sep-96	:	ı	:	:	:	i	:	;	1	:	:	:
	Dec-96	1	:	:	:	:	:	:	:	:	:	:	:
	Jun/Jul-97	:	:	:	;	:	:	:	:	:	:	:	:
ESMP-8DD	Feb-95		BLO	:	BLQ	BLQ	:	1	1	:	:	:	BCL
ESMP-9S	Feb-95	:	BLO	:	:	BLQ	:	:	:	:	:	:	BCL
ESMP-9D	Feb-95	1	;	1	:	:	:	•	:	;	:	;	:
ESMP-10S	Feb-95		BLQ	ı	1	BLQ	1	1	1	:	1	:	
	Aug-95	:	ŧ	•	:	:	:	:	ı	:	ı	ı	:
	Dec-95	:	:	:	1	:	:	:	ŀ	:	:	:	t
	Mar-96	1	:	:	:	:	:	:	:	1		:	:
	Jun-96	:	:	ı	:	:	:	:	:	:	:	1	•
	Sep-96	:	ı	:	1	1	1	:	1	;		:	:
	Dec-96	:	1	:	6.4	6.4	1	:	I.	1	:	:	ı
	Jun/Jul-97	1	:	:	1	1	1				-	•	ا:
ESMP-10D	Feb-95	:	:	:	1	;	:	:	:	:		•	:
ESMP-11D	Feb-95	BLQ	BLQ	BLQ	BLQ	BĽQ	:	:	:	:	:	1	
ESMP-12S	Mar-95	1	2.44	:	:	2.4	1	:	;	:	:	220	BCL
	Jun/Jul-97	:	:	:	;	:	:			:	:	1200D*	: 1
ESMP-13S	Feb-95	424.18	22.41	237.09	279.58	963.26	2.7	1.3	300	228	228	:	42.0
	Aug-95	500	:	130	100	430	:	:	430	:	1 3	·	:
	Dec-95	270	8.9 J	140	8	508.9	:	:	230	180	410	;	1
	Mar-96	280	8.2 J	120	34	442.2	62	:	230	210	440	:	:
	Jun-96	300	6.7	120	20	476.7	120	:	190	200	390	:	:
	Sep-96	390	10	160	100	099	72	:	51	8.6	151	:	:
	Dec-96	510	251	190	77	777	. 41	:	:	:	1	:	1
	Mar-97		1	180	84	864	45	:	:	;	:	: 3	:
	Jun/Jul-97	7 670D	12	190E	:	872	:	:	:		:	72	:

TABLE 4.10 (Continued)

BTEX AND CHLORINATED VOCS DETECTED IN SHALLOW GROUNDWATER
SEPTEMBER 1988 - JUNE/JULY 1997
HAZARDOUS WASTE STORAGE AREA
RICKENBACKER ANGB, OHIO

Samely	County	9	7	Dittille	Total	Total	Vinyl	1 1 PCE	T.12.PCE	C.1.2.TOF	1 2. DCE (total)	1 2.DC4	TCE
Location	Onte	(µg/L)	(µg/L)	(µg/L)	(Fg/L)	(µg/L)	(µg/L)	(rg/L)	(#g/L)	(µg/L)	(# g /L)	(µg/L)	(µg/L)
ESMP-13D	Feb-95	,		BLQ	1	BLQ	:	ł	:	;	t		:
ESMP-13D(D)	Feb-95	NA	NA	NA	NA	NA	;	:	:	•			
ESMP-14S	Feb-95	:	1	:	:	ı	:	:		1		:	1
ESMP-14S(D)	Feb-95	Y.V	٧٧	٧×	٧٧	NA.	:	1	:	***		:	:
ESMP-14D	Feb-95	:	:	1	:	:	-		1	2.2	2.2	:	:
	Aug-95	1	1	ı	:	1	1	1	:	ı	1	:	:
	Dec-95	ı	:	:	:	1	1	1	:	4.11	4.13	:	;
	Mar-96	:	:	1	;	:	1	;	ı	5.8	5.8	:	:
	Jun-96	ı	:	:	:	:		:	ı	s	s	:	;
	Sep-96	:	:	ı	:	1	2.1	1	8.5	:	8.5	:	;
	Dec-96	1	:	ı	ı	1	2.8	:	:	11	==	t	;
	Mar-97	ı	i	:	;	:	4.8	:	:	15	15	:	;
	Jun/Jul-97	1.7	:	:	:	1.7	9.1	:	:	22	22	:	:
ESMP-14DD	Feb-95	:	BLQ	:	:	BLQ	:	:	;	;	:	:	BCL
ESMP-15S	Feb-95	:	1.08	1	ı	1.08	:	:	:	:	:	:	:
	Jun/Jul-97	:	;	:	:	:	:	;	:	1	:	1.5	:
ESMP-15D	Feb-95	:	:	:	:	:	:	:	:	;	:	:	1
ESMP-16S	Mar-95	89.18	BLQ	228.5	153.65	471.33	1	:	:	BCL	BCL	1	;
	Aug-95	23	:	23	7	23	:	:	:	:	ı	:	:
	Dec-95	S	:	S	S	15	:	ŧ	:		:	:	:
	Mar-96	18	:	1	:	18	1	1	:	:	:	:	1
	Jun-96	3.13	:	:	:	3.13	:	:	:	:	:	;	:
	Sep-96	=	:	1	ı	11	:	:	:	1	:	:	:
	Dec-96	9.1	:	:	:	9.1	:	:	:	:	ı	ı	ı
	Jun/Jul-97	1.7	:	:	:	1.7	:	:	:	:	:	:	:
ESMP-16D	Mar-95	6.46	BLQ	26.87	62.03	95.36	;	1	:	BCL	BCL	:	:
	Aug-95	:	:	:	:	:	:	:	:	ł	:	;	;
	Dec-95	1	:	:	:	:	:	:	;	1	:	:	1
	Mar-96	:	:	ı	:	:	:	:	i	1	:	:	:
	Jun-96	:	:	:	1	1	1	1	ı	1	:	:	:
	Sep-96	:	:	:	:	: ;	1	ı	:	:	:	ı	1
	Dec-96	:	:	1	3.8	3.87	:	1	:	:	:	:	:
	Jun/Jul-97	:				:		:		-	:	:	:
ESMP-17S	Feb-95	:	:	1	:	:	1570	11.7	152	4913	2065	:	BCL
	Aug-95	•	1	:	:	:	930E	:	7730E	1	7730E	:	:
	Dec-95	:	:	:	;	:	1200	6.6	130	4200	4330	:	:
	Mar-96	t	:	:	:	:	670	:	130	2800	2930	:	:
	Jun-96	:	:	:	:	:	200	5.7	83	3000	3093	:	:
	Sep-96	:	:	;	1	:	460E	S	1800E	78	1878E	:	;
	Dec-96	1	:	:	:	:	210D	2.31	48	1500E	1548E	1	;
	Mar-97	1	:	:	:	:	49E	2.2	29E	700E	729E	1	:
	Jun/Jul-97	:	:	:		:	9009	3.9	•	1100E	1100E	:	;
MW-1	Sep-88	ま	:	:	:	1	;	:	1	:	:	1	1
	Feb-90	260	:	110	121	;		•	:	:	:	:	1

TABLE 4.10 (Continued) BTEX AND CHLORINATED VOCS DETECTED IN SHALLOW GROUNDWATER SEPTEMBER 1988 - JUNE/JULY 1997 HAZARDOUS WASTE STORAGE AREA RICKENBACKER ANGB, OHO

Semale	Sample	Renzene	Tohrene	Fehrelhenzene	Yvlence	RTEX	Chloride		T-1 2-DCE	C-1 2-DCE		1.2-DCA	TCE
Location	Date	(µg/L)	(µg/L)	(µg/L)	(µg/L)	(µg/L)	(µg/L)	(HR/L)	(µg/L)	(µg/L)	(µg/L)	(µ8/L)	(µg/L)
MW-2	Sep-88	:				,			:	;		:	;
	E.h. 00	1	1	:		:	:		:	1		:	1
	200	1	1	} ;	: :	: :	: :		1	1		:	1
	76-150	: ;	: ;	: ;	: ;		}	ł	}				
	Mar-95	A M	A T	PTY T	PITS	YIN B	:	:	1	ı	:	:	:
	Sep-96	:	:	:	:	:	:	ŧ	:	ı	:	:	ı
	Dec-96	:	1	ı	3.21	:	:	1.	•	•	:	:	:
	Mar-97	ı	t	:	:	:	:	t	:	ı	:	:	ł
	Trin/Int.07	1	:	:	:	:	:	:	1	8.4	8.4	;	ł
1007	00 0												44
MW-5	2ch-02	ŧ	:	:	!	:	:			ł	}	ł	ŧ
	Feb-90	:	1	ı	ı	:	:	ı	:	:	:	:	- :
	Oct-91	ı	:	ı	:	:	:	1	:	:	:	:	01
	Feb-95	BIQ.	BLQ	BLQ	BLQ	BLQ	:	:	:	BIC	BIC	:	-
	Aug-95	i	:	:	:	:	:	:	:	:	;	:	:
,	Dec-95	ı	:	:	:	ı	:	:	:	:	:	:	:
	Mar-96	ì	i	:	ŧ	:	:	:	:	1	:	;	ŧ
	Jun-96	ı	1	:	ı	:	:	:	:	1	:	1	:
	Sep-96	ı	:	:	:	:	;	;	:	:		;	SI
	Dec-96	:	ı	:	:	:	:	:	:	:	:	:	1
	Mar-97	:	:	:	:	:	:	:	i	:	;	:	0.53J
	Jun/Jul-97	:	:	:	:	:	;	;	;	:	:	1	1.6
MW-4	Feb-90	;	:	:		;	:	!	:		:	:	:
	Oct-91	ı	:	:	:	:	;	1	:	:	:	:	:
	Mar-95	BLQ	BLQ	BLQ	BLQ	BLQ	:	:	:	:	:	:	:
	Aug-95	:	:	:	:	;	:	:	:	;	:	:	:
	Dec-95	ı	:	:	;	:	:	:	;	ï	:	:	ı
	Mar-96	:	:	:	:	;	:	:	:	:	:	1	:
	Jun-96	;	:	:	:	:	1	:	:	;	i	:	1
	Sep-96	ı	:	:	:	;	1	:	:	1	:	:	1
	Dec-96	:	:	i	6.2	6.2	;	:	:	i	;	:	;
	Mar-97	:	:	1	:	:	;	:	:	1	:	ı	1
	Jun/Jul-97	;	:	:	:	;	:	:	:	;	:	1	
MW-5	Mar-95	17.19	41.94	317.97	375.93	753.03	;	:		1	1	:	:
	Aug-95	:	15	170	210	395	:	1	:	:	1	:	ı
	Dec-95	1	12.1	140	190	342	:	:	1	:	ı	;	t
	Mar-96	ŧ	4.8 J	¥	100	158.8	:	:	1	:	:	:	ı
	Jun-96	:	131	190	210	413	:	1	:	1	1	:	:
	Sep-96	;	7.6	150	180	337.6	;	;	:	:	;	:	:
	Dec-96	:	:	230	290	520	:	:	:	1	:	:	:
	Mar-97	7.91	4	210	218	449.9	:	:	:	:	1	;	1
	Jun/Jul-97	:	9.4	260D	280D	549.4	:	1	1	:	1	;	ŀ

TABLE 4.10 (Continued)
BTEX AND CHLORINATED VOCS DETECTED IN SHALLOW GROUNDWATER
SEPTEMBER 1988 - JUNE/JULY 1997
HAZARDOUS WASTE STORAGE AREA
RICKENBACKER ANGB, OHIO

					Total	Total	Vinyl						
Sample	Sample	Benzene	Toluene	Ethylbenzene	Xylenes	BTEX	Chloride	1,1-DCE	T-1,2-DCE	C-1,2-DCE	1,2-DCE (total)	1,2-DCA	TCE
Location	Date	(μg/L)	(πg/L)	(μg/L)	(µg/L)	(µg/L)	(μg/L)	(µg/L)	(# g /L)	(μg/L)	(μg/L)	(µg/L)	(µg/L)
MW-6	Feb-90	ı	:	:	1	ı	1	:	••	ı	~	1	78
	Oct-91	1	:	ı	1	1	:	1	1	1	190	1	2000
	Feb-95	¥	Ϋ́	ΥV	Y _N	AN	21.9	6.5	43.2	ž	837.2	ı	9580
	Aug-95	ı	ı	:	1	1	:	:	ı	:	1	ŧ	770
	Dec-95	:	:	•	1	1	ı	1	1	27	27	1	180
	Mar-96	ı	:	1	ŀ	:		;	1	:	1	1	11
	Jun-96	i	:	1	1	ŀ	:	:	1.73	26	287	:	280
	Sep-96	ı	:	:	1	:	1	;	18		18	1	220
	Dec-96	:	ı	:	:	:	1	:	1	10	10	ı	130
	Jun/Jul-97	1	1	:	:	:	:	:	:	1.8	1.8	:	24
MW-7	Feb-90	7 7 7 8	1	જ	70	ŀ	:	:	:	ı	:	:	ı
	Oct-91	19	77	62	36	1	17	:	:	ł	:	•	1
MW-8	Feb-90	:	:	:	1	:	:	:	:	:	:	:	1
	Oct-91	:	:	:	:	ı	:	:	:	:	·	:	;
	Feb-95	BĽQ	ı	:	;	BĽQ	1	:	:	ı	:	:	BCL
	Aug-95	1	:	:	:	1	:	:	:	:	:	:	;
	Dec-95	1	:	1	:	1	1	ı	1	ı	ı	:	ı
	Mar-96	1	:	:	:	:	:	:	:	:	:	1	:
	Jun-96	1	:	:	ı	:	:	:	:	:	:	:	:
	Sep-96	:	:	:	:	:	1	:	1	:	:	:	:
	Dec-96	ı	:	:	4.81	4.83	1	:	ı	:	:	:	;
	Mar-97	1	:	:	ı	:	:	:	:	1	:	:	:
	Jun/Jul-97	:	:	:	;	:	:	:	:	:	:	:	
MW-9	Feb-90	1	:	:	:	:	;	:	:	:	ŧ	:	:
	Oct-91	:	:	:	:	:	:	:	:	:	:	:	;
	Mar-95	:	:	:	1	:	1	:	;	:	:	:	:
	Jun/Jul-97		:	:		:	-	:	1	:	:	1	
MW-10	Oct-91	ı	:	:	ı	:	:	1	:	:	:	,	:
	Mar-95	1	1	:	ı	:	:	:	:	ı	:	;	1
	Jun/Jul-97	:	:	:	:	:	:		:	:	:	:	:
MW-11	Oct-91	1	:	:	1	:	:	:	:	:	:	ı	:
	Feb-95	1	:	1	1	1	:	:	:	1	1	:	1
	Aug-95	ı	1	:	:	:	:	1	:	ı	:	1	:
	Dec-95	:	:	:	:	:	:	:	:	1	:	:	ı
	Mar-96	:	ŧ	:	1	:	ı	:	ı	:	ı	:	:
	Jun-96	:	:	:	:	:	:	:	:	:	:	:	:
	Sep-96	:	:	1	:	ı	:	:	:	:	1	ı	:
	Dec-96	:	:	:	:	ı	:	:	:	:	:	ı	1
	Jun/Jul-97	:	:	•	:	;		:	:	:	ı	-	1

TABLE 4.10 (Continued) BTEX AND CHLORINATED VOCs DETECTED IN SHALLOW GROUNDWATER SEPTEMBER 1988 - JUNE/JULY 1997 HAZARDOUS WASTE STORAGE AREA RICKENBACKER ANGB, OHIO

					Total	Total	Viny						
Sample	Sample	Benzene	Benzene Toluene	Ethylbenzene	Xylenes	BTEX	Chloride	1,1-DCE	T-1,2-DCE	C-1,2-DCE	1,1-DCE T-1,2-DCE C-1,2-DCE 1,2-DCE (total)	1,2-DCA	TCE
Location	Date	(µg/L)	(#g/L)	(µg/L)	(µg/L)	(µg/L)	(µg/L)	(µg/L)	(µg/L)	(μg/L)	(µg/L)	(µg/L)	(µg/L)
MW-12	Oct-91	:											
	Feb-95	:	BLQ	1	ŧ	BLQ	1	:	1	t	1	:	;
	Aug-95	:	:	:	:	ı	1	:	:	1	:	ı	:
	Dec-95	:	ı	1	1	1	1	:	1	1	:	1	:
	Mar-96	:	:	1	ı	1	1	1	1	1	:	1	:
	Jun-96	:	ŀ	ı	1	1	1	:	1	1	:	:	:
	Sep-96	:	:	t	:	1	:		:	:	1	1	:
	Dec-96	:	:	:	;	:	ı	:	:	:	:	:	1
	Jun/Jul-97	1	1	t	:	ı	:	ı	:	1	1	1	:
MW101S	76-lnf/unf	:	:	:	:	1	14	:	4.1	29E	33.1E		:
MW101D	Jun/Jul-97	:	;	:		1			:		:	1	:
MW102S	Jun/Jul-97	:	:	:	:	:	:			1	:	:	1
MW102D	Jun/Jul-97	;	:	:	:		:	:	1	•	1	:	:
MW103S	Jun/Jul-97	:	1	:	:	•	1	:		•	:	1	:
MW103D	Jun/Jul-97	1	:	:	:	1		:		ł	:	:	:
MW104D	Jun/Jul-97	:	:	:	:	:	:			1	1	:	:
MW105S	76-Inf/unf	1.7	1	:	:	1.7	:	1		:	ī	:	:
MW105D	Jun/Jul-97	:	:	:	•	:	:	•		:	:	:	:
MW106D	76-lnf/unf	2.4	:	;	:	2.4	:	14	•	1200E	1200E		1500D

n/ -- not detected.

b/ BLQ - below limit of quantitation.

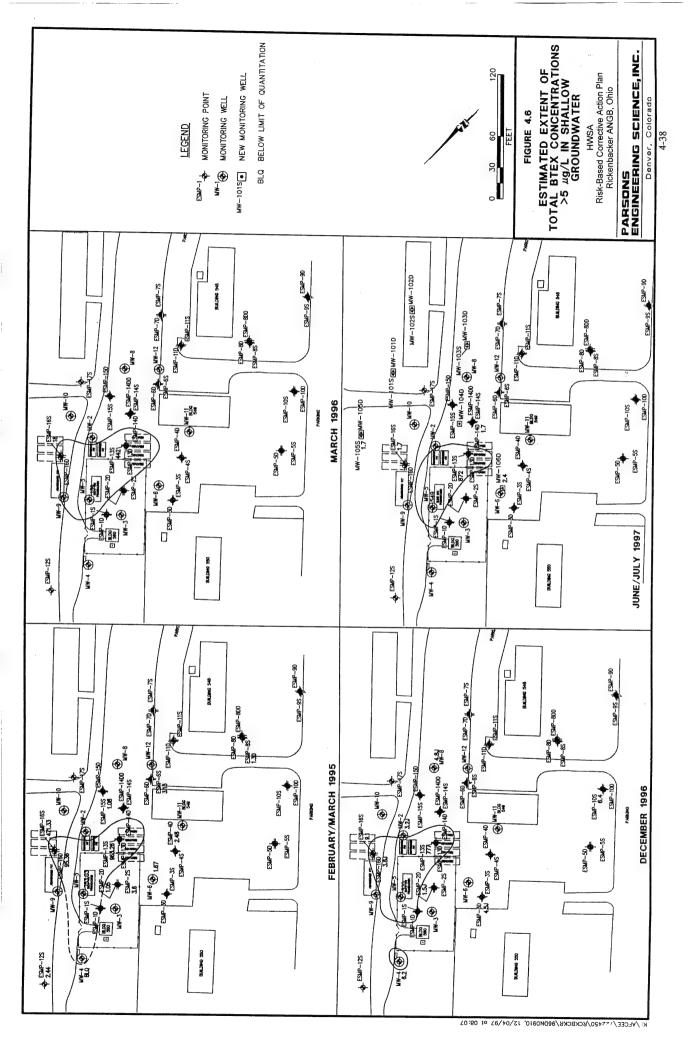
c/ BCL = below calibration limit (1.0 mg/L).

d/ 1 = compound detected, but below the laboratory reporting limit; value shown represents a laboratory estimated concentration.

e/E = concentration exceeded upper calibration standard; while shown represents an estimated concentration.

f/ NA - not analyzed.

g/ D = compound analyzed at a secondary dilution factor.



was measured at MW-5 (549 μ g/L), where mobile LNAPL has been observed in the past and where the mobile LNAPL thickness in February 1995 was measured as 0.35 foot. Analysis of a sample of the free product collected from MW-5 during the February 1995 sampling event indicated a total BTEX concentration of 627 μ g/L with a nondetectable quantity of benzene, indicating a severely weathered product (Parsons ES, 1997a). Mobile LNAPL thickness at MW-5 ranged from a sheen to 0.1 foot during 1996 quarterly monitoring (IT, 1997a) and a sheen was observed during sampling performed in June/July 1997.

Figure 4.6 illustrates that the dissolved BTEX plume does not appear to be expanding in areal extent over this 1995 to 1997 monitoring period. Dissolved BTEX concentrations at ESMP-16S have consistently decreased during this period, and it appears that BTEX concentrations are generally decreasing, or remaining stable, throughout the plume without significant horizontal or vertical migration. Data collected at the site since February 1995 suggest that natural chemical attenuation processes, specifically biodegradation, are reducing BTEX concentrations at this site. RNA of these petroleum constituents in groundwater is discussed further in Section 4.4.

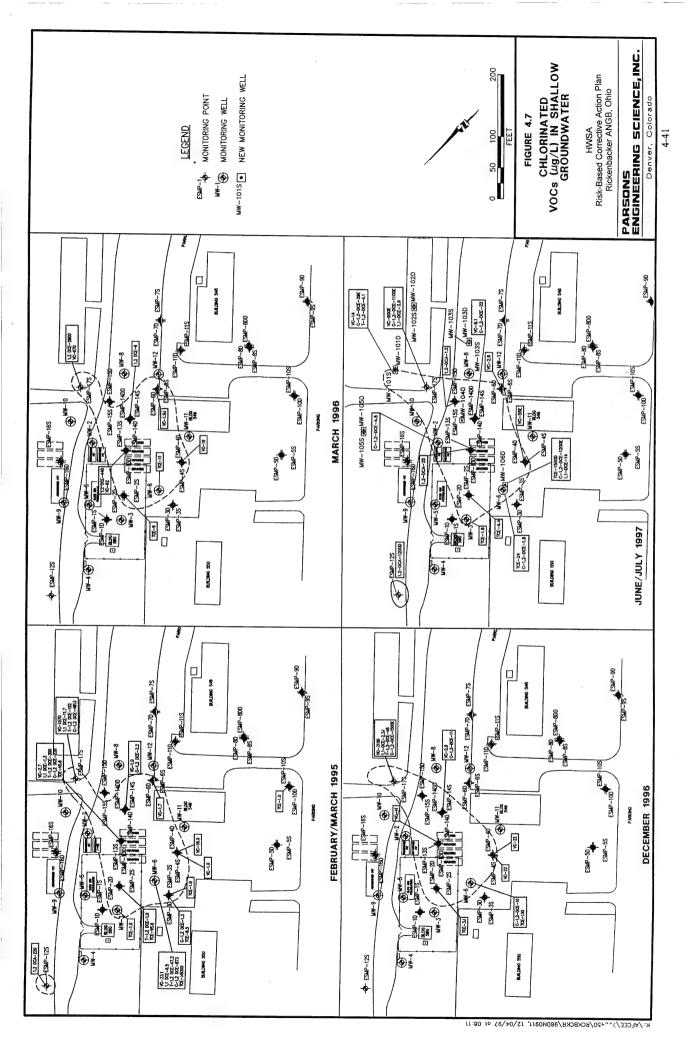
In addition to BTEX, CAHs have been historically detected in groundwater samples collected at and near the HWSA. Prior to the installation of several groundwater monitoring points in February/March 1995 (Section 3.1.1), elevated chlorinated VOC concentrations were routinely detected in samples collected from MW-3 and MW-6 (Table 4.10). Groundwater sampling performed subsequent to the groundwater monitoring point installation has shown elevated CAH concentrations in groundwater collected from ESMP-4D, ESMP-13S, and ESMP-17S. All five of these monitoring wells/points are screened in the sand and gravel zone approximately 14 to 18 feet bgs (Figures 2.5 and 2.6). CAHs have been detected in ESMP-14D, which is screened in the deeper sand seam at approximately 21 to 25 feet bgs; however, no VOCs were detected from ESMP-14DD (which is screened in the basal clay at approximately 26 to 29 ft bgs) during the February/March 1995 sampling event. The areal extent of CAHs

in groundwater, considering sampling results from the last 2 years of site monitoring, is shown on Figure 4.7.

Historically, the highest chlorinated VOC concentrations consistently have been measured in groundwater samples collected from ESMP-17S; however, elevated CAH concentrations also were detected in June/July 1997 samples collected from newly installed MW-106D and ESMP-12S. At MW-106D, TCE was detected at 1,500 μ g/L, cis-1,2-DCE was detected at 1,200 μ g/L, and 1,1-DCE was detected at 14 μ g/L. At ESMP-12S, 1,2-DCA was detected at 1,200 μ g/L. During this same sampling event, VC (600 μ g/L), 1,1-DCE (3.9 μ g/L), and cis-1,2-DCE (1,100 μ g/L) were detected at ESMP-17S. Chlorinated VOCs also were detected in the sample collected from newly installed MW-101S. At this well, VC (14 μ g/L), cis-1,2-DCE (29 μ g/L), and trans-1,2-DCE (4.1 μ g/L) were detected.

The CAH results from the June/July 1997 sampling event further define the extent of groundwater contamination in the vicinity of the HWSA. MW-106D is screened from approximately 17.5 to 22.5 ft bgs and was installed near MW-6 to evaluate the VOC contamination in the lower sand seam, above the basal clay layer. Other lower sand seam monitoring points in which chlorinated VOCs have been detected are ESMP-6D (VC detected at $2.8~\mu g/L$) and ESMP-14D (cis-1,2-DCE detected at $2.2~\mu g/L$). Lesser chlorinated VOC concentrations at these two groundwater MPs, combined with repeated non-detect measurements at ESMP-3D, suggest that the elevated concentrations detected at MW-106D are localized, and minimal, if any, migration of contaminants is occurring within this zone.

Minimal groundwater sampling has been performed at ESMP-12S since its installation in early 1995; however, 1,2-DCA was detected during both the March 1995 sampling event (at 220 μ g/L) and the June/July 1997 comprehensive sampling event (1,200 μ g/L). ESMP-12S is screened from approximately 12 to 16 feet bgs in the same upper sand seam as ESMP-4D, ESMP-13S, and ESMP-17S, but groundwater elevation data for this point (Tables 2.2 and 4.4) consistently indicate that water is



present at very shallow depths (as little as 0.40 ft bgs in March 1996). Based on the established easterly groundwater flow in the upper sand seam (Figures 2.7, 2.8, and 4.2) and typical depths to water (8 to 10 feet bgs), it does not appear that the VOC contamination evident in samples collected from ESMP-12S originated from within the HWSA.

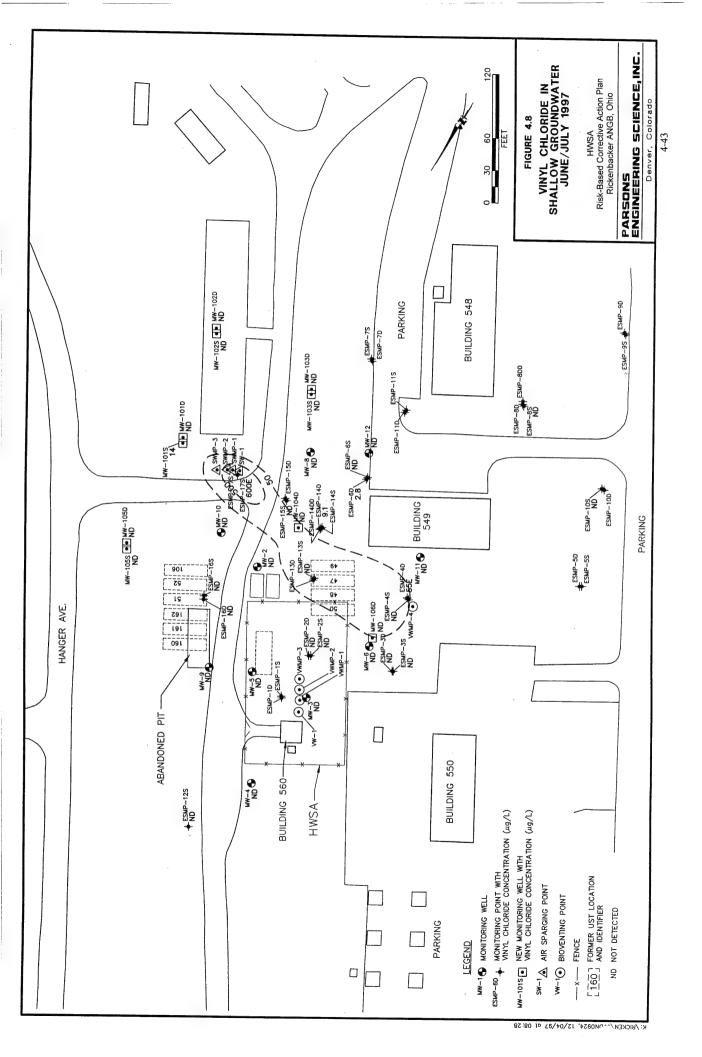
Chlorinated VOCs detected in newly installed downgradient well MW-101S indicate that the downgradient extent of the CAH plume within the upper sand seam has not been completely delineated. While VC and 1,2-DCE (cis- and trans-) were detected at MW-101S, contaminant concentrations are 1 to 2 orders of magnitude less than concentrations detected in the June/July 1997 sample from ESMP-17S. Based on the relatively short distance between ESMP-17S and MW-101S (approximately 35 feet), and the significant difference in contaminant concentrations, it is anticipated that the leading edge of the chlorinated VOC plume does not extend far past MW-101S.

The most prevalent chlorinated VOC detected at the HWSA is VC. Vinyl chloride is a degradation product of DCE, which in turn is a degradation product of TCE. During the June/July 1997 sampling event, VC was detected at ESMP-4D (55 μg/L), ESMP-6D (2.8 μg/L), ESMP-14D (9.1 μg/L), ESMP-17S (600 μg/L), and MW-101S (14 μg/L). The estimated areal extent of VC in groundwater during the June/July 1997 sampling event is shown on Figure 4.8. The prevalence of VC at this site indicates that CAH degradation is occurring through reductive dehalogenation. The historical contaminant data (Table 4.10) suggest that VC is not accumulating in site groundwater, and appears to be slowly degrading. Conversely, total 1,2-DCE and TCE concentrations generally appear to be decreasing in site monitoring wells. Further analysis of chemical fate in groundwater is provided in Section 4.4.

4.3 TREATABILITY TESTING RESULTS

As part of the 1997 assessment activities, air injection bioventing and air sparging treatability tests were performed to evaluate the effectiveness of these source reduction technologies at remediating contaminated soils near Building 560 and groundwater

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downgradient from the site near ESMP-17S (Figure 4.1). In the event that risk reduction through source area treatment of soils and/or groundwater proved necessary considering future land use and findings of the risk assessment (Section 5), treatability testing results would be used to design the remedial approach.

Bioventing and air sparging treatability tests at the HWSA were initiated by IT (1997c) in August 1997 following installation of the bioventing and air sparging wells/monitoring points in June 1997. Treatability testing procedures and methods as outlined in the June 1997 additional assessment activities work plan (Parsons ES, 1997b) were generally followed except where site conditions prevented adherence to these procedures. This section reviews the preliminary results of the treatability testing effort. A more detailed summary of testing results and system effectiveness will be presented by IT (1997c) in the treatability testing results report for the HWSA.

4.3.1 Bioventing Treatability Test Objective and Results

The bioventing pilot-scale system was installed near Building 560 because of the elevated concentrations of petroleum hydrocarbons (primarily ethylbenzene and xylenes) present in vadose zone soils in the vicinity of borings HB-1 (Section 4.1.1.1) and SB102 (Table 4.1). In addition, if bioventing treatment of vadose zone soils in this area proved to be effective, it was presumed that extended operation of the pilot-scale system would provide incremental risk reduction without hindering the ongoing biodegradation of CAHs in groundwater. Conversely, a bioventing system installed and operated closer to the four former USTs (47 through 50) might alter the reducing groundwater conditions favorable for reductive dehalogenation of chlorinated VOCs, and/or diminish the supply of fuel hydrocarbons which may be acting as a carbon source for microbial growth and biodegradation of CAHs in groundwater. Air injection VW and MP completion details for the bioventing system and a cross section of the system configuration are presented in Table 4.5 and on Figure 4.3.

4.3.1.1 Soil Gas Sampling

Following installation of the bioventing VW and MPs, initial soil gas samples were collected to establish baseline conditions within vadose zone soils in this area. Soil gas samples were collected from the shallow and deep MP screens at VWMP-1, VWMP-2, and VWMP-3 and field screened for oxygen, carbon dioxide, and VOC concentrations. Soil gas samples from these locations also were sent to Air Toxics, Ltd., in Folsom, California for analysis of VOCs by USEPA Method TO-14. The initial field screening and laboratory analytical soil gas results are presented in Tables 4.11, and 4.12, respectively.

Soil gas results from installed vapor monitoring points (Table 4.11) suggest that vadose zone soils near Building 560 are oxygenated and may not benefit significantly from air injection bioventing treatment. Although some oxygen depletion has occurred, soil gas samples had oxygen concentrations greater than 5 percent, which is generally considered to be sufficient for aerobic degradation of petroleum hydrocarbon contaminants. VOCs were detected by PID in soil gas samples and the most significant concentration of VOCs were detected from the VWMP-1D sample. Elevated carbon dioxide concentrations in soil gas also suggest that some natural hydrocarbon biodegradation is occurring (i.e., that carbon dioxide is being created as an aerobic biodegradation byproduct).

Laboratory soil gas analytical results (Table 4.12) suggest that significant concentrations of VOCs, including ethylbenzene and xylenes, are present in vadose zone soils near Building 560. Laboratory results support the elevated PID readings measured in the field at VWMP-1D. Ethylbenzene and xylenes concentrations were greatest at VWMP-1D, which is located approximately 8 feet bgs (Figure 4.3) and approximately 10 feet from VW-1, and soil borings HB-1, and SB102 (Figure 4.1). Ethylbenzene and xylenes concentrations generally lessen with increasing distance from this apparent source area and decrease with depth at VWMP-2 and VWMP-3.

TABLE 4.11 INITIAL SOIL GAS FIELD RESULTS AUGUST 1997

HAZARDOUS WASTE STORAGE AREA RICKENBACKER ANGB, OHIO

Monitoring Point	Oxygen (percent)	Carbon Dioxide (percent)	PID ^{a/} (ppmv) ^{b/}
VWMP-1S	20.9	0.3	25.8
VWMP-1D	13.5	5.4	100
VWMP-2S	11.8	6.7	23.7
VWMP-2D	15.5	4.2	7.0
VWMP-3S	18.6	1.3	25.2
VWMP-3D	6.4	6.7	8.1
SWVZ-4	17.8	4.2	2.3

PID = photoionization detector measurements of volatile organics using an Hnu DL-101.

by ppmv = parts per million, volume per volume.

TABLE 4.12
INITIAL SOIL GAS ANALYTICAL RESULTS
AUGUST 1997
HAZARDOUS WASTE STORAGE AREA
RICKENBACKER ANGB, OHIO

Analyte			S	Sample Location"			
ij	VWMP-1S	VWMP-1D	VWMP-2S	VWMP-2D	VWMP-3S	VWMP-3D	VZMP-4
Benzene	- -	1	1	1	1	ı	:
Ethylbenzene	2,700	61,000	4,200	460	1,800	099	17
Toluene	41	:	1	1	1	i	ı
Xylenes	16,500	157,500	14,660	1,530	6,370	2,760	71.5
Methylene Chloride	42	ŀ	1	1	1	1	1
cis-1,2-Dichloroethene	ı	ı	1	1	i	1	4.7
1,1,2,2-Tetrachoroethane	ı	1	!	ı	1	1	4.7
1,2,4-Trimethylbenzene	58	1	1	11	1	1	1
Acetone	5,900	1	1	89	1	1	ı
2-Propanol	240	1		1	1	1	:
2-Butanone	2,400		22,000	089	10,000	3,400	1
Tetrahydrofuran	1,800	4,100	26,000	1,400	76,000	16,000	370
Ethanol	300	•	1	1	1	-	1

Note: Samples analyzed by USEPA Method TO-14. All results in parts per billion, volume per volume.

a/ Bioventing system sample locations, except VZMP-4, which is a sparge point.

b/ - = Not detected.

Similar to the difficulty encountered during Geoprobe® soil gas sample collection in May 1997, tight vadose zone soils (clay and silt) and high soil moisture content hindered soil gas sample collection, and may have resulted in some ambient air leaking into the sampling train and diluting the samples. In addition, all but one of the MP screens produced water during sample extraction, and soil gas samples only could be collected by inserting a water trap in the sampling train.

4.3.1.2 Air Permeability, Radius of Influence, and In Situ Respiration Testing

Following collection of the initial soil gas samples, air permeability and radius-of-influence testing was performed using a temporary power supply and a positive-displacement blower. The objective of the testing was to determine the pressure influence and the extent of the subsurface that could be oxygenated by air injection into VW-1. During the testing, pressure response was measured at each MP screened interval with differential pressure gauges to determine the region influenced by the unit. Soil gas oxygen concentrations also were measured at each MP screen to evaluate oxygen radius of influence. The permeability test was performed at various flow rates to determine the optimum rate for the system.

Based on conversations with IT (1997e), no pressure response was indicated at any of the MPs, except when the air flow injection pressure was increased to more than 90 inches of water. At this injection pressure, pressure influence was indicated only at VWMP-1S, located approximately 10 feet from the point of air injection (VW-1). Initial high concentrations of oxygen at this point, prevented determination of oxygen influence. An *in situ* respiration test to evaluate oxygen utilization and aerobic biodegradation of hydrocarbon contaminants by soil microbes was not performed because of the difficulty in injecting air into this moist, low-permeability soil.

4.3.1.3 Discussion and Conclusions

Results obtained from the bioventing treatability test were limited and did not support bioventing as an effective means of remediating vadose zone soils within the HWSA. While elevated concentrations of VOCs appear to be present both in soil and

soil gas, tight, impermeable, wet vadose zone soils prevented collection of conclusive soil gas samples and prevented performance of treatability tests critical for bioventing system design and long-term operation. Bioventing data available for 145 Air Force sites, suggest that a combination of high soil moisture content and fine-grained soils (silt and clay) can prevent bioventing from being a feasible technology (AFCEE, 1996b). If it is determined that contaminated source area soils within the HWSA require remediation, bioventing does not appear to be a viable technology for further consideration. Further information regarding the bioventing treatability test at the HWSA is provided by IT (1997c).

4.3.2 Air Sparging

Air sparging was conducted to evaluate the effectiveness of this technology to provide an oxygenated barrier to downgradient migration of CAHs (e.g., VC). Two air sparging treatability tests, one continuous air injection test and one pulsed air injection test, were conducted near ESMP-17 to confirm the effectiveness of the technology (Figure 4.1). Sparge well/MP completion details for the air sparging system and a cross section of the system configuration are presented in Table 4.8 and on Figure 4.4.

Initial treatability testing was performed in August 1997. Results from the continuous air injection and pulsed air injection tests were inconclusive. An extended treatability test was initiated in September 1997 to evaluate the long-term effectiveness of pulsed air injection in this area. Extended testing was completed in December 1997 and no increase in DO concentrations at the various MP screened intervals was apparent. In addition, quarterly groundwater sampling results from December 1997 indicate that dissolved CAH contaminant concentrations in the vicinity of the sparging system were not reduced by the extended testing (IT, 1998). Results of the air sparging treatability testing are further discussed in the Soil and Groundwater Treatability Test Report (IT, 1997c).

4.4 CHEMICAL FATE IN GROUNDWATER

This section estimates the effects natural chemical attenuation processes have had and may have on the migration, mass, concentration, persistence, and toxicity of various dissolved contaminants in groundwater, especially petroleum hydrocarbons and chlorinated solvents. Inorganics (metals) also are considered. These processes must be evaluated when determining whether a compound poses or has the potential to pose, an actual risk to human health and the environment. If the contaminant is not likely to migrate to a potential receptor, the contaminant poses no risk.

Several physical/chemical and biological processes influence how a chemical behaves in groundwater, and the role these processes play is in turn influenced by the chemical characteristics of each particular contaminant. The most important mechanisms affecting chemical fate in groundwater include sorption, dispersion, dilution from recharge, abiotic chemical reactions, and biological degradation. Sorption, dispersion, and dilution are nondestructive attenuation mechanisms that will limit contaminant migration and possibly decrease contaminant concentrations without a permanent reduction in contaminant mass. Conversely, destructive mechanisms such as abiotic chemical reactions and biological degradation typically result in loss of contaminant mass. However, the daughter products of some of these reactions may also be considered contaminants (e.g., DCE and VC are produced via biological dehalogenation of TCE). The positive effects of these natural processes on reducing the actual mass of contaminants and/or minimizing the extent of contaminant migration in groundwater has been termed remediation by natural attenuation (RNA).

4.4.1 Nondestructive Attenuation Mechanisms

The relative solubility, sorptive nature, and volatility of a chemical can govern the effectiveness of nondestructive chemical attenuation processes, particularly sorption. Dispersion and dilution from recharge are more dependent upon the physical hydrogeologic setting encountered at a given site.

4.4.1.1 Solubility and Sorption

The water solubility of a chemical species defines how that particular chemical could partition (leach) from a contaminant source and dissolve into and migrate with groundwater. For example, benzene is fairly water soluble relative to the heavier organics and to most inorganics. Consequently, even though benzene may make up a low mass fraction of the initial source of contamination, it will preferentially leach from residual LNAPLs present in soils and migrate as dissolved contamination (Lyman et al., 1992). The differences in solubility between benzene and other fuel-related compounds is the cause of the disproportionate effect that benzene can have on groundwater quality in comparison to other hydrocarbons commonly found in fuel. For chlorinated compounds, solubility and mobility are controlled by the number of chlorine atoms and carbon atoms in the compounds. For example, TCE and PCE are less mobile relative to other chlorinated ethenes such as DCE and VC. Usually the less soluble compounds become adsorbed or occluded in the soils and have relatively low mobility.

For inorganic compounds such as lead and arsenic, the relative solubility can also be a function of environmental conditions. The solubility of these types of compounds can change as the groundwater becomes more or less reducing, more or less acidic, or more or less saturated with other compounds. For example, under highly reducing conditions, such as those found when organic contamination is present, compounds such as arsenic, manganese, and iron persist as species with low oxidizing potential (e.g., Masscheleyn et al., 1991a, and 1991b). These reduced species are generally more soluble, which means that the compounds desorb from the solid soil matrix and are available to be transported as dissolved constituents. Consequently, localized changes in environmental conditions can dramatically affected the solubility (and therefore the concentration of a contaminant in solution) of certain inorganic compounds. As soluble inorganics migrate into more oxidizing environments, they often precipitate or stabilize as metal hydroxides and are then naturally removed from the groundwater.

Sorptive properties also govern how a chemical is attenuated in soil and groundwater. Most of the contaminants or potential contaminants detected at the HWSA sorb to that portion of the soil matrix that is composed of organic carbon and fine clay particles. If a contaminant can be strongly sorbed to organic carbon and/or clay particles in the aquifer matrix, the compound will be less mobile and less likely to be transported great distances from the source area. Benzene is considered the most mobile of the BTEX compounds (Abdul et al., 1987), but it does sorb slightly to soils. The other BTEX compounds sorb more strongly, although they too are relatively mobile when compared to other compounds present in fuel mixtures. Heavier fuel hydrocarbons such as naphthalene sorb even more strongly to the soil matrix, and migration of those compounds tends to be limited in both soil and groundwater. As noted above, heavier and more chlorinated solvents such as PCE and TCE will sorb more strongly than DCE, DCA, or VC. In general, VC sorbs so weakly that it is not effectively attenuated by sorption and is thus seen further downgradient from a source area than are TCE and DCE. Sorption of metals detected in HWSA groundwater will likely be dependent on redox and pH conditions.

4.4.1.2 Volatilization

Volatilization can potentially remove contaminants from a groundwater system. In general, factors affecting the volatilization of contaminants from groundwater into soil gas include the contaminant concentration, the change in contaminant concentration with depth, the Henry's Law constant and diffusion coefficient of the compound, mass transport coefficients for the contaminant in both water and soil gas, and the temperature of the water (Larson and Weber, 1994). Partitioning of a contaminant between the liquid phase and the gaseous phase is governed by Henry's Law. Thus, the Henry's Law constant of a chemical determines the tendency of a contaminant to volatilize from groundwater into the soil gas. With the exception of VC, the physiochemical properties of chlorinated solvents and the BTEX compounds give them relatively low Henry's Law constants.

Because of the small surface area of the groundwater flow system exposed to soil gas, volatilization of chlorinated solvents and BTEX compounds from groundwater is a relatively slow process that, in the interest of being conservative, generally can be ignored as a significant attenuation process. Chiang *et al.* (1989) demonstrated that less than 5 percent of the mass of dissolved BTEX is lost to volatilization in the saturated groundwater environment. Moreover, Rivett (1995) observed that for plumes more than about 3 feet below the air/water interface, little, if any, solvent concentrations will be detectable in soil gas due to the downward groundwater velocity in the vicinity of the water table. This suggests that for portions of plumes more than 3 feet below the water table, very little, if any, mass will be lost due to volatilization. In addition, vapor transport across the capillary fringe can be very slow (McCarthy and Johnson, 1992), thus further limiting mass transfer rates. Because of this, the impact of volatilization on dissolved contaminant reduction can generally be considered negligible.

4.4.1.3 Dispersion and Dilution

Hydrodynamic dispersion is the process whereby a contaminant plume spreads out in directions that are longitudinal and transverse to the direction of plume migration. Dispersion of organic solutes in an aquifer is an important consideration when modeling RNA. Dispersion of a contaminant dilutes the concentrations of the contaminant, and introduces the contaminant into portions of the aquifer with different geochemical conditions, where it may be attenuated more or less rapidly than in the main plume area. Hydrodynamic dispersion consists of mechanical dispersion and molecular diffusion. Molecular diffusion is only important where groundwater velocities are very low. At the HWSA, as in most cases, mechanical dispersion dominates.

As defined by Domenico and Schwartz (1990), mechanical dispersion is mixing that occurs as a result of local variations in velocity around some mean velocity of flow. With time, a given volume of solute will gradually become more dispersed as different portions of the mass are transported at the differing velocities. In general, the main cause of variations of both rate and direction of transport velocities is the heterogeneity

of the porous aquifer medium. These heterogeneities are present at scales ranging from microscopic (e.g., pore to pore) to macroscopic (e.g., well to well) to megascopic (e.g., a regional aquifer system). In general, these heterogeneities are a function of the site geology. At the HWSA, the unconsolidated aquifer materials are relatively heterogeneous, as is common with glacial till deposits. Thus, dispersion is likely to be an important factor governing contaminant fate and transport. However, quantifying dispersion is difficult; it is dependent upon the scale (length) of the plume and upon the nature of the heterogeneities that control mechanical dispersion. Dispersion is typically only quantified where needed for transport modeling.

Recharge from infiltrating precipitation is the result of a complex series of processes in the unsaturated zone. Description of these processes is beyond the scope of this discussion; however, it is worth noting that the infiltration of precipitation through the vadose zone brings the water into contact with the soil and thus may allow dissolution of additional electron acceptors and possibly organic soil matter (a potential source of electron donors). Infiltration therefore provides fluxes of water, inorganic species, and possibly organic species into the groundwater. Consequently, the water entering the groundwater system will not only aid in dilution of a contaminant plume but it may also add electron acceptors and possibly electron donors to the groundwater. At the HWSA, dilution from recharge may be minimized by the clayey nature of the shallow soil and future land use plans which call for an impermeable concrete over much of the site.

4.4.1.4 Discussion

The preceding discussion shows that solubility and sorptive characteristics are important chemical characteristics to consider when assessing whether groundwater contaminants may present an unacceptable risk to human health or the environment. Site contaminants characterized by relatively high water solubility and low sorptive properties readily partition from soils or LNAPL and can be transported with groundwater. Less soluble and more sorptive compounds are less likely to leach and do not travel as rapidly. Consequently, more mobile compounds detected at the HWSA, such as VC or benzene, provide an indication of the "leading edge" of the contaminant

plume(s) and should be monitored during routine and confirmatory groundwater sampling.

4.4.2 Destructive Attenuation Mechanisms

In comparison to nondestructive chemical attenuation processes, destructive chemical attenuation processes result in the permanent removal of contaminant mass from the environment thereby reducing the incremental risk contributed by these contaminants. The effectiveness of destructive attenuation processes at reducing contaminant mass at a site depends on how amenable the chemical is to biodegradation and whether the site is characterized by physical, chemical, and biological conditions favorable to such processes. In general, there are two destructive attenuation mechanisms that must be considered when evaluating chemical fate. The most significant of these is biodegradation, although abiotic chemical reactions can affect some solutes.

4.4.2.1 Overview of Biodegradation Processes

Microorganisms obtain energy for cell production and maintenance by facilitating thermodynamically advantageous redox reactions involving the transfer of electrons from electron donors to available electron acceptors. This results in the oxidation of the electron donor and the reduction of the electron acceptor. Electron donors may be natural organic carbon, fuel hydrocarbon compounds, and less-chlorinated solvents (e.g., VC, DCE, or DCA). Fuel hydrocarbons or solvents are completely degraded or detoxified if they are utilized as the primary electron donor (i.e., as a primary substrate or carbon source) for microbial metabolism (Bouwer, 1992). Electron acceptors are elements or compounds that occur in relatively oxidized states, and include oxygen, nitrate, ferric iron, sulfate, manganese, carbon dioxide, and highly chlorinated solvents (e.g., PCE, TCE, TCA, and polychlorinated benzenes).

The driving force of biodegradation is electron transfer, which is quantified by the Gibbs free energy of the reaction (ΔG°_r) (Stumm and Morgan, 1981; Bouwer, 1994; Godsey, 1994). The value of ΔG°_r represents the quantity of free energy consumed ($\Delta G^{\circ}_r > 0$) or yielded ($\Delta G^{\circ}_r < 0$) to the system during the reaction. Although

thermodynamically favorable, most of the reactions involved in biodegradation of fuel hydrocarbons or chlorinated solvents cannot proceed abiotically because of the lack of activation energy. Microorganisms are capable of providing the necessary activation energy; however, they will facilitate only those redox reactions that have a net yield of energy (i.e. $\Delta G^{\circ}_{r} < 0$). Microorganisms preferentially utilize electron acceptors while metabolizing hydrocarbons (Bouwer, 1992). DO is utilized first as the prime electron acceptor. After the DO is consumed, anaerobic microorganisms typically use electron acceptors in the following order of preference: nitrate, manganese, ferric iron hydroxide, sulfate, and finally carbon dioxide. Chlorinated solvents can also be used as electron acceptors when aquifer conditions are such that sulfate or carbon dioxide are preferred electron acceptors.

Depending on the types and concentrations of electron acceptors present (e.g., nitrate, sulfate, carbon dioxide), as well as pH and redox conditions, anaerobic biodegradation (using the contaminant as a substrate/electron donor) can occur by denitrification, manganese reduction, ferric iron reduction, sulfate reduction, or methanogenesis. Other, less common anaerobic degradation mechanisms such as manganese or nitrate reduction may dominate if the physical and chemical conditions in the subsurface favor use of these electron acceptors. Vroblesky and Chapelle (1994) show that the dominant terminal electron accepting process can vary both temporally and spatially in an aquifer with fuel hydrocarbon contamination. Biodegradation in which the contaminant is used as an electron acceptor is also dependent upon site geochemical conditions and is more likely to occur under reducing conditions.

4.4.2.2 Biodegradation of BTEX

Numerous laboratory and field studies have shown that hydrocarbon-degrading bacteria can participate in the degradation of many of the chemical components of fuel hydrocarbons, including the BTEX compounds (e.g., Jamison *et al.*, 1975; Atlas, 1981, 1984, 1988; Gibson and Subramanian, 1984; Reinhard *et al.*, 1984; Young, 1984; Bartha, 1986; Wilson *et al.*, 1986, 1987, and 1990; Barker *et al.*, 1987; Baedecker *et al.*, 1988; Lee, 1988; Chiang *et al.*, 1989; Grbic-Galic, 1989 and 1990;

Cozzarelli et al., 1990; Leahy and Colewell, 1990; Altenschmidt and Fuchs, 1991; Alvarez and Vogel, 1991; Baedecker and Cozzarelli, 1991; Ball et al., 1991; Bauman, 1991; Borden, 1991; Brown et al., 1991; Edwards et al., 1991 and 1992; Evans et al., 1991a and 1991b; Haag et al., 1991; Hutchins and Wilson, 1991; Hutchins et al., 1991a and 1991b; Beller et al., 1992; Bouwer, 1992; Edwards and Grbic-Galic, 1992; Thierrin et al., 1992; Malone et al., 1993; Davis et al., 1994). Fuel hydrocarbons biodegrade naturally when an indigenous population of hydrocarbon-degrading microorganisms is present in the aquifer and sufficient concentrations of electron acceptors and nutrients are available to these organisms.

During biodegradation of fuel hydrocarbons, the fuel compounds are used as electron donors. They are the primary substrate ("food") for the microbes, while the electron acceptors provide the oxidant used to reduce (metabolize) the substrate and produce energy. In most subsurface environments, both aerobic and anaerobic degradation of fuel hydrocarbons can occur, often simultaneously in different parts of the plume. The reduction of highly oxidized electron acceptors (e.g., oxygen/DO) results in an overall decrease in the oxidizing potential of the groundwater. The reduction of oxygen and nitrate will reduce the oxidizing potential to levels at which ferric iron (Fe³⁺) reduction can occur. As each chemical species that can be used to oxidize the contaminants is exhausted, the microorganisms are forced to use other available electron acceptors with lower oxidizing capacity. When sufficiently reducing conditions have been developed as a result of these redox reactions, sulfate reduction and methanogenesis can occur almost simultaneously (Stumm and Morgan, 1981).

The ubiquitous nature of these processes in the groundwater environment is well documented (e.g., Rice et al., 1995; Wiedemeier et al., 1995; Kuehne and Buscheck, 1996). Aerobic destruction of BTEX compounds dissolved in groundwater results in the consumption (reduction) of DO and the formation of carbon dioxide. Anaerobic destruction of the BTEX compounds is associated with the accumulation of fatty acids, production of methane, solubilization of iron, and reduction of nitrate and sulfate (Cozzarelli et al., 1990; Wilson et al., 1990).

Groundwater data collected at the HWSA indicate that biodegradation of fuel hydrocarbons (especially BTEX) is an ongoing process. One indication that biodegradation is ongoing is provided by BTEX plume data collected over time (Figure 4.6 and Table 4.11). In general, the data show that although dissolved BTEX concentrations in the source area are stable, the extent of the BTEX plume has not changed significantly, suggesting that biodegradation is limiting migration of BTEX at the site. It is unlikely that this is attributable solely to sorption, because there would still be some movement of the plume, and also because sorption is typically limited by the amount of organic carbon available in the aquifer matrix. A stable plume of this nature implies that some residual soil contamination is likely continuing to introduce a minor amount of BTEX into groundwater, approximately at the same rate that natural attenuation processes are removing mass or decreasing dissolved concentrations.

Moreover, geochemical evidence indicates that biodegradation reactions are taking place at the site. These geochemical changes generally coincide with the BTEX plume at the HWSA, and given the ubiquitous nature of fuel hydrocarbon biodegradation, are almost certainly the result of such degradation. In general, DO is depleted in monitoring wells and points in the vicinity of the BTEX plume (Table 4.13), as is sulfate (Figure 4.9). Also, ferrous iron and methane concentrations are elevated within or just downgradient of the BTEX plume (Figures 4.10 and 4.11). Such patterns are typically observed at sites where dissolved fuel hydrocarbons are being biodegraded, because DO is consumed first, and as conditions become more reducing, ferrous iron is produced (due to use of ferric iron as an electron acceptor), then sulfate is consumed, and finally methane is produced.

Elevated concentrations of ferrous iron (Fe²⁺) are of note because they often are found in anaerobic portions of fuel-contaminated groundwater systems. Recent evidence suggests that the reduction of ferric iron cannot proceed at all without microbial mediation (Lovley and Phillips, 1988; Lovley *et al.*, 1991; Chapelle, 1993). None of the common organic compounds found in low-temperature, neutral, reducing groundwater could reduce ferric oxyhydroxides to ferrous iron under sterile laboratory

NATURAL ATTENUATION GROUNDATER GEOCHEMICAL DATA HAZARDOUS WASTE STORAGE AREA RICKENBACKER ANGB, OHIO JUNE/JULY 1997 **TABLE 4.13**

				FIELD DATA	V				LABORAT	LABORATORY DATA
		Dissolved	Rodox	Ferrous						
	Sample	Oxygen	Potential	Iron	Nitrate	Nitrite	NH3	Chloride	Methane	Ethene
Sample Location	Date	$(mg/L)^{a/}$	(mV)b/	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)
ESMP-2D	Jul-97	6.0	-62.2	1.7	4.0	0.003	0.07	01	0.081	NDc/
ESMP-2S	Jul-97	/pLN	Ä	¥	¥	¥	¥	ž	0.055	S
ESMP-3D	Jun-97	0.77	-19	0.87	4.0	0.001	0.15	15	0.024	S S
ESMP-4D	Jun-97	0.67	11.4	0.33	0.3	0.001	0.7	10	0.012	Q.
ESMP-4S	Jun-97	0.74	38.2	99.0	0.3	0.002	0.23	15	0.791	Q.
ESMP-6D	Jun-97	1.34	ጀ	0.148	6.4	0.003	0.25	10	0.016	N Q
ESMP-8S	Jun-97	5.43	189.1	ž	Ħ	Ħ	Ä	Ä	QZ	S
ESMP-10S	Jun-97	0.69	212.8	0.72	0.1	9000	0.28	10	0.002	2
ESMP-12S	Jun-97	2.77	186.7	L	¥	Ħ	¥	Ä	0.004	QN
ESMP-13S	Jun-97	1	-120.3	2.94	0	0	0.94	15	22.6	960.0
ESMP-14D	Jun-97	1.29	10.3	2.17	0.3	0.002	0.32	10	0.279	0.001
ESMP-15S	Jun-97	0.84	-128.5	3.02Del	0	0	0.21	15	0.16	Š
ESMP-16D	Jun-97	0.65	-96.1	2.9D	0	0	0.48	15	0.15	Q.
ESMP-16S	76-unf	0.84	-115.5	14.5D	0	0	2.72	15	4.85	QZ Q
ESMP-17S	Jun-97	2.42	-323	13.5D	0	0	0.51	10	0.604	0.011
MW-2	Jun-97	1.58	-30.5	0.2	0.003	0.5	0.23	10	0.248	QN Q
MW-3	Jul-97	5.95	198.5	0.14	9.0	9000	0.08	10	QN Q	S S
MW-4	Jun-97	1.32	236	0	0.3	0.00	0.05	2	£	S
MW-5	Jul-97	¥	¥	4.34D	0	0	0.7	10	4.62	æ
MW-6	Jun-97	3.76	225	0	9.0	0.002	0.22	10	Q.	g
MW-8	Jun-97	1.36	178.3	0.13	0	0	0.23	10	N	Q.
WW-9	Jun-97	0.59	4 .8	0.32	0.2	0.001	0.93	15	0.004	Ą
MW-10	Jun-97	99.0	-59.4	1.36	0.1	0	0.36	15	0.112	£
MW-11	Jun-97	99.0	195.5	0	0.3	0.004	0.03	10	0.005	S
MW-12	Jun-97	1.36	218.1	0	0.3	0.002	1.5	15	0.001	£
MW-101D	Jul-97	0.74	-56.8	1.95	0.3	0	8.0	15	0.37	Ð
MW-101S	Jul-97	1.06	-111.8	4.16D	0	0	0.35	10	0.635	Q.
MW-102D	Jul-97	0.75	T.TT-	3.1	0.3	0	0.25	15	0.217	Ą
MW-102S	Jul-97	1.02	-74.2	5.1	0.2	0	0.2	10	1.16	Ð
MW-103D	Jul-97	0.76	6.6	2.22	0.4	0.002	0.5	20	0.089	Q.
MW-103S	Jul-97	0.87	-11.9	1.61	0.4	0.00	0.13	01	10.0	g
MW-104D	Jul-97	0.75	-36.4	1.52	0.3	0.002	90.0	15	0.112	g
MW-105D	Jul-97	0.87	-51.9	2.57	0.3	0.001	0.16	01	0.045	8
MW-105S	Jul-97	0.79	-61.8	2.35	0.3	0	0.12	15	0.239	æ
MW-106D	Jul-97	0.81	-24.4	1.19	0.3	0.001	0.15	15	0.066	0.002
				1 7000	-000		-	4 - 1	7.7	

Note: A more complete listing of geochemical data for December 1996, March 1997, and June/July 1997 are presented in Appendix C. **I milligrams per liter.

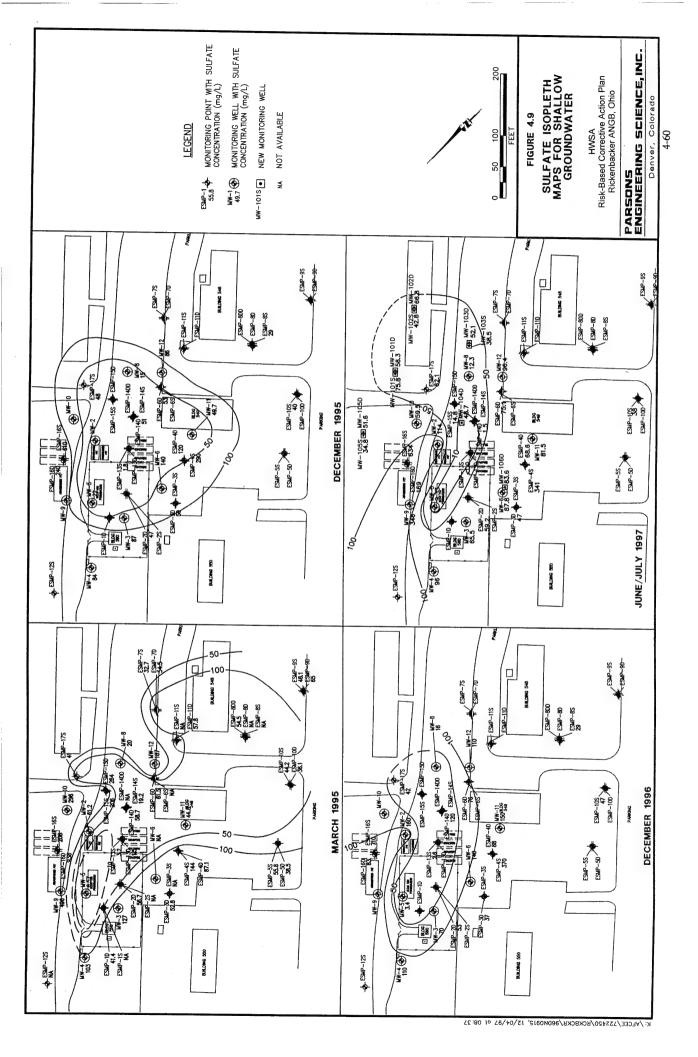
b/ mV = milligrams per liter.

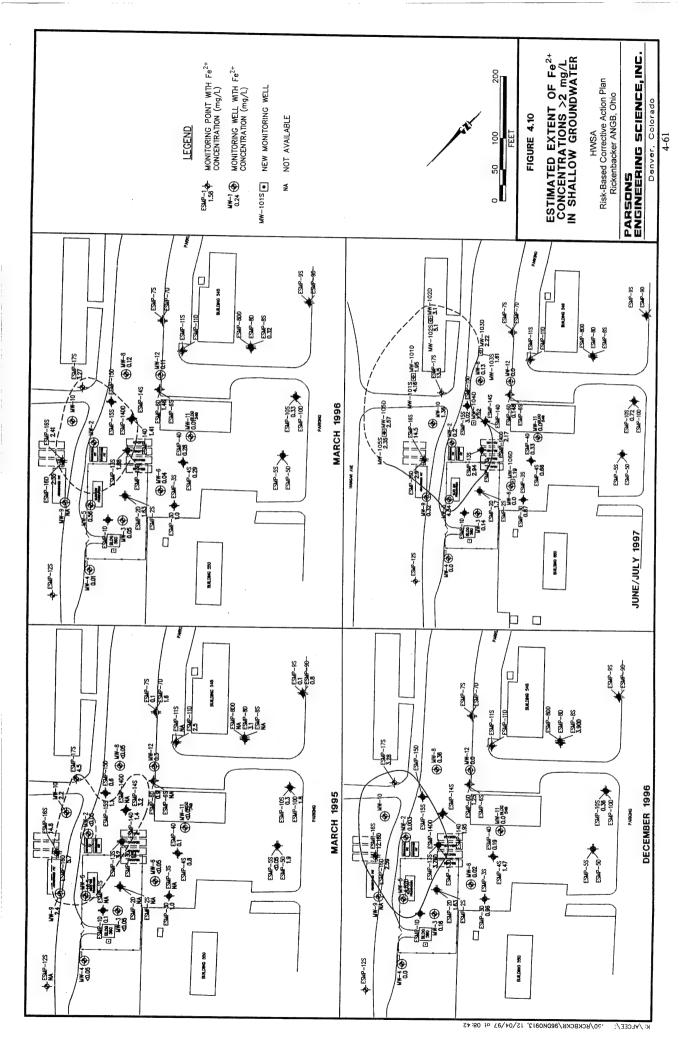
b/ mV = millivolts.

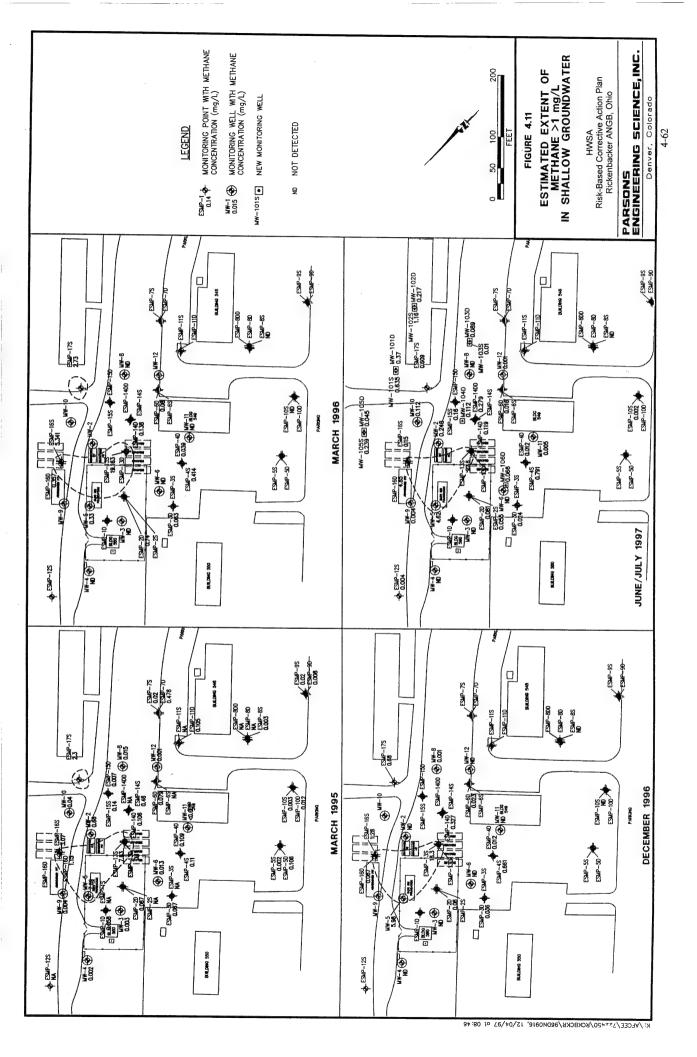
c/ ND = not detected.

d/ NT = sample not tested for this parameter.

c/ D = compound snalyzed at a secondary dilution factor.







conditions (Lovley et al., 1991). This means that the reduction of ferric iron requires mediation by microorganisms with the appropriate enzymatic capabilities, and those microbes likely use fuel hydrocarbons as a substrate/electron donor. This is further evidence of biodegradation of fuel hydrocarbons in HWSA groundwater.

4.4.2.3 Biodegradation of Chlorinated Solvents

Chlorinated solvents can be transformed, directly or indirectly, by biological processes (e.g., Bouwer et al., 1981; Miller and Guengerich, 1982; Wilson and Wilson, 1985; Nelson et al., 1986; Bouwer and Wright, 1988; Little et al., 1988; Mayer et al., 1988; Arciero et al., 1989; Cline and Delfino, 1989; Freedman and Gossett, 1989; Folsom et al., 1990; Harker and Kim, 1990; Alvarez-Cohen and McCarty, 1991a, 1991b; DeStefano et al., 1991; Henry, 1991; McCarty et al., 1992; Hartmans and de Bont, 1992; McCarty and Semprini, 1994; Vogel, 1994; Bradley and Chapelle, 1996; Klier et al., 1996). Biodegradation of chlorinated solvents results from the same general processes that bring about biodegradation of fuel hydrocarbons. However, a more complex series of processes often is involved, and chlorinated solvents may act as both a substrate (electron donor) and an electron acceptor. Chlorinated solvents may undergo biodegradation through three different pathways: use as an electron acceptor, use as an electron donor, or cometabolism, which is degradation resulting from exposure to a catalytic enzyme fortuitously produced during an unrelated process. At a given site, one or all of these processes may be operating, although at many sites the use of chlorinated solvents as electron acceptors appears to be most likely.

Under anaerobic conditions, biodegradation of chlorinated solvents usually proceeds through a process called reductive dehalogenation. During this process, the halogenated hydrocarbon is used as an electron acceptor, not as a source of carbon, and a halogen atom is removed and replaced with a hydrogen atom. In general, reductive dehalogenation of chlorinated ethenes occurs by sequential dechlorination from PCE to TCE to DCE to VC to ethene. Depending upon environmental conditions, this sequence may be interrupted, with other processes then acting upon the products.

During reductive dehalogenation, all three isomers of DCE can theoretically be produced; however, Bouwer (1994) reports that under the influence of biodegradation, cis-1,2-DCE is a more common intermediate than trans-1,2-DCE, and that 1,1-DCE is the least prevalent intermediate of the three DCE isomers. Reductive dehalogenation of chlorinated solvent compounds is associated with the accumulation of daughter products (e.g., cis-1,2-DCE, VC) and an increase in chloride concentrations (which may not be distinguishable, depending upon background concentrations).

Reductive dehalogenation affects each of the chlorinated ethenes differently. Of these compounds, PCE is the most susceptible to reductive dehalogenation because it is the most oxidized. Conversely, VC is the least susceptible to reductive dehalogenation The rate of reductive because it is the least oxidized of these compounds. dehalogenation also has been observed to decrease as the degree of chlorination decreases (Vogel and McCarty, 1985; Bouwer, 1994). Murray and Richardson (1993) have postulated that this rate decrease may explain the accumulation of VC in PCE and TCE plumes that are undergoing reductive dehalogenation. In addition to being affected by the degree of chlorination of the compound, reductive dehalogenation can also be controlled by the redox conditions of the site groundwater system. In general, reductive dehalogenation has been demonstrated under anaerobic nitrate- and sulfatereducing conditions, but the most rapid biodegradation rates, affecting the widest range of chlorinated solvents, occur under methanogenic conditions (Bouwer, 1994). Dehalogenation of PCE and TCE to DCE can proceed under mildly reducing conditions such as nitrate reduction or iron (III) reduction (Vogel et al., 1987), while the transformation of DCE to VC, or the transformation from VC to ethene requires more strongly reducing conditions (Freedman and Gossett, 1989; DeStefano et al., 1991; De Bruin et al., 1992).

Where chlorinated compounds are used as electron acceptors, there must be an appropriate source of carbon for microbial growth in order for reductive dehalogenation to occur (Bouwer, 1994). Potential carbon sources can include fuel hydrocarbons, low-

molecular-weight compounds (e.g., lactate, acetate, methanol, or glucose) present in natural organic matter, or less-chlorinated solvents (as discussed below).

Under aerobic conditions some chlorinated compounds can be utilized as the primary substrate (i.e., electron donor) in biologically mediated redox reactions (McCarty and Semprini, 1994). In this type of reaction, the facilitating microorganism obtains energy and organic carbon from the degraded solvent. In contrast to reactions in which the chlorinated solvent is used as an electron acceptor, only the least oxidized solvents (e.g., VC, DCE, or chlorobenzene) may be utilized as electron donors in biologically mediated redox reactions. For example, while Murray and Richardson (1993) write that microorganisms are generally believed to be incapable of growth using TCE and PCE, other less chlorinated compounds have been shown to be used as substrates. Davis and Carpenter (1990) describe the aerobic oxidation of VC in groundwater. McCarty and Semprini (1994) describe investigations in which VC and 1,2-DCA were shown to serve as primary substrates. Klier et al. (1996) describe aerobic mineralization of all three isomers of DCE. In addition, Bradley and Chapelle (1996) show evidence of oxidation of VC under iron-reducing conditions so long as there is sufficient bioavailable iron (III).

When a chlorinated solvent is biodegraded through cometabolism, it serves as neither an electron acceptor nor a primary substrate in a biologically mediated redox reaction. Instead, the degradation of the solvent is catalyzed by an enzyme or cofactor that is fortuitously produced by organisms for other purposes. The organism receives no known benefit from the degradation of the solvent; rather the cometabolic degradation of the solvent may in fact be harmful to the microorganism responsible for the production of the enzyme or cofactor (McCarty and Semprini, 1994). Cometabolism is best documented in aerobic environments, although it potentially could occur under anaerobic conditions. It has been reported that under aerobic conditions chlorinated ethenes, with the exception of PCE, are susceptible to cometabolic degradation (Murray and Richardson, 1993; Vogel, 1994; McCarty and Semprini, 1994). In the cometabolic process, TCE is indirectly transformed by bacteria as they

use BTEX or another substrate to meet their energy requirements. Therefore, TCE does not enhance the degradation of BTEX or other carbon sources, nor will its cometabolism interfere with the use of electron acceptors involved in the oxidation of those carbon sources. Given this relationship, it would follow that depletion of suitable substrates (BTEX or other organic carbon sources) likely limits cometabolism of chlorinated solvents.

Available site data for the HWSA show that conditions are favorable for reductive As noted in the discussion of BTEX dehalogenation of chlorinated solvents. biodegradation, reducing conditions have been created via consumption of fuel hydrocarbons. The reduced sulfate concentrations (Figure 4.9) and the presence of methane (Figure 4.11) indicate optimal conditions for reductive dehalogenation. Evidence of reductive dehalogenation is provided by the presence of cis-1,2-DCE and VC in concentrations greater than the observed TCE concentrations (Table 4.10), especially in the vicinity of ESMP-17S. While TCE and PCE are commonly used as solvents, the presence of cis-1,2-DCE and VC demonstrates that chlorinated degradation is occurring in site groundwater because the only potential source of these compounds is as a byproducts of reductive dehalogenation. In addition, ethene has been detected in monitoring points ESMP-13S and ESMP-17S, indicating that some of the VC is also being dehalogenated (Table 4.13). Because the microbial dehalogenation of VC is the slowest of such reactions, VC concentrations actually increase downgradient from the source area where TCE and DCE concentrations are greatest. However, it also should be noted, that VC will be consumed as a substrate/electron donor as it migrates into more oxidizing groundwater which surrounds the HWSA.

Determining if solvents such as VC, DCE, or DCA are being used as electron donors is difficult, generally because the reactions generally produce only carbon dioxide and water, or other intermediates that are rapidly biodegraded. However, based on chlorinated VOC contaminant data presented in Table 4.10 and Figure 4.7, it appears that the CAH plume is relatively stable as a result of the reductive dechlorination of TCE and DCE and biological oxidation of VC in downgradient

groundwater. If RNA is to be relied upon as the primary remediation/risk-reduction strategy for the site, additional monitoring will be necessary to validate this assumption.

Cometabolism of solvents is also difficult to detect, and the most well-understood pathways involve aerobic metabolism of organic compounds such as butane, methane, propane, toluene, ethene, or VC. Cometabolism is a potential mechanism for biodegradation of chlorinated solvents in groundwater downgradient of the source area.

With BTEX compounds the occurrence of complete biodegradation to less toxic compounds is ubiquitous, and the HWSA data show these effects with a stable BTEX plume, but for chlorinated compounds, biodegradation produces toxic intermediates (e.g., VC), and at the HWSA, the downgradient extent of this plume, is not as yet completely defined (relatively low levels of DCE and VC have been detected in newly installed monitoring well MW-101S). In order to determine the applicability and effectiveness of destructive attenuation processes in reducing the contaminant mass of chlorinated compounds (particularly VC), additional monitoring data are needed. For RNA to be considered as an effective risk-reduction strategy for the site, additional data would need to be collected in downgradient areas to verify plume stability. A steady or retreating plume observed over several years is the most unequivocal evidence of the effectiveness of natural attenuation processes.

4.4.2.4 Abiotic Degradation

Chlorinated solvents dissolved in groundwater may also be degraded by abiotic mechanisms, although the reactions may not be complete and often result in the formation of a toxic intermediate. The most common abiotic reactions affecting chlorinated solvents are hydrolysis and dehydrohalogenation. Hydrolysis is a substitution reaction in which a halogen substituent is replaced with a hydroxyl (OH) group from a water molecule. Dehydrohalogenation is an elimination reaction in which a halogen is removed from a carbon atom, followed by removal of a hydrogen atom from an adjacent carbon atom, with a double bond between the carbon atoms being produced. Other possible reactions include oxidation and reduction, although Butler

and Barker (1996) note that no abiotic oxidation reactions involving common halogenated solvents have been reported in the literature, and that reduction reactions are most commonly microbially mediated. Butler and Barker (1996) also note that attributing changes in the presence, absence, or concentration of halogenated solvents to abiotic processes is usually difficult, particularly on the field scale. Solvents may undergo both biotic and abiotic degradation, and discerning the effects of each mechanism (on the field scale), if possible, would be very difficult. Also, as Butler and Barker (1996) note, the breakdown products of some reactions such as hydrolysis (e.g., acids and alcohols) may be more easily degraded (biotically or abiotically); these products also require additional analyses that may not be feasible for a field investigation. This makes collection of field evidence to demonstrate hydrolysis very difficult to collect and interpret. Further, evidence of abiotic degradation of the particular solvents detected at the HWSA (e.g., TCE, DCE, VC, and DCA) is limited and suggests that if it occurs, it occurs at extremely slow rates.

4.4.3 Discussion and Conclusions

The purpose of this subsection was to describe the likely processes affecting contaminant fate and to describe how the available data indicate which destructive and nondestructive attenuation processes are significant in groundwater in the vicinity of the HWSA. The fate of contaminants in groundwater at the site is affected by several factors. In general, available data show that concentrations of these contaminants are being attenuated by natural nondestructive and destructive processes. However, while evidence of these processes is apparent, the long-term effects cannot be predicted without additional long-term monitoring. The important points of this subsection are summarized as follow:

 Nondestructive attenuation mechanisms may be partially responsible for observed contaminant attenuation. Of these processes, dispersion is likely to be the most important, and sorption also will be important for most organic and inorganic compounds. However, VC is not likely to be affected by sorption due to its chemical characteristics.

- Geochemical evidence indicates that BTEX compounds are biodegrading in saturated soils and groundwater at the HWSA via aerobic respiration, ferric iron reduction, sulfate reduction, and methanogenesis. Further evidence of the effectiveness of biodegradation and nondestructive attenuation processes affecting fuel hydrocarbons is provided by data showing that the BTEX plume is stable (i.e., not expanding) and that BTEX concentrations are fluctuating without a clear trend towards increasing or decreasing.
- Reductive dehalogenation of chlorinated hydrocarbons is occurring, fostered by
 the reducing conditions produced through biodegradation of fuel compounds. It
 appears that TCE is being degraded to DCE (primarily cis-1,2-DCE) and VC,
 and that some VC is converted to ethene. However, this process may not
 completely remove all of the solvent mass, particularly because the step from VC
 to ethene is slow.
- If the VC plume moves outside of the area with the most reducing conditions, as
 illustrated by Figures 4.9 through 4.11, it can be oxidized. To observe whether
 this process or any of the nondestructive processes is an effective means for
 reducing VC contaminant mass in site groundwater, additional monitoring will be
 required.
- Apparently elevated concentrations of several inorganic compounds may in part be attributable to the localized changes in reduction-oxidation conditions brought about by the microbially mediated reactions involving organic contamination. Once sufficient organic mass has been biodegraded, groundwater redox conditions will be restored to pre-release (more oxidizing) characteristics. Consequently, the attenuation of inorganic compounds appears to be tied to the short- and long-term fate of the hydrocarbon compounds that are driving changes in groundwater geochemical conditions.
- The BTEX compounds are unlikely to migrate beyond the boundaries of the observed plume; however, once a sufficient amount of petroleum compounds are

removed from the system, reductive dehalogenation will cease and the remaining chlorinated compounds will only be affected by oxidative processes. Additional groundwater monitoring is the key to ultimately understanding the effect of destructive and non-destructive attenuation processes upon dissolved contaminants.

To evaluate whether current levels of site contamination pose an unacceptable risk to human health, Section 5 documents the potential risks to human receptors based on the observed site concentrations and the types of exposures that could occur at the HWSA under current and foreseeable future conditions. This initial review suggests that dissolved petroleum hydrocarbons and chlorinated compounds are unlikely to migrate off-Base in concentrations posing unacceptable risk, based on current and future land use scenarios. However, complete delineation of the chlorinated plume is recommended, as is continued groundwater monitoring at the site.

SECTION 5

RISK ASSESSMENT

A comprehensive risk assessment was performed to quantitatively evaluate potential risk to human and ecological receptors based on site-specific conditions and assumptions regarding exposure to chemicals of potential concern (COPCs) at the former HWSA, Building 560, Rickenbacker ANGB. All COPCs were evaluated using standard USEPA (1989, 1991a-c, and 1992a-e) and Ohio EPA (1993a and 1993b) risk assessment practices. Individual risks or hazards were then combined to determine the overall carcinogenic risks and noncancer hazards for all complete exposure pathways at the site.

The results of this risk assessment show that existing concentrations of organic and inorganic chemicals found at the HWSA do not pose an unacceptable level of carcinogenic risk to current or future receptors. The assessment also demonstrates that exposure to noncancer hazards is at or below the threshold level that could potentially trigger remedial action. Based on the conceptual site model (CSM), which considers current and future (industrial) land use scenarios, affected media, release mechanisms, and potential exposure routes, a nonintrusive worker performing grounds maintenance activities (a groundskeeper) and a general (hypothetical) onsite worker represent the two current human receptors considered in this risk assessment. The developed nature Construction workers of the site precludes exposure of ecological receptors. performing intrusive activities at the site represent the most likely future receptors. The most significant exposure route of concern for the intrusive worker is dermal contact with groundwater by construction workers; 91 percent of the cancer risk and 76 percent of the noncancer hazard are associated with dermal exposure to groundwater. Typically, groundwater at the site is present between 8 and 12 feet bgs, which is well below any foreseeable intrusive site activities such as utility burial. This notwithstanding, groundwater contact is conservatively included in the principal future receptor scenario (a construction worker installing a taxiway). Considering this future intrusive scenario, the estimated total cancer risk is not expected to exceed 1.3 x 10⁻⁶ and the total hazard index (HI) for noncancer risks is not expected to exceed 0.2. This exposure scenario, and all other scenarios presented in this section, consider the additive effects of multiple COPCs and multiple exposure routes. Risk-based health standards, or SSTLs, for COPCs that potentially could pose a risk to current and future receptors are discussed in Section 5.7.

5.1 IDENTIFICATION OF COPCS IN SOIL AND GROUNDWATER

In order to identify COPCs in soil and groundwater, the most recent sample data characterizing soil and groundwater contaminant levels were used. However, based on the limited collection of soil data during the 1997 additional assessment activities (Section 4); all available historical soil data, as discussed in Sections 3 and 4 and presented in Appendix B, were combined with the 1997 soil data for the purpose of COPC identification. Groundwater contaminant data collected during the June/July 1997 comprehensive groundwater sampling event (Section 4.2.2) were used for the identification of the current COPCs in groundwater. Groundwater analytical data from the June/July 1997 sampling event are provided in Appendix C. All statistical analysis of these data are presented in Appendix E. These analyses include identification of data distributions (normal, lognormal, or nonparametric) and the development of 95-percent UCLs on the means for contaminants in groundwater.

For comparative purposes, background data for all inorganics were obtained. Background concentrations for inorganic compounds in both soil and groundwater were collected and statistically analyzed for Rickenbacker ANGB during previous IRP efforts. Background locations were selected and approved by the Base Closure Team (BCT) consisting of the Air Force, USEPA, and Ohio EPA (IT, 1997d). Background data for soil and groundwater are presented in Appendix E.

USEPA (1992a) has provided recommendations and guidance on appropriate estimated exposure concentrations for use in risk estimates. Because of the uncertainty associated with estimating a true average concentration at a site, the 95-percent UCL on the mean typically is used. This is referred to as the reasonable maximum exposure (RME) concentration term. If data for a site are limited (e.g., less than five samples), it may be necessary to use the maximum value as the concentration term in risk calculations. In general, the data sets for soil and groundwater taken from the HWSA were sufficient in number to compute the 95-percent UCL for use as the concentration term, as prescribed by USEPA (1992a) and Ohio EPA (1993b). USEPA finds that the 95-percent UCL provides reasonable confidence that the true site average will not be underestimated, and increases the consistency and comparability of risk assessments. In those instances where determination of a 95-percent UCL was complicated by consideration of substantial historical data (soils) or where the 95-percent UCL could not be specifically quantified for a data set considering its distribution, the maximum measured site concentration of the analyte was used.

5.1.1 Identification of COPCs in Soil

Per Ohio EPA (1993b), all organic chemicals detected in soil were retained as final COPCs. Validated results for non-nutritive metals detected in background and site soils were compared statistically to identify constituents present at concentrations significantly above background levels. Any non-nutritive metal detected at the site but not detected in background samples was assumed to be site-related and retained as a COPC. Per Ohio EPA (1993b), the maximum concentration of each detected inorganic was compared with the 95-percent UCL on the mean background concentration (Appendix E). Maximum soil concentration values (Appendix E) were determined based on 1997 and all previous soil sampling data collected from the unsaturated soil zone within the boundaries of the HWSA (Figure 2.2). Soil sample results from below the water table (i.e., greater than 10 feet bgs) were not included in this analysis. As presented in Table 5.1, all of the detected site inorganics, with the exception of

TABLE 5.1 COMPARISON OF MAXIMUM SITE SOIL CONCENTRATIONS

TO BACKGROUND CONCENTRATIONS

HAZARDOUS WASTE STORAGE AREA RICKENBACKER ANGB, OHIO

Detected Analytes	Maximum Detected Site Concentration ^{a/} (mg/kg) ^{b/}	95-Percent UCL Background Concentration ^{c/} (mg/kg)	Maximum Site Concentration Exceeds Background
Inorganics			
Aluminum	18,000	15,095	Yes Yes
Antimony	6	ND ^{d/}	
Arsenic	42	15.9	Yes Yes
Barium	190	149	
Beryllium Cadmium	1.2 9.1	0.89 0.77 18.8	Yes Yes Yes
Chromium Cobalt Copper	25.6 17 73.1	14.8 29.3	Yes Yes
Lead	382	22.5	Yes
Manganese	640	746.0	No
Mercury	2.6	ND	Yes
Nickel	60	44.1	Yes, 1999
Selenium	1.9	ND	Yes
Silver	7.2	1.2	Yes
Thallium	10.5	ND	Yes
Vanadium	38	36.1	
Zinc	522	92.1	

^{a/} Maximum detected site concentration based on all available historic data (Appendix B) and data collected during 1997 (Appendix C) for the hazardous waste storage area.

b/ mg/kg = milligrams per kilogram.

^{c/} Source: (IT, 1997d). Background concentrations represent the 95-percent upper confidence limit (UCL) on the mean for surface soil samples (samples collected from 0 to 2 feet below ground surface) collected at Rickenbacker ANGB in areas outside the zone of influence of the HWSA and other environmental sites.

^{d/} ND = This analyte was not detected in any of the Rickenbacker ANGB background soil samples (IT, 1997d).

manganese, were retained as COPCs. The final list of soil COPCs retained for further analysis is shown in Table 5.2.

TABLE 5.2 SOIL COPCS RETAINED FOR QUANTITATIVE RISK/HAZARD CALCULATIONS HAZARDOUS WASTE STORAGE AREA

HAZARDOUS WASTE STORAGE AREA RICKENBAKCER, ANGB, OHIO

Volatile Organic Compounds

1,1,1-Trichloroethane, 1,1,2,2-Tetrachloroethane, 1,2,3-Trichlorobenzene, 1,2,4-Trimethylbenzene, 1,2-Dichloroethane, 1,2-Dichloroethene, cis-, 1,2-Dichloroethene, trans-, 1,3,5-Trimethylbenzene, 4-Methyl-2-Pentanone, Acetone, Benzene, Butylbenzene, n-, Butylbenzene, sec-, Ethylbenzene, Isopropylbenzene, Methyl ethyl ketone, Methylene chloride, Propylbenzene, n-, Styrene, Toluene, Trichloroethene, Vinyl chloride, Xylene, o- Xylenes, m- & p-

Semi-Volatile Organic Compounds

2-Chlorophenol, 3-Nitroaniline, 4-Nitroaniline, bis(2-Ethylhexyl)phthalate, bis(2-Chlorethyl)ether, di-n-Butylphthalate

Polynuclear Aromatic Hydrocarbons

2-Methylnaphthalene, Acenaphthene, Anthracene, Benzo(a)anthracene, Benzo(a)pyrene, Benzo(b)fluoranthene, Benzo(ghi)perylene, Benzo(k)fluoranthene, Chrysene, Dibenz(a,h)anthracene, Fluoranthene, Fluorene, Indeno(1,2,3-cd)pyrene, Naphthalene, Phenanthrene, Pyrene

Metals

Aluminum, Antimony, Arsenic, Barium, Beryllium, Cadmium, Chromium, Cobalt, Copper, Lead, Mercury, Nickel, Selenium, Silver, Thallium, Vanadium, Zinc

5.1.2 Identification of COPCs in Groundwater

Per Ohio EPA (1993b), all organic chemicals detected in groundwater were retained Validated results for metals detected in background and site as final COPCs. groundwater were compared statistically to identify constituents present at concentrations significantly above background levels. Any metal detected at the site but not detected in background samples was assumed to be site-related and retained as a COPC. Per Ohio EPA (1993b), the maximum concentration of each detected inorganic was compared with the 95-percent UCL on the mean background concentration (Appendix E). Groundwater contaminant data collected from the HWSA during the June/July 1997 comprehensive sampling event was used for the comparison process. For inorganics, total and dissolved concentrations were considered. The comparison of maximum detected groundwater concentrations to the 95-percent UCL background concentrations (Appendix E) is presented in Table 5.3. Based on this comparison, several inorganics were either below background or considered essential human nutrients (calcium, magnesium, potassium, and sodium), and therefore were eliminated The final list of groundwater COPCs retained for further from further analysis. analysis is shown in Table 5.4. The highest value for each remaining inorganic (total or dissolved) was retained for further analysis.

5.2 EXPOSURE ASSESSMENT

The objective of an exposure assessment is to estimate the type and magnitude of exposures to chemicals that are present or migrating from a site (USEPA, 1989). An exposure pathways analysis describes the migration paths a chemical takes from the source of contamination to a potentially exposed individual (USEPA, 1989). Site-related contamination can present a potential risk to receptors only if exposure pathways are completed. A completed exposure pathway must consist of a source, a release mechanism (e.g., leaching or volatilization), a transport medium (e.g., groundwater or air), a potential receptor (e.g., current and future onsite workers), a potential exposure point (i.e., locations where receptors could come into contact with site-related contamination), and potential routes of exposure (i.e., ingestion, dermal

TABLE 5.3 COMPARISON OF MAXIMUM SITE GROUNDWATER CONCENTRATIONS TO BACKGROUND CONCENTRATIONS

HAZARDOUS WASTE STORAGE AREA RICKENBACKER ANGB, OHIO

	Maximum	,	Maximum Site	
	Detected Site	Background ^{c/}	Concentration	Retained for
	Concentration ^{a/}	Concentration	Exceeds	Comparison to
Detected Analytes	$(mg/L)^{h/}$	(mg/L)	Background	PRGs
Aluminum ^{d/}	2.2	5.5	No	
Aluminum-D ^{e/}	0.11	5.5	No	
Antimony	ND ^{f/}	ND	No	
Antimony-D	0.057	ND	Yes	Yes
Arsenic	0.013	0.0055	Yes	No ^{g/}
Arsenic-D	0.019	0.0055	Yes	Yes was
Barium	0.36	0.097	Yes	Yes
Barium-D	0.36	0.097	Yes	No ^{g/}
Beryllium	ND	ND	No	
Beryllium-D	ND	ND	No	
Cadmium	0.034	ND	Yes	Yes
Cadmium-D	0.026	ND	Yes	No ^{g/}
Calcium	210	170	Yes	No ^{h/}
Calcium-D	210	170	Yes	No ^{h/}
Chromium	ND	ND	No	•
Chromium-D	ND	ND	No	
Cobalt	ND	ND	No	
Cobalt-D	ND	ND	No	
Copper	0.0096	ND	Yes	No ^{g/}
Copper-D	0.017	ND	Yes	Yes
Iron	7.4	363	No	
Iron-D	5.7	363	No	
Lead	0.026	ND	Yes: 7	No ⁱ
Lead-D	0.018	ND	Yes	Yes
Magnesium	76	63.9	Yes Yes	No ^{h/}
Magnesium-D	76	63.9	Yes	No ^{h/}
Manganese	6.7	1.34	Yes	Yes
Manganese-D	1.0	1.34	No	
Mercury	0.00032	ND	Yes Very	No ^{g/}
Mercury-D	0.00077	ND	Yes	Yes
Nickel	ND	ND	No	
Nickel-D	ND	ND	No	
Potassium	1.7	$< 0.6^{j/}$	Yes	No ^{h/}
Potassium-D	2.3	< 0.6	Yes	No ^{h/}
2 Olubbiani D	~			

TABLE 5.3 (Continued)

COMPARISON OF MAXIMUM SITE GROUNDWATER CONCENTRATIONS

TO BACKGROUND CONCENTRATIONS

HAZARDOUS WASTE STORAGE AREA RICKENBACKER ANGB, OHIO

Detected Analytes	Maximum Detected Site Concentration ^{a/} (mg/L) ^{b/}	Background ^{c/} Concentration (mg/L)	Maximum Site Concentration Exceeds Background	Retained for Comparison to PRGs
Selenium	0.0098	ND	Yes	Yes
Selenium-D	0.0054	ND	Yes	No ^{g/}
Silver	ND	< 0.005	No	
Silver-D	0.019	< 0.005	Yes.	Yes :
Sodium	35	11.3	Yes With	No ^{h/}
Sodium-D	36	11.3	Yes	No ^h
Thallium	0.013	ND	Yes	Yes
Thallium-D	0.0062	ND	Yes	No ^{g/}
Vanadium	ND	ND	No	
Vanadium-D	ND	ND	No	
Zinc	0.060	0.028	Yes	Y Yes
Zinc-D	0.022	0.028	No	

^{a/} Maximum detected site concentration based on June/July 1997 groundwater sampling event.

b/ mg/L = milligrams per liter.

c/ Source: (IT, 1997d). Background concentrations represent the 95 percent upper confidence limit (UCL) on the mean (see Appendix E).

d/ Unless otherwise noted, value shown represents total (unfiltered) analyte concentration.

e/ D = dissolved concentration of analyte.

f/ ND = not detected.

^{2&#}x27; The highest detected concentration for each analyte is retained for further analysis.

h/ Essential human nutrients are omitted from further analysis.

i' Since the lead concentration is compared to an 'at the tap' federal action level for drinking water in the subsequent table, the dissolved lead concentration is retained for further analysis.

 $^{^{\}rm j\prime}$ < = 95% UCL value for background not determined, but is less than the value shown.

TABLE 5.4

GROUNDWATER COPCS RETAINED FOR QUANTITATIVE RISK/HAZARD CALCULATIONS

HAZARDOUS WASTE STORAGE AREA RICKENBACKER ANGB, OHIO

Volatile Organic Compounds

1,1-Dichloroethene, 1,2-Dichloroethane, 1,2-Dichloroethene, cis-, 1,2-Dichloroethene, trans-, 1,2-Dichloropropane, 4-Methyl-2-Pentanone, Acetone, Benzene, Carbon disulfide, Chloroethane, Ethylbenzene, Toluene, Trichloroethene, Vinyl chloride, Xylene, o-, Xylenes, m- & p-, Xylenes, total

Semi-Volatile Organic Compounds

bis(2-Ethylhexyl)phthalate, Carbazole, Dibenzofuran, di-n-Octylphthalate

Polynuclear Aromatic Hydrocarbons

2-Methylnaphthalene, Acenaphthene, Fluoranthene, Fluorene, Naphthalene, Phenanthrene

Metals

Antimony, Arsenic, Barium, Cadmium, Copper, Lead, Manganese, Mercury, Selenium, Silver, Thallium, Zinc

contact, or inhalation). Each of these elements must be present before a particular exposure pathway can be considered completed. If any of these elements is missing, the exposure pathway is considered incomplete, and there is no risk. Thus, the overall objective of this assessment is to determine which, if any, exposure pathways are or may be completed at the HWSA (USEPA, 1992a).

A site-specific exposure pathways analysis was completed for the HWSA to determine the likelihood of receptor contact with site-related contamination. Those exposure pathways that were considered potentially complete and significant were retained for quantitative evaluation. The potential cumulative risks to human receptors due to exposure to each COPC was then quantitatively characterized. For non-cancer

effects, the potential hazards presented by all COPCs were summed for all exposure routes for each receptor. Cancer risk was summed for all COPCs and across all exposure routes for each potential receptor. Consequently, the final risk estimates account for multiple chemicals in, and receptor exposure to, the affected media. This risk assessment was prepared in accordance with *Risk Assessment Guidance for Superfund* (RAGS) documents (USEPA, 1989, 1991a-c, 1992a-e) and Ohio EPA (1993a and 1993b) guidance.

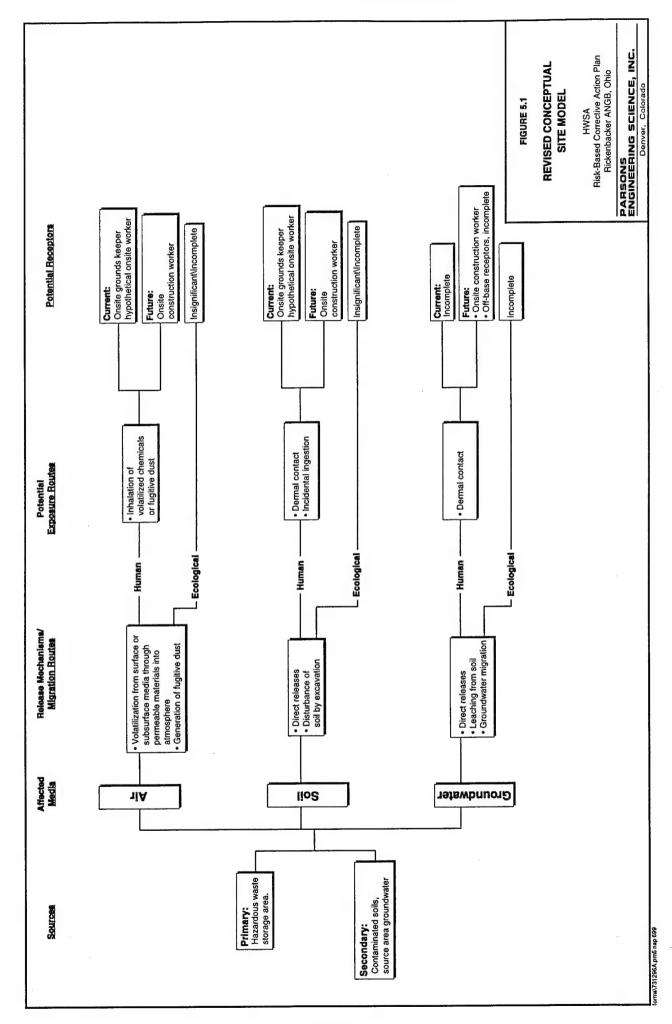
5.2.1 Conceptual Site Model

A CSM is used to qualitatively define the type of potential exposures to contaminants at and migrating from a site (i.e., to systematically evaluate the impact of chemicals in relevant media to potential receptors). The CSM describes onsite release points, the affected physical media, the types of contaminant transport mechanisms that may be involved at the site, potentially exposed current and future receptors, and how each receptor group could come into contact with site-related contamination. The CSM is used to summarize existing site characterization data, including assumptions about land and groundwater use, and to complete the qualitative exposure pathway screening assessment. A preliminary CSM, which was used to identify data gaps and guide data collection activities, was included in the June 1997 work plan (Parsons ES, 1997b). The revised CSM for this site (Figure 5.1), which is briefly discussed in the following sections, identifies only those exposure pathways that may be involved in actual or hypothetical future exposure scenarios.

5.2.1.1 Sources and Release Mechanisms

As described in Sections 3 and 4, site investigations and ongoing monitoring events have identified metals, SVOCs, fuel-related hydrocarbons such as benzene, and CAHs such as VC, as site-related contaminants in soil and groundwater at the site.

COPCs are present at the HWSA in both surface (0 to 2 feet bgs) and subsurface soils (2 to 10 feet bgs). Although it is conceivable that there may be direct exposure to COPCs in surface soil at the HWSA (e.g., inhalation of fugitive dust), direct exposure



to soil COPCs (e.g., through dermal contact or incidental ingestion) is more likely to be of significance during intrusive/construction activities at the HWSA. Direct contact with COPCs in groundwater is a remote possibility in an intrusive construction scenario. Leaching of COPCs from soil and into groundwater, and subsequent offsite migration also could occur.

Volatilization of organic vapors from soil is not considered to be a significant release mechanism for undisturbed surface media. However, if subsurface soils are exposed during excavation activities, volatilization of VOCs could represent a more significant release/transport mechanism. Volatilization of COPCs from soils conservatively was considered in this risk assessment.

There is little opportunity for fugitive dust generation within the HWSA because the majority of the site is covered with horticultural grasses and weeds. The generation of fugitive dust is further reduced because the HWSA (excluding the adjacent tank area outside of the fence) is surrounded by a chain-link fence with a locked gate; thus, minimizing surface-disturbing activity at the site. Nevertheless, it was conservatively assumed that groundskeepers may be exposed to fugitive dust while performing their activities (e.g., grass cutting) at or near the HWSA. Intrusive excavation activities within the HWSA could also result in fugitive dust exposures for future site construction workers. Consequently, exposure to fugitive dust emissions is evaluated for all current and future onsite receptors.

Exposure to COPCs in surface water can be a significant pathway on sites where direct releases from groundwater or surface runoff is a significant migration route. At the HWSA, surface water pathways are considered incomplete because of the absence of surface water drainages at or near the site, and the lack of evidence of groundwater at, or near the surface of the HWSA (groundwater is typically present at 8 to 12 feet bgs). Therefore, surface water is not included as an exposure medium in the CSM (Figure 5.1).

5.2.1.2 Contaminant Environmental Transport

Contaminant transport, transformation, and fate in the environment following a release is important to consider when assessing the potential for exposure. Due to the reducing environment of the subsurface soil and groundwater, inorganics may be subject to transport via groundwater flow. As groundwater containing mobilized inorganics move out of the area impacted by the presence of fuel hydrocarbons and the area returns to a more oxidizing environment, the inorganics should become less soluble and be removed from solution. The increased solubility and transport of inorganics observed in the reducing environment immediately downgradient from the organic contamination should be a temporary geochemical imbalance.

As discussed in Section 4.4, the extent of the fuel hydrocarbon and CAH contamination is expected to gradually decrease due to natural attenuation processes. The downgradient transport of inorganics via groundwater should also be a transient process. The extent of downgradient transport via this mechanism is expected to be limited, and is not expected to impact offsite receptors. Although the effects of offsite migration of dissolved COPCs is expected to stabilize, maximum and/or 95-percent UCL concentrations for groundwater based on current sampling results (June/July 1997) are used in this risk evaluation to assess potential risks to future receptors. A qualitative evaluation of the transport, transformation, and fate of COPCs in groundwater is provided in Section 4.4.

5.2.1.3 Potentially Exposed Populations and Exposure Routes

Once the source and release mechanisms have been determined, the routes of exposure and potential receptors need to be identified in order to determine the likelihood and extent of human or ecological receptor contact with site-related contaminants (USEPA, 1989 and 1992a). Land use assumptions are critical to defining the types of receptors that are now present or may be reasonably expected to be present in the foreseeable future at the HWSA, or in immediately adjacent areas that could potentially be impacted by site-related contamination in soils or groundwater.

5.2.1.3.1 Current Onsite Receptors

Groundskeeper

If no construction activities occur at the site, onsite groundskeepers (e.g., landscape maintenance workers) are the most likely human receptors. As previously discussed, the HWSA (except for the former tank area in the southern corner of the site) is surrounded by a chain-link fence with a locking gate resulting in little or no activity within the fence line. Because of this, incidental contact with environmental media (i.e., surface soil) by onsite workers is minimal. However, groundskeepers could potentially be exposed to COPCs via incidental ingestion of and dermal contact with contaminated soil. If it is assumed that the groundskeeper activity disturbs the surface of the soil, inhalation of fugitive dust is also possible within the HWSA, even though the majority of the site is covered with vegetation. Because groundwater is 8 to 12 feet bgs and there is no evidence of seepage near the surface, groundwater exposure pathways are considered incomplete for the groundskeeper.

Hypothetical Worker

Although there are no workers that perform daily activities within or directly adjacent to the HWSA, it was conservatively assumed that such a worker could potentially exist at the site. Thus, it is assumed that this hypothetical worker would be exposed to site soils 8 hours a day, 250 days per year, for 25 years, as directed by Ohio EPA guidance for an industrial scenario (Ohio EPA, 1993b). It is further assumed that these workers potentially could be exposed to COPCs via incidental ingestion of and dermal contact with contaminated soil and inhalation of fugitive dust.

Construction/Utility Worker

Because it is unlikely that construction/excavation activities will occur at the HWSA in the immediate future, potential risks to construction workers are considered as a future scenario in this analysis. Although it is conceivable that a utility worker may occasionally need to excavate a utility line within the HWSA, it was assumed in this assessment that the risks associated with a future construction worker would exceed

risks due to intermittent exposure of a utility worker. It is also unlikely that the same utility worker would work on a given site such as the HWSA more than once.

Trespasser

Trespassing by potential residential or recreational receptors is not a significant concern at the HWSA due to institutional and physical access restrictions (a gate house for entry onto the Base, and a chain-link fence with locking gate around most of the HWSA). There also are no environmental or man-made features that would attract unauthorized personnel to this area. Thus, trespassers are not considered potential receptors at the site.

Groundwater User

No shallow groundwater is withdrawn for use by Rickenbacker ANGB or Rickenbacker Port Authority facilities to meet potable or nonpotable water requirements (see Section 1.3.5). Because of this, onsite groundwater consumption is considered an incomplete exposure route.

Ecological Receptors

As noted earlier, the industrial nature of the HWSA, which includes a paved driveway, a decontaminated building with a concrete floor (Building 560) and a chain-link fence to limit access, precludes the existence of significant wildlife habitat. The areas surrounding the HWSA also are industrial or commercial in nature, with runways, hangar facilities, parking lots, and a number of office buildings. Although it is possible that certain rodents or birds may enter or be near the HWSA, it can be assumed that no special concern species (i.e., threatened, endangered or rare species) frequent the site. Typically, species associated with disturbed/industrial sites are either introduced or commonly found species. During the site visit, no resident ecological receptors were identified for which soils and/or groundwater are likely contaminant exposure media. Given this information, terrestrial biota are not considered receptors at the site.

The vegetation within the HWSA is typical of a disturbed/landscaped site, consisting primarily of horticultural grasses and a variety of weeds. Although this vegetation may be exposed to COPCs in soil domesticated plants are not considered to be ecological receptors (USEPA, 1996b).

5.2.1.3.2 Future Onsite Receptors

Construction Worker - Taxiway

Based on future land use plans, it is likely that some form of construction will occur at the HWSA in the future. The most likely scenario is that the HWSA and the former tank area will be covered by a taxiway and paved support aprons. Associated airfield support facilities will be located in areas adjacent to the site. Building 560 and the surrounding area are included in Parcel D1 (Phase I) of the redevelopment plan (Section 2.3.3). Supporting documentation on the future land use of Parcel D1 based on the approved redevelopment plan is provided in Appendix A.

The construction of a taxiway could represent a short-term exposure to site contaminants for construction workers (twenty 8-hour days) because of the limited extent of intrusive work required at any one area to be graded and paved. Potential exposure routes include incidental ingestion of and dermal contact with soil, inhalation of fugitive dusts and dermal contact with groundwater. Because shallow groundwater typically is encountered at 8 to 12 feet bgs (Section 2.2.2), workers engaged in deep excavations could hypothetically come into contact with groundwater and saturated soils. Construction of a taxiway would require relatively shallow excavation (less than 2 feet); therefore the potential for dermal contact with groundwater is extremely remote. Nonetheless, it was conservatively assumed that direct dermal contact with groundwater could possibly occur during taxiway construction. Incidental ingestion of groundwater is considered an incomplete pathway because it is even more improbable than dermal contact.

Construction Worker - Building or Hangar

The construction of a building or hangar would result in a slightly longer-term exposure for construction workers (90 8-hour days) to site contaminants because of their potential for intrusive work within the HWSA. Potential exposure routes include incidental ingestion of and dermal contact with soil, inhalation of fugitive dusts, and dermal contact with groundwater. If a building or hangar is constructed onsite, deeper excavation may occur than that assumed for the taxiway scenario. However, substantial worker contact with site groundwater is still unlikely because basements and deep building foundations are not typically built in the area because of the relatively high water table (Friedstrom, 1997). Nevertheless, dermal contact with groundwater is conservatively assumed in the analysis. Incidental ingestion of groundwater is considered an incomplete pathway in this scenario.

Utility Worker

As discussed in Section 5.2.1.3.1, it is assumed that a utility worker scenario would be similar to a construction worker scenario but less frequent and shorter exposures would occur. Because of this, a future utility worker scenario was not included in this analysis.

Groundskeeper

It can also be concluded that there would be little or no future exposure of groundskeepers in the area following construction of a taxiway, hangar or building. If part of the HWSA remains uncovered following the construction of the taxiway, there may be limited exposure of a groundskeeper cutting grass along the edge of the taxiway. However, the potential for exposure would be less than or equal to the current exposure scenario proposed for a groundskeeper. Because of this, the keeper is considered only as a current onsite scenario.

Ecological Receptors

Future ecological receptors are considered insignificant or nonexistent, particularly if a taxiway, hangar, or building is built over the HWSA. The rationale for concluding that ecological pathways are incomplete is similar to the rationale described in Section 5.2.1.3.1 for current ecological pathways.

5.2.1.3.3 Current and Future Offsite Receptors

The use of groundwater from the UWBZ is the only conceivable offsite receptor exposure scenario associated with COPCs originating from the HWSA. As discussed in Section 1.3.5, very few residents use private wells screened in the UWBZ in the area surrounding Rickenbacker ANGB and RPA facilities. The nearest offsite well is 1.5 miles southwest of the HWSA. Considering the relative stability of the contaminants in groundwater (Sections 4.2 and 4.4) and the substantial distance from the site to offsite wells, there is no reason to assume that COPCs associated with the HWSA could potentially affect the water quality in these offsite well locations. Therefore, offsite receptor exposures to COPCs in groundwater associated with the HWSA are considered unlikely. Nonetheless, Rickenbacker ANGB will propose institutional controls for the HWSA and other groundwater contaminated sties at the Base, through deed restrictions, to ensure shallow groundwater is not extracted for potable or nonpotable use (Friedstrom, 1997).

5.2.1.4 Summary of Potentially Completed Exposure Pathways

Incidental exposure to onsite COPCs in surface soil and associated fugitive dust are considered completed pathways for the nonintrusive groundskeeper and the hypothetical onsite worker. Incidental exposure pathways for COPCs in groundwater to these receptors are considered incomplete because of the typical depth to groundwater (8 to 12 feet bgs).

For intrusive construction workers, exposure to COPCs may occur through incidental ingestion of and dermal contact with surface and subsurface soils, inhalation of fugitive dust, and dermal contact with COPCs in groundwater. Although exposure

to shallow groundwater is unlikely unless deep excavation activities occur at the site, it was conservatively assumed that the proposed construction of a taxiway or the construction of an airfield support structure (building or hangar) could result in dermal contact with groundwater. The ingestion of groundwater during construction activities is highly unlikely, therefore, this potential pathway is considered incomplete in this analysis.

Because of the nature of the site (an industrial/disturbed site), significant use of the site by ecological receptors appear to be limited or nonexistent. Therefore, ecological pathways are considered incomplete in this analysis.

5.2.2 Estimating Exposure-Point Concentrations

The representative exposure-point concentration is defined as the concentration that represents the highest exposure that could reasonably be expected to occur for a given exposure pathway. This value is intended to account for both the uncertainty in environmental data and the variability in exposure parameters (USEPA, 1992a). Given the substantial amount of historical data for COPCs in soil and the format of the historical database, 95-percent UCLs were not calculated. Therefore, maximum concentrations of COPCs detected in soils were used as conservative estimates of COPC exposure-point concentrations for each exposure pathway considered in the risk assessment.

Fugitive dust in outdoor air may contain SVOCs and metals found in the soil at the HWSA. Although the potential for fugitive dust emissions is low under relatively nondisruptive conditions (i.e., activities performed by groundskeepers and hypothetical onsite workers), intrusive construction workers could generate fugitive dust during deep excavation activities. Concentrations of fugitive dust were developed by calculating a particulate emission factor (PEF) based on USEPA (1996c) guidance for the intrusive and the nonintrusive onsite workers. The use of a PEF allows development of estimated concentrations to which receptors could be exposed via inhalation of contaminated surface dust. COPCs also may volatilize from soil (USEPA, 1996c).

The ambient air concentrations of COPCs volatilized from soil were estimated per USEPA (1996c) guidance using a volatilization factor (VF) term.

To more accurately simulate intrusive construction conditions, it was assumed that there is no vegetative cover present at the site. Due to the potentially invasive nature of construction activities, it also was assumed that intrusive construction workers could be exposed to fugitive dust from subsurface soil as well as from surface soils. Thus, the COPC concentrations in fugitive dust to which the construction worker could be exposed via inhalation was assumed to originate from representative COPC soil concentrations (i.e., maximum detected concentrations) from 0 to 10 feet bgs at the HWSA.

5.3 QUANTIFICATION OF EXPOSURE: CHEMICAL INTAKES

Once the exposure pathways are described qualitatively and the exposure concentrations are defined quantitatively, the amount of any one chemical to which a receptor may be exposed during a specified time is estimated. Calculating chemical intakes hinges on reasonable, yet conservative, assumptions about how each group of potential receptors at a particular site may be exposed to site-related contamination. This step in the risk assessment process is called quantification of exposure. The risks to potential receptors exposed to site-related contamination are then estimated by coupling toxicity data with quantified exposure data.

Intake estimates are normally expressed as the amount of chemical at the exchange boundary in milligrams of contaminant per kilogram of body weight per day (mg/kg-day), which represents an intake normalized for body weight over time. The total exposure is then divided by the time period of interest to obtain an average exposure over time. The time used to average exposure is a function of the toxicity endpoint: for noncarcinogenic effects it is the exposure duration, and for carcinogenic effects it is a lifetime (70 years).

This risk assessment is based on chronic exposures to measured COPC concentrations. Short-term (i.e., subchronic) and acute exposures are not evaluated in this risk assessment. Although intrusive construction workers are assumed to have subchronic exposures, risks are evaluated using chronic toxicity values to be more conservative (health-protective). As required by USEPA (1992c) and Ohio EPA (1993b), an RME exposure scenario is used in this risk assessment. When coupled with the appropriate toxicity information, intakes calculated using the RME exposure parameters result in RME risk estimates.

Where appropriate, standard default intake variables as defined by USEPA (1991c) and Ohio EPA (1993b) were used in quantifying exposure for the intrusive construction worker (e.g., body weight, breathing rate, lifetime). However, some intake variables, particularly for the construction worker, are based on best professional judgment. For example, the taxiway construction worker is conservatively assumed to remain at the job site for an equivalent of 20 8-hour days. This assumption is based on best professional judgment of a field engineer at Rickenbacker ANGB (Haines, 1997). Twenty days is a conservative value because excavation activities likely would take only 3 work days (the other 17 days are related to construction of the concrete pavement - see Appendix G). It is also conservatively assumed that excavation would reach the upper water-bearing zone (8 to 12 feet bgs) at the HWSA even though excavation for a taxiway is typically 2 feet deep.

An exposure scenario with a more detailed construction-related activity (e.g., building a foundation) is included in this analysis for comparison. This scenario assumes that the construction worker would spend up to 3 months performing construction-related activities where exposure to site contaminants in soil and groundwater would be possible. This scenario is also very conservative, as the depth of a foundation or support piers for existing buildings or hangars at Rickenbacker ANGB is typically well above the UWBZ (8 to 12 feet bgs). Finally, workers involved in excavation activities would not remain at the site once the foundation or pavement has been poured and the aboveground construction activities have been initiated.

Calculation of dermal exposure to groundwater also required development of several additional intake variables. An exposed surface area of 3,280 square centimeters (cm²) was used based on the assumption that an adult worker's hands, arms and head would not be covered by clothing, and therefore could be exposed to contaminated groundwater (USEPA, 1997). Consistent with the development of other exposure parameters, the exposure assumption for dermal contact with groundwater was developed to characterize an RME.

If groundwater is encountered during construction, dermal contact by a construction worker is not likely to be continuous over an 8-hour work day. Therefore, it was conservatively assumed that a construction worker would be in contact with groundwater for 50 percent of the 8-hour work day (i.e., exposure time [ET] = 4 hours).

A chemical-specific permeability constant (Kp) value was used to calculate dermal intakes from groundwater. All Kp values were either calculated or obtained from USEPA (1992e) Dermal Exposure Assessment: Principles and Applications. Using these Kp values, a dose absorbed per unit area per event (DA_{event} in mg/cm² per event) was calculated. Detailed methodology, along with the intake variables, resulting exposure factors, and the formulas used to calculate intake for intrusive construction workers and nonintrusive groundskeepers and hypothetical onsite workers are shown in Appendix F.

Dermal absorption factors for soil were developed based on studies of absorption from soil (USEPA, 1992e; California Environmental Protection Agency, 1994). Ten percent of most organics, and one percent of inorganics were assumed to be absorbed dermally, unless chemical-specific absorption factors were available.

A groundskeeper was conservatively assumed to work at the HWSA (e.g., cutting the grass and weeds) six times per year for 5 years. It is also assumed that the task would take up to 8 hours to complete, even though the HWSA is a small site. These assumptions are based on professional judgment and are conservative particularly,

because it is unlikely that the same individual would perform the activity each time, for 8 hours, for 5 years.

It is conservatively assumed that the hypothetical onsite worker would remain at the site 8 hours per day, 250 days per year for 25 years. Although these exposure durations are very unlikely given the nature of the HWSA site, this scenario reflects the standard assumptions for an industrial worker scenario (Ohio EPA, 1993b).

5.4 TOXICITY ASSESSMENT

The objective of the toxicity assessment is to weigh available evidence regarding the potential for particular contaminants to cause adverse effects in exposed individuals and to provide, where possible, an estimate of the relationship between the extent of exposure to a contaminant and the increased likelihood and/or severity of adverse effects. For humans, USEPA has conducted numerous toxicity assessments that have undergone extensive review within the scientific community.

The types of USEPA toxicity values used in this risk assessment include oral reference doses (RfDs), inhalation reference concentrations (RfCs), oral carcinogenic slope factors (CSFs), and inhalation unit risk factors (IURs). RfDs and RfCs are used to evaluate systemic (noncancer) effects. CSFs and IURs are used to evaluate carcinogenic effects. Toxicity values for the noncarcinogens and carcinogens evaluated in this risk assessment are presented in Appendix F. The toxicity information used in this risk assessment was obtained from the Integrated Risk Information Service (IRIS) (USEPA, 1999). If values were not available from IRIS, *Health Effects Assessment Summary Tables* (HEAST) values (USEPA, 1995) were used. In some cases, toxicity values have either been withdrawn or are provisional. For example, there is currently no RfD or CSF for TCE. Because of the potential importance of TCE, provisional values were used in this assessment.

USEPA has not derived toxicity values for all routes of exposure. Most of the available toxicity values are for oral exposure, although many inhalation values are

available. No values are currently available for dermal exposure. Dermal toxicity values were developed by modifying oral toxicity values with chemical-specific oral absorption factors (Appendix F). For those chemicals for which toxicity values are not available for any route of exposure, it may be appropriate to use toxicity values derived for similar chemicals (i.e., surrogates). This is appropriate for chemicals where the toxicity values for one chemical may be used for another related chemical (e.g., polynuclear aromatic hydrocarbons). For most chemicals, however, there is no chemical that is similar enough to justify the use of toxicity information for a surrogate; therefore, these chemicals cannot be quantitatively evaluated in the risk assessment.

5.5 RISK CHARACTERIZATION

To characterize risk, toxicity and exposure assessments were summarized and integrated into quantitative expressions of risk. To characterize potential noncancer effects, comparisons were made between projected intakes of chemicals and chronic toxicity values. To characterize potential carcinogenic effects, probabilities that an individual would develop cancer over a lifetime of exposure were estimated from projected intakes and chemical-specific dose-response information. Major assumptions, scientific judgments, and, to the extent possible, estimates of the uncertainties embodied in the assessment were also considered in the analysis.

The current and future land uses at the HWSA were assumed to involve predominantly nonintrusive maintenance and intrusive construction activities, respectively. The RME chemical-specific hazard quotients (HQs), total HIs, and cancer risk estimates for the aforementioned exposure routes are presented in Appendix F. Table 5.5 summarizes the RME risk estimates developed for current and future receptors that could be exposed to existing levels of site contamination at the HWSA.

5.5.1 Risk Estimates for Hypothetical Current Receptors

Ohio EPA (1995b) states that a cumulative HI greater than 1 or a cumulative carcinogenic risk greater than 10⁻⁵ is unacceptable and necessitates remediation and/or appropriate control measures to protect human health. The Ohio EPA (1995b)

SUMMARY OF QUANTITATIVE RISK ESTIMATES FOR CURRENT AND FUTURE RECEPTORS HAZARDOUS WASTE STORAGE AREA TABLE 5.5

RICKENBACKER ANGB, OHIO

'ay - RME" HI ^b /	Risk Level	p	Risk Level ^c
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Construction Worker (Intrusive)	Taxiway	way	Ha	Hangar
Incidental ingestion of soil	1.42E-02	5.06E-08	6.67E-02	2.28E-07
Dermal contact with soil	7.60E-03	5.91E-08	3.42E-02	2.66E-07
Inhalation of fugitive dusts	2.29E-02	1.18E-08	1.03E-01	5.33E-08
Dermal contact with groundwater	1.43E-01	1.19E-06	6.44E-01	5.36E-06
Inhalation of volatiles from soil	-	-		1
TOTAL	TOTAL 1.88E-01	1.31E-06	8.48E-01	5.91E-06

Groundskeeper (Nonintrusive)		
Incidental ingestion of soil	4.45E-03	7.59E-08
Dermal contact with soil	1.14E-03	4.43E-08
Inhalation of fugitive dusts	6.86E-03	1.78E-08
Inhalation of volatiles from soil	-	
TOTAL	TOTAL 1.24E-02	1.38E-07

Hypothetical Worker (Nonintrusive)		
Incidental ingestion of soil	1.85E-01	1.58E-05
Dermal contact with soil	4.75E-02	9.24E-06
Inhalation of fugitive dusts	2.86E-01	3.70E-06
Dermal contact with groundwater		
Inhalation of volatiles from soil		
TOTAL	TOTAL 5.18E-01	2.87E-05

a' RME = reasonable maximum exposure.

 $^{^{}bV}$ HQ = chemical-specific hazard quotients; summed across all pathways for each receptor group to obtain cumulative HI (target is cumulative < I).

^e Carcinogenic risk level; expressed as probability; summed across all pathways for each receptor group to obtain cumulative risk level (target is cumulative 10⁻⁵ for potential exposure).

memorandum goes on to state that "when site-specific risk assessments are conducted for an industrial-use scenario, a risk goal of 10⁻⁴ may be used provided a demonstration is made that the cumulative risk from potential offsite exposure on neighboring property is less than 10⁻⁵". In this assessment, the risk target goal is 10⁻⁵; however, given that the HWSA is clearly an industrial site, risks that slightly exceed this goal are considered acceptable, particularly if the exposure scenario and assumed exposure pathways are unlikely.

The cumulative RME HI for a groundskeeper is 0.01, well below the threshold value of 1. The cumulative cancer risk estimate for this receptor is 1.4×10^{-7} , which is well below the Ohio EPA target of 1×10^{-5} .

The cumulative RME HI for a hypothetical onsite worker is 0.5, which is below the threshold value of 1. The cumulative cancer risk estimate is 2.8×10^{-5} . Because the potential for actual exposure for this receptor is believed to be low or nonexistent at the HWSA, and because of the hypothetical nature of this scenario, the cumulative cancer risk is believed to be acceptable even though it exceeds the target risk level (1×10^{-5}) . In addition, it is below the 1×10^{-4} risk level for industrial sites discussed above. As stated earlier, there is little reason to assume that a worker would be working at this small site, 8 hours per day, 250 days per year, for 25 years.

5.5.2 Risk Estimates for Hypothetical Future Receptors

The cumulative RME HI for a taxiway construction worker is 0.2, which is below the threshold value of 1. The cumulative cancer risk estimate for the taxiway construction worker is 1.3×10^{-6} . The cumulative RME cancer risk total is below the Ohio EPA target of 1×10^{-5} , indicating an acceptable risk level. As suggested earlier, the construction of a taxiway is the most likely future scenario for the HWSA.

If a more conservative and intrusive construction scenario is considered (i.e., construction of a hangar), the cumulative HI is 0.8, which is below the threshold value of 1. The cumulative cancer risk estimate for the hangar construction worker is 5.9 x

10-6, which is below the Ohio EPA target of 1 x 10-5, indicating an acceptable risk level.

5.5.3 Summary/Discussion of Results

The results of this risk assessment demonstrate that measured concentrations of COPCs in soils and groundwater at the HWSA do not result in HQs greater than the noncancer threshold limit of 1 for any of the receptors evaluated. While cumulative risk to the nonintrusive hypothetical onsite worker (2.6 x 10⁻⁵) slightly exceeds the cancer target risk level of 10⁻⁵, the exposure assumptions used in this scenario also are extremely conservative and unlikely. In addition, the estimated cumulative risk to the nonintrusive hypothetical onsite worker (2.6 x 10⁻⁵) is well below the 10⁻⁴ risk level that is considered acceptable for industrial sites.

5.6 UNCERTAINTY CHARACTERIZATION AND ANALYSIS

All risk assessments involve the use of assumptions, judgments, and imperfect data to varying degrees. This results in uncertainty in the final estimates of hazard and risk. This section describes the likelihood that the approaches incorporated into this risk assessment either overestimate or underestimate the actual risks associated with exposure to site-related organic chemical concentrations, pursuant to Ohio EPA (1993a and 1993b) guidance. Risk assessment in general, as it is currently practiced, is highly conservative and often based on extremely conservative assumptions and scenarios.

There are several types of uncertainty associated with the risk assessment process. The first type concerns the initial selection of substances for analyses and the uncertainties associated with data quality. Another area of uncertainty involves the selection of exposure scenarios that are conservative (i.e., protective of human health), and yet are realistic and probable. Additional uncertainties are inherent in the exposure assessment for individual substances and individual exposures. Those uncertainties are driven by the degree of reliability of the chemical monitoring data, the models used to estimate exposure-point concentrations in the absence of monitoring data, and the receptor intake parameters (e.g., exposure factors). The availability of toxicity

information for the COPCs at the site to address all routes of potential exposure (e.g., dermal exposure) is another area of uncertainty. Finally, additional uncertainties are incorporated into the risk assessment process when exposures to several substances are summed to develop cumulative risks.

5.6.1 Data Uncertainties

Inorganic chemicals were compared to background levels and eliminated if site concentrations did not exceed background levels. It is possible that some of the chemicals not retained for risk analysis may be present as a result of anthropogenic activities. It is possible, although unlikely, that the elimination of the anthropogenic fraction of these inorganic chemicals present at these sites could lead to an underestimation of risk. Chemicals were not eliminated from the quantitative risk assessment based on any detection-frequency analysis, and this approach possibly resulted in an overestimation of risks (e.g., SVOCs in soil).

To calculate the 95-percent UCL, chemicals detected in at least one sample were assumed to be present in all nondetect samples. Statistical analyses (mean, standard deviation, 95-percent UCL, etc.) were then performed on all data by evaluating detects in combination with nondetects. In some cases, especially for VOCs and SVOCs in groundwater, maximum values were used instead of 95-percent UCL values if a nonparametric UCL could not be accurately determined. Especially for these types of scenarios, use of the 95-percent UCL or the maximum measured concentration as the concentration term may overestimate the overall amount of chemical present in the exposure medium and, consequently, the risk posed by the chemical.

5.6.2 Exposure Uncertainties

A significant part of a risk assessment is the estimation of risks that are based on receptor exposure; if exposure of receptors does not occur, no risks are present. Although this assessment qualitatively identifies the probability of the exposure pathway occurring, the quantitative risk estimates for those receptor groups where exposure is possible but unlikely will be overestimated. Additionally, in the risk assessment, it is

assumed that each unique receptor is exposed to the same contaminant concentrations. This assumption tends to overestimate risk because each individual receptor will not realistically be exposed to precisely the same contaminant concentrations.

5.6.3 Uncertainty in Exposure-Point Concentration and Intake Values

There is also uncertainty associated with estimating exposure-point concentrations and the matrix-specific intake factors, including uncertainty associated with intake values and their respective default values for the RME exposure scenarios. Uncertainty arises in the assumption that current and future nonintrusive receptors will be exposed to a mixed (surface plus subsurface) soils stratum. By not evaluating the soil strata on a discrete basis, and assuming all strata are combined, underestimation or overestimation of risks may occur. Assuming that a receptor will be exposed to the maximum concentration detected, regardless of the soil strata from which the detection occurred, could result in an overestimation of risk. In groundwater and soil, exposure-point concentrations are assumed to remain constant over time, while all organic compounds are subject to concentration decreases due to destructive and non-destructive attenuation mechanisms (Section 4.4).

The models used to estimate exposure via a particular route can introduce a considerable amount of uncertainty. In particular the calculations of dose for dermal exposure to groundwater or soil introduce relatively large amounts of uncertainty. The exposure estimates incorporate several factors for which there is little chemical-specific experimental information available (i.e., permeation rates through skin, bioavailability of chemicals bound in soil, or length of time soil remains in contact with skin). Without specific information, it is necessary to make assumptions about the behavior of chemicals in soils. Uncertainty will be introduced into the risk assessment to the degree that the assumptions used in these models are inaccurate. In general assumptions are made which will tend to make the model results conservative (health protective). For example the non-steady state model recommended by USEPA (1992e) in *Dermal Exposure Assessment: Principles and Applications* was used to calculate the dose of organic contaminant absorbed dermally from groundwater. USEPA states that

this approach provides a more conservative total absorbed dose over the traditional steady-state equation.

Standard assumptions regarding body weight, duration of exposure, life expectancy, receptor population characteristics, and lifestyle were made to reflect the RME exposure to individuals for each pathway evaluated. The assumption for RME reflects a conservative (health-protective) approach. Because of this, the RME most likely will overestimate actual risk.

Exposure variables for this risk assessment were taken from current USEPA and Ohio EPA guidance, when available. However, some exposure scenario intake values are location-specific and reflect professional judgment. For example, the duration and frequency of exposure to soils via several exposure routes for the groundskeeper and the future construction worker are unknown, and professional judgment was used. Judgment was applied with the intention of overestimating, rather than underestimating the upper-bound risk estimate.

5.6.4 Uncertainty in the Toxicity Assessment

Some uncertainty is inherent in the toxicity values used for the assumed duration of exposure assessed. These uncertainties are compounded under the assumption of dose additivity for multiple substance exposure. That assumption ignores possible synergisms or antagonisms among chemicals, and assumes similarity in mechanisms of action and metabolism. Another assumption is that all the toxicity values used have an equal degree of reliability, which in reality is not the case. Overall, these assumptions would tend to overestimate hazards and risks. Because toxicity constants for cancer generally are based on the 95-percent UCL, risks tend to be overestimated.

Not all organic chemicals detected can be evaluated quantitatively for health effects because toxicity values do not exist for all chemicals. The lack of toxicity data tends to underestimate risk, therefore, the more chemicals that lack toxicity data, the greater the tendency is for risk underestimation. To help avoid this potential underestimation,

provisional or withdrawn toxicity values were used in the assessment particularly for identified COPCs (e.g., TCE).

There is uncertainty associated with the use of dermal absorption factors used in the conversion of oral toxicity values to dermal toxicity values. Use of dermal toxicity values based on converted oral toxicity values introduces uncertainty due to potential differences in point-of-entry effects and potential differences in metabolic activation/deactivation in dermally absorbed doses. This could tend to either overestimate or underestimate risk.

Whenever multiple absorption factors exist for a particular chemical, the most stringent factor for the various chemical forms was selected. This also may lead to an overestimation of risk.

Regarding noncancer health effects, the application of uncertainty factors to noobserved-adverse-effect level (NOAEL) for a chemical in an animal study for animalto-human extrapolation adds additional uncertainty to the toxicity assessment. The application of scaling or uncertainty factors may result in an overestimation of risk.

5.6.5 Uncertainty in Risk Characterization

Uncertainties in the risk characterization process reflect the cumulative effects of uncertainties in all preceding risk analysis steps. Overall, the assumptions in this risk assessment tend to overestimate risk.

5.7 SITE-SPECIFIC TARGET LEVELS

As discussed in Section 5.5.3, the results of this risk assessment demonstrate that measured concentrations of COPCs in soils and groundwater at the HWSA do not result in HQs greater than the noncancer threshold limit of 1 for any of the receptors evaluated. In addition, the HWSA is clearly an industrial site and cumulative cancer risk levels for all receptors were less than or only slightly exceeded the Ohio EPA (1993b) target risk level of 1 x 10-5 and were well below the Ohio EPA (1993b) 1 x 10-4 risk level that is considered acceptable for industrial sites. Based on the results of this

risk assessment, measured concentrations of COPCs in soils and groundwater at the HWSA would not be expected to result in unacceptable cancer risks and/or noncancer hazards. Therefore, site-specific target levels were not calculated.

SECTION 6

IMPLEMENTATION OF PROPOSED CLOSURE ACTIVITIES

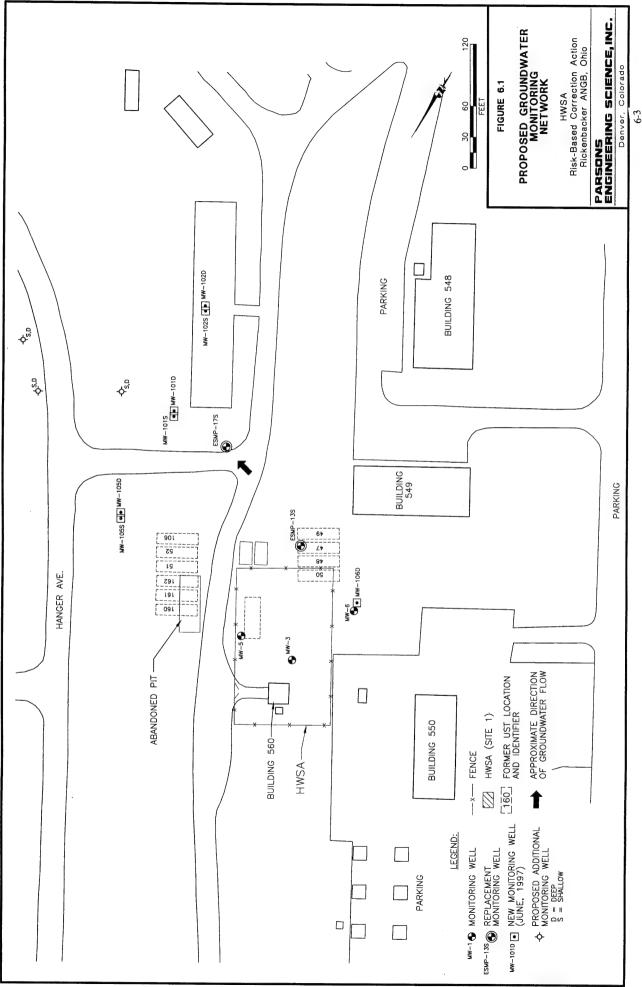
This section describes the actions proposed to be implemented at the HWSA to facilitate closure of the unit. Additional soil, soil gas, and groundwater sampling at the HWSA has been performed to supplement site data from earlier investigations in determining the nature and extent of contamination at the site, and for use in a site-specific risk assessment based on current and future land use scenarios. The risk assessment demonstrated that concentrations of organic and inorganic compounds in soil and groundwater do not pose unacceptable carcinogenic risks or noncancer hazards to current or future receptors assuming continued commercial/industrial land use. Considering these findings a closure approach based on annual groundwater monitoring and institution of land-use controls is proposed.

6.1 PROPOSED CLOSURE APPROACH

To implement closure of both soils and groundwater, the following activities have been completed or are proposed to be completed:

- Removal of the remaining four USTs (completed February 1995);
- Decontamination of Building 560 by cleaning the building and the drum wash pad (completed April 1996);
- Installation of 10 new monitoring wells in the vicinity of the HWSA and comprehensive groundwater sampling from 37 site monitoring wells (completed July 1997);
- Sampling of soil and soil gas in the vicinity of Building 560 (completed August 1997);

- Treatability testing for determination of the effectiveness of air injection bioventing, in the event residual organic soil contamination required limited remediation based on the findings of the risk assessment (completed August 1997);
- Treatability testing for determination of the effectiveness of air sparging near ESMP-17S in the event groundwater in the vicinity of the site required in situ remediation based on the findings of the risk assessment (completed December 1997);
- Performance of a comprehensive risk assessment (Section 5) to quantitatively determine the carcinogenic risks and noncancer hazards associated with soil and groundwater contamination at the site;
- Natural oxidation of dissolved BTEX and natural reductive dehalogenation of dissolved chlorinated VOCs (in progress);
- Installation of one additional monitoring well cluster downgradient of MW-101S and MW-101D in order to completely delineate the extent of the chlorinated VOC plume (proposed);
- Installation of two monitoring well clusters east of Hangar Avenue in order to evaluate groundwater quality and plume stability at the property boundary (proposed);
- Continued monitoring of the groundwater COPCs identified in the risk assessment (proposed). Based on the findings presented in Section 4 and Section 5, a revised monitoring well network is shown on Figure 6.1 and discussed in Sections 6.3.1 and 6.3.2;



- Institution of land-use controls through a deed restriction upon transfer of the HWSA property (proposed); and
- Eventual exposure control by installation of taxiway (proposed as future land use).

6.2 ACCESS CONTROL

During closure activities, the source area near Building 560 will be secured with a chain-link fence, with padlocked gates to limit access in compliance with OAC 3745-68-9. The fence will be removed prior to construction of the taxiway or implementation of contingency actions, if necessary. Limiting access to the site will be unnecessary following construction of the taxiway.

6.3 CLOSURE FIELD ACTIVITIES

Many of the activities to be completed as a part of the closure will be field-related activities. These activities are described in detail below.

Because of known contamination in groundwater, an alternate groundwater monitoring system, in accordance with OAC 3745-65-90 (D), has been implemented at the site. The monitoring well network as detailed in the February 1997 amended closure/post-closure plan (Parsons ES, 1997a) relies on quarterly sampling from 19 monitoring wells installed prior to 1997. While this network complies with OAC 3745-65-91 (A, B, C) and 3745-65-92 (A, E), it does not incorporate the findings of the 1997 assessment activities (Section 4) or the site-specific risk assessment (Section 5) as presented in this plan. The proposed monitoring network and sampling and analysis procedures as outlined below comply with the above-cited OAC rules, and consider findings that have been previously reported. The following section provide details for the proposed expanded monitoring well network.

6.3.1 Monitoring Well Installation

Eighteen monitoring wells or points will be used to conduct annual monitoring at the site (Figure 6.1). Twelve of these sampling locations have been used previously to define the nature and extent of contamination at the site. However, six additional sampling points are required to determine the downgradient extent of chlorinated VOC contamination in groundwater. Also, the two 0.5-inch ID monitoring points (ESMP-13S and -17S) in the proposed monitoring network should be abandoned and replaced with 2-inch monitoring wells screened at the same depths as the existing monitoring points.

Besides the replacement of ESMP-13S and -17S, six additional wells (three shallow and three deep) should be installed and sampled as part of the proposed long-term monitoring activities. Two wells should be installed approximately 60 feet downgradient from monitoring wells MW-101S and D, and used with other site monitoring wells to assess the extent of dissolved contamination downgradient from the HWSA. Four wells should be installed immediately east of Hangar Avenue in order to monitor plume stability and groundwater quality at the property boundary. Based on the sand zones encountered during the drilling of MW-101S and D, the new shallow wells should be drilled to approximately 13.5 and the new deep wells should be drilled to approximately 21.5 feet bgs. The proposed wells should be completed as described in OAC 3745-65-91 (C). These wells and the other 12 wells shown on Figure 6.1 will be sampled using methods described in Appendix G. Groundwater samples collected from these wells will be used to confirm plume stability and contaminant reduction progress via natural attenuation.

6.3.2 Groundwater Monitoring, Sampling and Analysis

Following the installation of the three well pairs downgradient from MW-101S and D, the 18 wells comprising the monitoring well network to be used during closure monitoring (Figure 6.1) will be sampled annually to confirm stability of the CAH plume, and to ensure downgradient migration does not extend beyond the Rickenbacker

ANGB boundary (Hangar Avenue). Results of these groundwater monitoring activities will be submitted to Ohio EPA in an annual report as required by OAC 3745-65-93. If groundwater monitoring results demonstrate that the plume is stable or diminishing with time, and if consecutive annual monitoring results at the downgradient property boundary demonstrate that contaminant groundwater concentrations are below MCLs, groundwater monitoring will be discontinued.

Groundwater samples will be collected in accordance with methods described in Appendix G. Groundwater samples will be collected for analysis of VOCs by USEPA Method SW8260. Samples will be collected and groundwater elevations will be determined during each annual monitoring event. Based on their relative immobility, low frequency of detection (Section 4.2.2.2 and Appendix E), and low contribution to site risk (Section 5.7), groundwater samples will not be analyzed for SVOCs or metals.

6.3.3 Site Inspection and Maintenance

Applicable RCRA interim status requirements described in OAC 3745-66-13(b)(2) will be maintained during closure activities to prevent threats to human health and the environment. The applicable requirements include the following:

- Maintenance of site security through the upkeep of the existing fence, prevention
 of unauthorized entry, and preventing unnecessary physical contact with or
 disturbance of contaminated media;
- Training site personnel in proper hazardous waste management procedures (see Appendix H);
- Maintaining communication regarding the site with local police, fire, and spill response authorities; and
- Continued groundwater monitoring to confirm plume stability.

During each of the annual groundwater monitoring events, the site will be inspected, and regular maintenance and repair activities will be completed. The site building

(Building 560), fence, gate, and padlocks will be inspected for damage due to use, abuse, wear, vandalism, or weathering. If any damage is noted during the annual inspections, the damage will be immediately repaired or damaged items will be replaced. All damage and repair or maintenance actions will be noted in a field logbook and in the annual report to the Ohio EPA.

For the groundwater monitoring network, the inspection program will consist of checking each well for the following:

- · Damaged protective casings;
- · Damaged well casings;
- Missing or damaged well covers, caps, or locks;
- · Presence of foreign objects in wells;
- · Heaving or subsidence of the wells;
- Damaged concrete pads;
- · Silting of the wells; and
- Other signs of unauthorized use, abuse, vandalism, or deterioration.

If any of the above circumstances are observed, they will be noted in the dedicated field log book. The missing or damaged items will be repaired or replaced within 30 days, as appropriate. A record of any observations and repair/replacement activities will be included in the annual report.

6.3.4 Site Survey

The location and dimensions of the fenced area (HWSA) will be determined by a registered professional surveyor with reference to permanently installed and protected onsite benchmarks. The survey data will be used to prepare and maintain a survey plat of the HWSA, which will be kept at the Base. If it is determined through the deed

restriction and deed transfer process that areas downgradient from the HWSA also will be subject to land use controls, then these areas also will be surveyed and included in the survey plat.

6.4 CONTINGENCY PLAN

Should the proposed closure approach fail to retard dissolved contaminant migration and/or achieve the closure objectives for this site, there should be no significant impact on the land use plans for the site. Risk assessment results detailed in Section 5 demonstrated that there were no unacceptable risks or hazards associated with reasonable industrial land use. Industrial land use has been proposed; the site soon will be covered by a concrete taxiway to be constructed by the RPA.

Contingency actions would only be necessary if potential receptors may be subject to unacceptable exposure and/or if site-related contaminants are observed to be migrating beyond areas intended for industrial land use at concentrations that could pose an unacceptable risk to downgradient receptors. If either of these events occur, an additional assessment event would be conducted to determine the extent of contamination and to locate the zone of highest contamination. If contaminant migration beyond Hangar Avenue occurs, the Air Force should consider engineered barriers such as a sparging curtain or groundwater collection trench. As noted, the failure of the proposed closure approach will not impact the current or proposed uses of the HWSA.

Section 6.1 describes the closure activities deemed appropriate to pursue a risk-based closure of the HWSA. However, if annual groundwater monitoring data suggest that these approaches will not achieve the desired level of risk reduction at the site, possible contingency actions could include installation of a cover prior to taxiway construction and groundwater extraction and treatment. This level of remediation would be deemed excessive by most environmental professionals given the levels of contamination encountered at this site and the current and future uses of the land and groundwater.

6.5 AMENDMENT OF THE CLOSURE PLAN

Whenever changes in the operating plans or facility design affect the closure plan, or there is a change in the expected year of closure, or unexpected events occur during the partial or final closure activities, the closure plan will be modified by the procedures established in OAC 3745-66-12.

6.6 COST ESTIMATES

Rickenbacker Air National Guard Base is exempt from filing a closure cost estimate (40 CFR 265.142 and .143) because it is owned and operated by the federal government. However, preliminary cost estimates have been previously provided to support potential subsequent decision making regarding the need for, and cost of, contingency actions at the HWSA. Based on estimates presented in the February 1997 amended closure/post-closure plan (Parsons ES, 1997a), the estimated cost for risk-based closure of the HWSA, including additional well installation and annual confirmation groundwater sampling for 6 years, is approximately \$200,000. The estimated costs for contingency actions at the HWSA, including design and installation of a cover and installation and maintenance of an onsite groundwater extraction and treatment system exceed \$2,500,000 (Parsons ES, 1997a).

6.7 CLOSURE NOTICES

6.7.1 Annual Groundwater Monitoring System Reports

Groundwater sampling reports should be submitted to the Ohio EPA annually during facility closure activities. These reports will include the following information:

- Results of site inspection and maintenance activities;
- Groundwater elevation data;
- · An evaluation of the groundwater surface elevations;
- Results of groundwater analyses;

- An evaluation of the analytical results;
- A determination of the rate and extent of contaminant migration; and
- · Any other pertinent data or information.

6.7.2 Record of Hazardous Waste Disposal

As stipulated in OAC 3745-66-19, a record of the type, location, and quantity of hazardous wastes disposed of at the site will be submitted to the Ohio EPA and the local zoning authority no later than 60 days after the certification of closure.

6.7.3 Deed Restriction Notation and Certification

A notation will be recorded on the facility property deed in accordance with OAC 3745-66-19. A certification stating that said notation was made, and a copy of the document in which the notation was placed, will be submitted to the director of the Ohio EPA within 60 days of certification of closure.

The Closure Certification and Closure Plan will ensure a risk-based closure under the industrial scenario requiring no further remediation. The AFBCA will cause, through the sale and transfer of the property the filing of a notice and deed restriction to restrict future land use of the property, other than industrial within the Franklin County Recorder's Office. In addition, land-use controls will be instituted through the deed restriction that minimize contact with site groundwater. Specifically, potable and nonpotable water extraction and use from the UWBZ at the site (0 to 40 ft bgs) will be restricted.

Because the deed restriction for the HWSA will not be developed until the time of deed transfer, specific deed restriction language for the site is not submitted with this plan. Example deed restriction language similar to that which will be required for the HWSA is provided in Appendix A. The final and complete deed restriction will be submitted to the Ohio EPA following completion of the last annual monitoring event.

The deed restriction for the HWSA will be developed based on the most recent Ohio EPA guidance.

6.7.4 Survey Plat

In accordance with OAC Rule 3745-66-16, a survey plat of the facility will be submitted to the Ohio EPA and the local zoning authority no later than the submission of the certification of closure. The plat will be prepared and certified by a professional land surveyor. The survey plat will contain a note, prominently displayed, that states the obligation of the owner/operator to restrict disturbance of the facility in accordance with OAC 3745-66-10 to 3745-66-20.

6.7.5 Closure Certification

Rickenbacker ANGB will submit certification of closure that states that closure of the facility was performed in accordance with the specifications in the closure plan. The certification will be submitted, by registered mail, to the Director of the Ohio EPA and to the Regional Administrator of the USEPA within 60 days of the completion of the established closure period. The certification will be signed by an agent of the owner/operator and by an independent, qualified, Ohio-registered professional engineer.

In accordance with Ohio Revised Code, OAC 3745-50-42(D), the signatories to the certification of closure will make the following certification:

"I certify under penalty of law that this document and all attachments were prepared under my direction or supervision in accordance with a system designed to assure that qualified personnel properly gather and evaluate the information submitted. Based on my inquiry of the person or persons who manage the system, or those persons directly responsible for gathering the information, the information submitted is to the best of my knowledge and belief, true, accurate, and complete. I am aware that there are significant penalties for submitting false information, including the possibility of fine and imprisonment for knowing violations.

6.8 CLOSURE CONTACT

The closure contact for the referenced site is:

Mr. Alan Friedstrom AFBCA/DB Rickenbacker IAP 7556 South Perimeter Road Columbus, OH 43217-5910

Telephone: (614) 492-8065 Ext. 13

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SECTION 7

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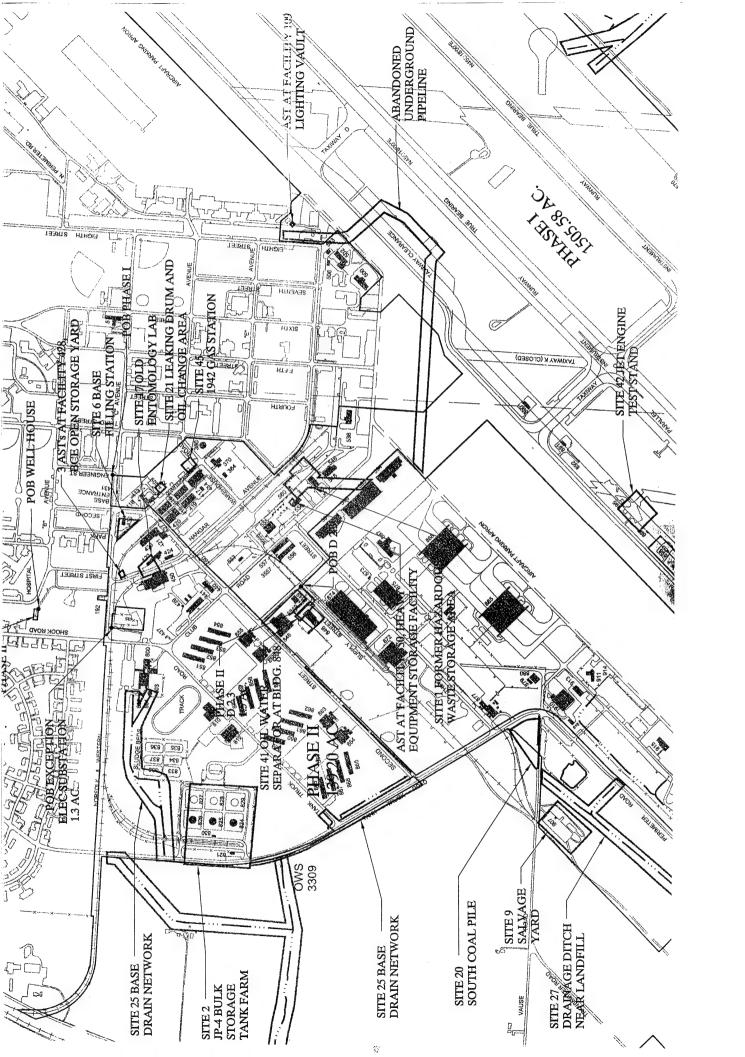
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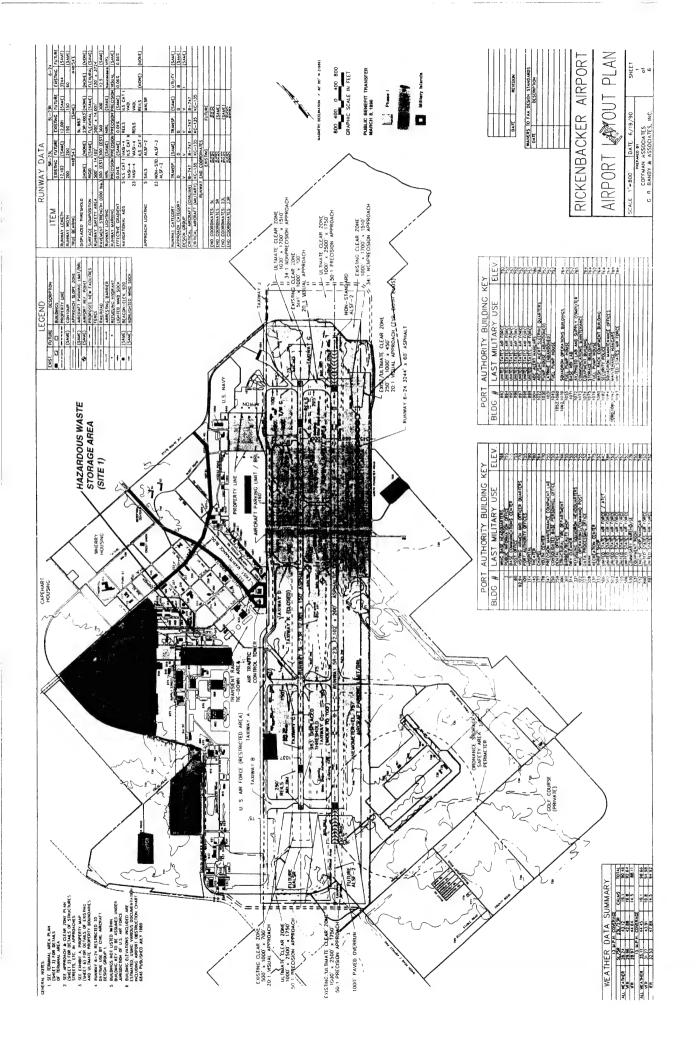
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APPENDIX A AVAILABLE FUTURE LAND USE DOCUMENTATION

APPENDIX A-1

HWSA LOCATION WITHIN PHASE I (PARCEL D1) OF RPA DEVELOPMENT PLAN





APPENDIX A-2

EXAMPLE DEED RESTRICTION LANGUAGE FROM AIR FORCE PLANT 85 (JAMES ROAD), COLUMBUS, OHIO

Tn: 913038318208

JAMES ROAD ENVIRONMENTAL RESTRICTION.

The Grantor, United States Air Force hereby provides public notice about, and imposes restrictions on, the following described real estate commonly known as the James Road parcel which is a sub parcel of the Property, which sub parcel shall hereinafter be referred to as the "Area", more particularly described as follows:

Situated in the State of Ohio, County of Franklin, City of Columbus, in Quarter Township 4, Township 1, Range 17, United States Military Lands and being part of the 62.068 acre tract conveyed to The United States of America by deed of record in Deed Book 1733, page 399, Recorders Office, Franklin County, Ohio, and being more particularly described as follows:

Beginning for reference at a found railroad spike in the centerline of James Road at an inside corner of said 62.068 acre tract, at the northeast corner of a 2.353 acre tract conveyed to Modern Builders Supply, Incorporated, by deed of record in Official Record 21382, Page B11, Recorders Office, Franklin County, Ohio: thence across said 62.068 acre tract on a projection of the centerline of said James Road, North 53°12'32" East, a distance of 322.37 feet to the TRUE POINT OF BEGINNING in the westerly line of the herein described tract of land; thence with said westerly line.

North 34°11'05" West, a distance of 74.70 feet to a set iron pin; thence

North 55°48'55"; East a distance of 230.13 feet to a set iron pin; thence

South 34°11'05" East, a distance of 110.00 feet to a set iron pin about 10 feet southeast of the southeast edge of the old James Road pavement; thence with a line generally parallel to said edge of pavement and about 10 feet distant therefrom.

South 55°48'55" West, a distance of 230.13 feet to a set iron pin; thence North 34°11'05" West, a distance of 35.30 feet to the TRUE POINT OF BEGINNING. passing said edge of pavement at about 10 feet, containing 0.581 acre of land.

- 1. In consideration for the Director of the Ohio EPA's approval of the Hazardous Waste Closure Plan dated 9/24/97, 1997 for the Area, Grantor agreed to impose the restrictions described below on the Area.
- 2. The Grantor hereby imposes the following restrictions on the Area:

The Area shall not be used for residential activities, but may be used for industrial activities. The term "residential activities" shall include the following:

(1) Single and multi-family dwelling and rental units;

- (2) Day care centers and preschools;
- (3) Hotels and motels;
- (4) Educational (except as a part of the facility's industrial activities) and religious facilities;
- (5) Restaurants and other food and beverage services (except as a part of the facility's industrial activities);
- (6) Entertainment and recreational facilities (except as a part of the facility's industrial activities);
- (7) Hospitals and other extended care medical facilities; and
- (8) Transient or other residential facilities.

The term "industrial activities" shall include manufacturing, processing operations and office and warehouse use, including but not limited to production, storage and sales of durable goods and other non-foodchain products and parking/driveway use.

- 3. All of the above restrictions on the Area shall run with the land and be binding upon the Grantor, its successors, assigns, transferees, or any other owner of the Facility. All of the above restrictions shall continue in perpetuity, subject to termination as described below.
- 4. Any person may request written approval for a use of the Area which is not specifically permitted by the restriction set forth above in paragraph 3 by submitting a written petition, via certified mail, to the Director of Ohio EPA. The Director shall respond to such request within sixty (60) days of receipt of such written request. Any such request which constitutes a change in the specific prohibition may only be granted based on the standards described below. In such event, the petition for modification or termination shall state the specific provision(s) sought to be modified or terminated and shall further include evidence demonstrating;
- (a) The Area meets Ohio's risk-based closure performance standard in effect at the time of such petition for a residential (or equivalent) land-use scenario; or
- (b) The Area has been sample tested and compared with background samples taken from land surrounding the Area which has been unaffected by past treatment, storage, or disposal of hazardous waste, and such data shows that current conditions present at the Area are not statistically greater than background conditions; or

PAGE: 007

(c) The Area does not pose an unacceptable risk to human health or the environment.

614-728-3898

- 5. The petition will be considered by the Director of Ohio EPA only when it presents new and relevant information not previously considered by the Agency. The Director of Ohio EPA will issue a final determination within such sixty (60) day period based upon the criteria set forth above.
 - 6. For violation or breach of the foregoing Use Restriction, Ohio EPA shall have the right to proceed at law or in equity to compel compliance with the terms hereof or to obtain injunctive relief in order to prevent violation or breach of the foregoing Use Restriction. Failure to timely enforce the foregoing covenant and Use Restriction by any party shall not bar subsequent enforcement by such party and shall in no manner be deemed a waiver.

APPENDIX B

PREVIOUS SITE ANALYTICAL DATA (1988-1996)

APPENDIX B-1

SUMMARY OF ANALYTICAL DATA, 1988-1990

RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

2	DX/DW	2		Zinc	X	DE LONG TO	2	•		
- (mg/kg	CWND	0.47	Thalllum	₹ :		OUT		A-3	SU-20
-	mg/kg	2	0.66	Silvar	X		OUT	6	A-3	SU-20
0.5	mg/kg	AB	0.2	Solenlum	X		OUT	-		SU-20
*	mg/kg	,	15.2	Nickel	×		OUT	6		SU-20
0.1	mg/kg	BO.	0.082	Morcury	Σ		OUT			SU-20
0.3	mg/kg		22.9	Lead	Σ		OUT		A-3	SU-20
2.5	mg/kg	•	15.3	Copper	Σ		OUT			SU-20
_	mg/kg	7	9.1	Chrombm	Σ		OUT	6		SU-20
0.5	mg/kg	2	0.19	Cadmium	Σ		OUT			SU-20
0.5	mg/ku	7	0.1	Baryllium	×		OUT		A-3	SU-20
-	mg/kg	CNB	13.1	Arsanic	Σ		100	0-2		SU-20
9	mg/kg	25	4	Antimony	Z	•••	007	0-2		SU-20
¥	ug/kg	NA	2	ALL SEMI-VOLATILES	S	SURFACE SOIL SAMPLE	OUT	0-2	A-3	SU-20
								1	1	1
2	mg/kg	3	90.0	Zinc	2	SURFACE SOIL SAMPLE	THO	0-0	A-2	ī
-	mg/kg	CWNU	0.53	Thallium	×	SURFACE SOIL SAMPLE	OUT	1	A-2	SU-19
-	mg/kg	2	0.75	Silvar	Σ	SURFACE SOIL SAMPLE	DOUT	0-2	A-2	SU-19
0.5	mg/kg	2	0.2	Selonlum	Σ	SURFACE SOIL SAMPLE	DOUT	0-2	A-2	SU-19
7	mg/kg		14.1	Nickol	×	SURFACE SOIL SAMPLE	OUT	0-2	A-2	SU-19
0.1	mg/kg	8	660.0	Marcury	2	SURFACE SOIL SAMPLE	OUT	0-2	A-2	SU-19
0.0	mg/kg		110	Load	Σ	SURFACE SOIL SAMPLE	OUT	0-2	A-2	SU-18
2.5	mg/kg	•	15.0	Copper	2	SURFACE SOIL SAMPLE	OUT	0-2	A-2	SU-18
-	mg/kg	3	14.4	Chrombm	Σ	SURFACE SOIL SAMPLE	TUO	0-2	A-2	SU-19
0.5	mg/kg		0.60	Cadmium	Σ	SURFACE SOIL SAMPLE	OUT	0-2	A-2	SU-19
5.0	mg/kg	83	0.45	Boryllium	Σ	SURFACE SOIL SAMPLE	OUT	0-2	A-2	SU-19
•	по/ка	ZVB	9.6	Arsenic	Σ	SURFACE SOIL SAMPLE	OUT	0-2	A-2	SU-19
9	mg/kg	2	4.6	Antimony	X	SURFACE SOIL SAMPLE	DOCT	0-2	A-2	SU-19
330	ug/kg		200	Benzo(g,h,l)Porylana	25	SURFACE SOIL SAMPLE	OUT	0-2	A-2	SU-19
330	ug/ka		440	Indono(1,2,3-cd)Pyrana	S	SURFACE SOIL SAMPLE	OUT	0-2	A-2	SU-19
230	ug/kg		850	Bonzo(a)Pyrene	SV	SURFACE SOIL SAMPLE	OUT	0-2	A-2	SU-19
000	ug/kg		230	Banzo(k) Fluoranthone	SV	SURFACE SOIL SAMPLE	DO	0-2	A-2	SU-19
င္ပင	ug/kg		750	Bonzo(b) Fluoranthono	SV	SURFACE SOIL SAMPLE	DO	0-2	A-2	SU-19
000	ug/kg		07.2	Chrysono	S	SURFACE SOIL SAMPLE	DO	0-2	A-2	50-19
000	ug/kg		740	Bonzo(a) Anthraceno	SV	SURFACE SOIL SAMPLE	OUT	0-2	A-2	SU-19
000	ո <u>ց</u> /kg		1500	Pyreno	S	SURFACE SOIL SAMPLE	OUT	0-2	A-2	SU-19
99	ug/kg		1500	Fluoranthene	SV	SURFACE SOIL SAMPLE	DOUT	02	A-2	SU-19
330	ug/kg	7	001	- VIIII acollo	S	SURFACE SOIL SAMPLE	DOUT	0-2	A-2	200
		-		Anthracana					,	61110

RICKENBACKERANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS – SOIL

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								_	
1-4	0-2	OUT	SURFACE SOIL SAMPLE	S	ALL SEMI-VOLATILES	Q.	¥ X	ug/kg	Y Z
A-4	0-2	DOUT	SURFACE SOIL SAMPLE	×	Antimony	4.8	ZNO	mg/kg	60
A-4	0-2	OUT	SURFACE SOIL SAMPLE	×	Arsonic	6.0	CNB	mg/kg	-
A-4	0-2	TUO	SURFACE SOIL SAMPLE	×	Boryllium	0.36	80	mg/kg	0.5
A-4	0-2	OUT	SURFACE SOIL SAMPLE	×	Cadmium	0.24	80	mg/kg	0.5
A-4	0-2	OUT	SURFACE SOIL SAMPLE	¥	Chrombm	3.6	7	mg/kg	-
A-4	0-2	OUT	SURFACE SOIL SAMPLE	Σ	Coppor	=	•	mg/kg	2.5
A-4	0-2	DOOT	SURFACE SOIL SAMPLE	Σ	Load	25.7		mg/kg	0.3
A-4	0-2	OUT	SURFACE SOIL SAMPLE	Σ	Mercury	90.0	2	mg/kg	0.1
A-4	0-2	DOUT	SURFACE SOIL SAMPLE	¥.	Nickol	11.9		mg/kg	4
A-4	0-2	DOCT	SURFACE SOIL SAMPLE	Z	Solentum	0.23	BW	mg/kg	0.5
A-4	0-2	DOUT	SURFACE SOIL SAMPLE	Σ	Silvor	0.78	ס	mg/kg	-
A-4	- 1	TUO	SURFACE SOIL SAMPLE	Σ	Thallium	0.53	CWND	InU/kg	-
A-4	- 11	OUT	SURFACE SOIL SAMPLE	¥	Zinc	34.4	3	mg/kg	2
A - S	0-0	THO	SUBFACE SOIL SAMPLE	NS.	Phenaulycone	280	-7	na/ka	930
A-5		DOCT	SURFACE SOIL SAMPLE	S	Fluoranthana	009	'	UU/ka	33
A-5	0-2	OUT	SURFACE SOIL SAMPLE	SV	Pyrone	570		un/ka	330
A-5	0-2	OUT	SURFACE SOIL SAMPLE	S	Benzo(a) Anthracene	200	ז	ug/kg	330
A-5	0-2	OUT	SURFACE SOIL SAMPLE	S	Chrysene	320	7	UQ/kg	330
A-5	0-2	OUT	SURFACE SOIL SAMPLE	S	Benzo(b) Fluoranthene	350	7	ug/kg	330
A-5	0-2	DOC	SURFACE SOIL SAMPLE	SV	Bonzo(a)Pyrene	250	7	ug/kg	330
A-5	0-2	OUT	SURFACE SOIL SAMPLE	S	Indeno(1,2,3-cd)Pyrene	8	7	ug/kg	330
A-5	0-2	100	SURFACE SOIL SAMPLE	S	Benzo(g,h,l)Perylene	35	7	ug/kg	330
A-5	0-2	OUT	SURFACE SOIL SAMPLE	Σ	Antimony	4.0	TWO	mg/kg	80
A-5	0-2	OUT	SURFACE SOIL SAMPLE	×	Arsenic	1.9	BN	mg/kg	-
A-5	0-2	OUT	SURFACE SOIL SAMPLE	Σ	Beryllium	9.0		mg/kg	0.5
A-5	0-2	OUT	SURFACE SOIL SAMPLE	*	Cadmium	0.24	80	mg/kg	0.5
A-5	0-2	OUT	SURFACE SOIL SAMPLE	Σ	Chromkin	8.8	2	mg/kg	-
A-5	0-2	OUT	SURFACE SOIL SAMPLE	×	Coppor	12.6	•	mg/kg	2.5
A-5	0-2	OUT	SURFACE SOIL SAMPLE	×	Lead	43.0		mg/kg	0.3
A-5	0-2	DOUT	SURFACE SOIL SAMPLE	*	Mercury	0.053	>	mg/kg	0.1
A-5	0-2	OUT	SURFACE SOIL SAMPLE	×	Nickol	12.2		mg/kg	*
A-5	0-5	DO	SURFACE SOIL SAMPLE	*	Solenium	0.21	¥n	mg/kg	0.5
A-5	0-2	TUO	SURFACE SOIL SAMPLE	×	Silvar	0.79	כ	mg/kg	-
A-5		OUT	SURFACE SOIL SAMPLE	×	The	0.54	rwnn	mg/kg	-
A-5	0-2	DOUT	SURFACE SOIL SAMPLE	×	Zinc	60.4	2	mg/kg	2

RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

T	_	_	0	0	0	0	0	9		s	2	-	2.5	0.3	_	4	0.5	_	_	7	5	000	9		0.5	0.5	-	2.5	0.3	<u>-</u>	- 1	5.0		- 0
	330	330	330	330	330	330	330			0.5	0.5		6	o	0.1		Ó				,	ë			0	•		~	0	0		0		
	ug/kg	ug/kg	ug/kg	ug/kg	UU/KU	սց/kը	ug/kg	mg/kg	24/61	100/kg	ma/ka	mg/kg	DA/GE	ma/kg																				
							7	CN.	BNS		5	₹	•		83		2	כ	CWND	7		7	CNU	3		60	2	•		5		<u>₹</u>	0	2 2
	200 7	280	570	260	290	520	260	4.0	10.0	0.96	0.23	14.7	10.6	52.9	60.0	21.4	0.2	0.70	0.54	108	9	5	4.5	15.2	0.67	0.22	14.1	17.8	60.4	0.056	24.1	0.10	0.74	95.7
	Phenanthrene	Fluoranthene	Pyrana	Bonzo(a) Anthracono	Chryseno	Bonzo(b) Fluoranthono	Benzo(a)Pyrene	Antlmony	Arsenic	Boryllum	Cadmlum	Chromhm	Copper	Pead	Marcury	Nickol	Solonium	Silvor	Thallium	Zinc	Firefree	a de la constante de la consta	Antimony	Arsonic	Boryllium	Cadmlum	Chrombm	Coppor	Lead	Mercury	Nicko	Solonium	Silvor	Zinc
	SV	SV	SV	SV	SV	SV	SV	Z	Σ	Σ	Σ	Z	Σ	\$	×	×	Σ	×	Z	Σ	70	\ \frac{1}{2}	×	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ.	X :	\$ 3	₹ ₹
	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	a large collegion	SIDEACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE					
	DOUT	TUO	OUT	OUT	DOUT	DOUT	TUO	DOUT	DOUT	DOUT	OUT	OUT	OUT	TUO	OUT	OUT	DOOT	DOOT	DOUT	TUO	Ė	2 2	100	OUT	001	000								
	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-5	0-2	0-2	0-2	0-2	0-2	0-2	0-5	c	1 1	010	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-5	0-5	0 0	0 0
	9-V	A-6	9-Y	9-V	A-6	A-6	A-6	A-6	A-6	A-6	A-6	A-8	9-V	A-6	A-6	A-6	9-Y	9-V	9-V	A-6	•	- 1	A-7	V-7	V-7	A-7	A-7							
	SU-23	SU-23	SU-23	SU-23	SU-23	SU-23	SU-23	SU-23	SU-23	SU-23	SU-23	SU-23	SU-23	SU-23	SU-23	SU-23	SU-23	SU-23	SU-23	SU-23		200	511-24	SU-24										

RICKENBACKERANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

SU-25 B-2 O-2 IN SURFACE SOIL SAMPLE SV Flooranthone 150 J ug/kg SU-25 B-2 O-2 IN SURFACE SOIL SAMPLE N Antimone 150 J ug/kg SU-25 B-2 O-2 IN SURFACE SOIL SAMPLE M Antimone 130 UNI mg/kg SU-25 B-2 O-2 IN SURFACE SOIL SAMPLE M Antimone 0.47 mg/kg SU-25 B-2 O-2 IN SURFACE SOIL SAMPLE M Coppor 2.3 NU mg/kg SU-25 B-2 O-2 IN SURFACE SOIL SAMPLE M Coppor 2.3 NU mg/kg SU-26 B-2 O-2 IN SURFACE SOIL SAMPLE M Coppor 2.3 NU mg/kg SU-26 B-2 O-2 IN SURFACE SOIL SAMPLE M Coppor 2.24 NU mg/kg SU-26 B-2 O
Substitute
B
B-2 0-2 IN SURFACE SOIL SAMPLE SV Filtornal
85 B-2 0-2 IN SURFACE SOIL SAMPLE SS B-4
2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2
25 25 25 25 25 25 25 25 25 25 25 25 25 2
2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2
2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2

RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

		_	_							
63	<u> </u>	0-2	Z	SURFACE SOIL SAMPLE	SV	ALL SEMI-VOLATILES	Q	¥2	ug/kg	X
۵	7-1	0-2	z	SURFACE SOIL SAMPLE	Σ	Antlmony	4.3	CND	mg/kg	8
ω	B-4	0-2	Z	SURFACE SOIL SAMPLE	×	Arsenic	7.	3	mg/kg	0.5
Δ.	B-4	0-2	z	SURFACE SOIL SAMPLE	¥	Boryllium	0.65		mg/kg	0.5
w	B-4	0-2	Z	SURFACE SOIL SAMPLE	×	Cadmlum	0.2	כ	mg/kg	-
w	8-4	0-2	Z	SURFACE SOIL SAMPLE	∑.	Chrombm	14.1	z	mg/kg	_
ω	8-4	0-2	Z	SURFACE SOIL SAMPLE	×	Соррог	20.2	•	mg/kg	2.5
ш	B-4	0-2	Z	SURFACE SOIL SAMPLE	Σ	Lead	59.0		mg/kg	0.3
ш	B-4	0-2	z	SURFACE SOIL SAMPLE	Z	Morcury	0.052	5	mg/kg	0.1
w	8-4	0-2	z	SURFACE SOIL SAMPLE	Σ	Nickel	23.5		mg/kg	4
ш	3-4	0-2	z	SURFACE SOIL SAMPLE	Σ	Salonlum	0.28	8	mg/kg	0.5
ш	B-4	02	Z	SURFACE SOIL SAMPLE	Σ	Silvar	0.71	כ	mg/kg	-
w	8-4	0-5	Z	SURFACE SOIL SAMPLE	2	Thallium	0.40	CWNO	mg/kg	-
	3-7	0-2	Z	SURFACE SOIL SAMPLE	2	Zing	09.3	7	mg/kg	2
_	3-5	0-2	Z	SURFACE SOIL SAMPLE	S	ALL SEMI-VOLATILES	2	ž	ug/kg	۲ ۲
w.	8-5	0-2	z	SURFACE SOIL SAMPLE	Σ	Antimony	4.0	CND	mg/kg	9
	3-5	0-2	z	SURFACE SOIL SAMPLE	×	Arsenic	10.0	BNJ	mg/kg	0.5
	B-5	0-5	Z	SURFACE SOIL SAMPLE	Σ	Borylllum	09.0		mg/kg	0.5
	3-5	02	z	SURFACE SOIL SAMPLE	¥	Cadmium	0.22	D	mg/kg	-
	3-5	0-2	z	SURFACE SOIL SAMPLE	Σ	Chromium	13.0	z	mg/kg	-
SU-20 E	3-5	0-5	Z	SURFACE SOIL SAMPLE	Σ	Соррог	17.4	•	mg/kg	2.5
	3-5	0-2	Z	SURFACE SOIL SAMPLE	Σ		43.2		mg/kg	0.3
	3-5	0-2	z	SURFACE SOIL SAMPLE	Σ	Mercury	0.084	2	mg/kg	0.1
SU-20 E	3-5	0-2	z	SURFACE SOIL SAMPLE	×	Nickol	10.0		mg/kg	•
	B5	0-2	Z	SURFACE SOIL SAMPLE	\$	Solonlun	0.48	æ	mg/kg	0.5
~	3-5	0-5	z	SURFACE SOIL SAMPLE	×	Silvar	0.75	2	mg/kg	-
-20 E	B-5	0-2	z	SURFACE SOIL SAMPLE	×	Thalllum	0.50	CWNO	mg/kg	_
1	3-5	0-2	z	SURFACE SOIL SAMPLE	×	Zinc	95	3	mg/kg	8

RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

									1	
SU-29	B-6	0-2	Z	SURFACE SOIL SAMPLE	S	Pyrana	260		no/ka	C.C.
SU-29	B-6	0-2	Z	SURFACE SOIL SAMPLE	S	Fluoranthana	250		ממ/אט	86
SU-29	8-6	0-2	Z	SURFACE SOIL SAMPLE	S	Banzo(a) Anthracena	140	7	un/ko	
SU-29	8-6	0-2	Z	SURFACE SOIL SAMPLE	S	Chrysone	8		מט/גם	930
SU-29	8-8	0-2	Z	SURFACE SOIL SAMPLE	SV	Bonzo(b) Fluoranthene	8	2	na/ka	OCC
SU-29	8-0	0-2	Z	SURFACE SOIL SAMPLE	SV	Bonzo(k) Fluoranthene	61		ua/ka	occ
SU-29	8-6	0-2	Z	SURFACE SOIL SAMPLE	SV	Bonzo(a)Pyrona	140	7	ua/ka	330
SU-28	B-6	0-2	Z	SURFACE SOIL SAMPLE	×	Antimony	4.7	N _D	ma/kg	9
SU-29	8-6	0-2	Z	SURFACE SOIL SAMPLE	×	Arsenic	11.0	3	ma/kg	0.5
SU-29	B-0	0-2	Z	SURFACE SOIL SAMPLE	Σ	Baryllium	0.93		ma/kg	0.5
SU-29	8-6	0-2	Z	SURFACE SOIL SAMPLE	Σ	Cadmium	0.22	כ	mg/kg	
SU-29	B-6	0-2	Z	SURFACE SOIL SAMPLE	¥	Chromhm	11.1	z	mg/kg	
SU-29	B-6	0-2	Z	SURFACE SOIL SAMPLE	×	Соррог	20.0	•	mg/kg	ĸ
SU-29	B-6	0-2	Z	SURFACE SOIL SAMPLE	Σ	Load	32.1		ma/ka	0.0
SU-29	B-6	0-2	Z	SURFACE SOIL SAMPLE	Σ	Marcury	0.050	2	ma/ka	0.1
SU-28	8-6	0-2	Z	SURFACE SOIL SAMPLE	×	Nickel	21.5		ma/ka	
50-29	0-0	0-2	Z	SURFACE SOIL SAMPLE	×	Solontum	0.34	BW	ma/ka	o
SU-29	8-6	0-2	Z	SURFACE SOIL SAMPLE	Σ	Silvar	0.77	0	mg/kg	
SU-29	B-6	0-2	Z	SURFACE SOIL SAMPLE	Σ	Thallium	0.49	ONW	mg/kg	
SU-29	B-8	0-2	Z	SURFACE SOIL SAMPLE	2	Zinc	109	z	ma/ka	

RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

SURFACE SOIL SAMPLE MAIGONIA MAIGONIA SURFACE SOIL SAMPLE MAIGONIA MAIGON	OUT SURFACE SOIL SAMPLE SV OUT SURFACE SOIL SAMPLE MM SURFACE SOIL SAMPLE MM OUT SURFACE SOIL SAMPLE MM SURFACE SOIL SAMPLE MM OUT SURFACE SOIL SAMPLE MM SURFACE SOIL SAMPLE MM OUT SURFACE SOIL SAMPLE MM SURFACE SO	OUT SURFACE SOIL SAMPLE OUT SU
SURFACE SOIL SAMPLE SUBFACE SOIL SAMPLE	00 1 10 0 0 1 10 0 0 1 10 0 0 1 10 0 0 1 10 10	

RICKENBACKERANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

SAMPLE	GRID	DEPTH	IN/OUT	DESCRIPTION	САТЕВОПУ	FOR	RESULTS	HESULTS QUALFIER UNITS	1871	UMIT
SU-3		0-2	Z	SURFACE SOIL SAMPLE	SV	. ALL SEMI-VOLATILES	Q	A N	սց/kg	A X
SU-32	C-2	0-2	Z	SURFACE SOIL SAMPLE	×	Antimony	6.4	r N O	mg/kg	9
SU-32			Z	SURFACE SOIL SAMPLE	Σ	Arsonic	10.1	3	mg/kg	0.5
SU-3		0-2	<u>z</u>	SURFACE SOIL SAMPLE	Σ	Baryllium	0.73		mg/kg	0.5
SU-3		0-2	Z	SURFACE SOIL SAMPLE	Σ	Cadmlum	0.49	E	mg/kg	-
SU-32	2 C-2	0-2	Z	SURFACE SOIL SAMPLE	Σ	Chrombm	16.2	7	mg/kg	-
SU-3		0-2	Z	SURFACE SOIL SAMPLE	Σ	Copper	22.6	•	mg/kg	2.5
SU-3		0-2	Z	SURFACE SOIL SAMPLE	Σ	Load	41.6		mg/kg	0.3
SU-3		. 0-2	Z	SURFACE SOIL SAMPLE	Σ	Morcury	190.0	כ	mg/kg	0.1
SU-3		0-2	Z	SURFACE SOIL SAMPLE	×	Nickol	27.3		mg/kg	₹
SU-3			Z	SURFACE SOIL SAMPLE	Σ	Salanlum	0.10	3	mg/kg	0.5
SU-32		0-2	Z	SURFACE SOIL SAMPLE	×	Silvar	0.01	ס	mg/kg	-
C-OS			Z	SURFACE SOIL SAMPLE	*	Thallium	0.40	WLAU	mg/kg	-
SU-32		0-2	Z	SURFACE SOIL SAMPLE	M	Zinc	113	3	mg/kg	2

RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

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N. V.	L							
	2	SUBFACE SOIL SAMPLE	S	Phonanthrono	820		ug/kg	330
		SURFACE SOIL SAMPLE	S	Anthracono	100	7	ug/kg	330
	Z	SURFACE SOIL SAMPLE	SV	Fluoranthone	1900		ug/kg	330
	Z	SURFACE SOIL SAMPLE	>5	Pyrene	2300		ug/kg	330
	Z	SURFACE SOIL SAMPLE	S	Bonzo(a) Anthracono	1400		սց/kg	330
	Z	SURFACE SOIL SAMPLE	SV	Chrysono	1400		ug/kg	930
	Z	SURFACE SOIL SAMPLE	SV	Benzo(b) Fluoranthene	1400		ug/kg	330
	Z	SURFACE SOIL SAMPLE	SV	Benzo(k) Fluorantheno	1200		ug/kg	330
	Z	SURFACE SOIL SAMPLE	SV	Benzo(a)Pyrene	1300		ug/kg	230
	Z	SURFACE SOIL SAMPLE	SV	Indono(1,2,3-cd)Pyrana	8		ug/kg	330
	z	SURFACE SOIL SAMPLE	SV	Dibonz(a,h) Anthracone	240	7	ug/kg	330
	Z	SURFACE SOIL SAMPLE	S	Benzo(g,h,l)Porylana	200		ug/kg	330
	Z	SURFACE SOIL SAMPLE	Σ	Antimony	10	TNS	mg/kg	9
	Z	SURFACE SOIL SAMPLE	Σ	Arsonic	20		mg/kg	0.5
	Z	SURFACE SOIL SAMPLE	Σ	Borylllum	0.75	_	mg/kg	0.5
	Z	SURFACE SOIL SAMPLE	Σ	Cadmlum	0.63		mg/kg	_
	Z	SURFACE SOIL SAMPLE	Σ	Chrombm	13.2		mg/kg	-
	z	SURFACE SOIL SAMPLE	Σ	Coppor	25.5	3	mg/kg	2.5
	Z	SURFACE SOIL SAMPLE	Σ	Lond	112	7.2	mg/kg	0.3
	Z	SURFACE SOIL SAMPLE	Σ	Mercury	0.064	>	mg/kg	0.1
	Z	SURFACE SOIL SAMPLE	Σ	Nickel	24.5		mg/kg	4
	Z	SURFACE SOIL SAMPLE	Σ	Selonlum	0.49	BW	mg/kg	0.5
	Z	SURFACE SOIL SAMPLE	Σ	Silvar	0.03	5	mg/kg	-
	Z	SURFACE SOIL SAMPLE	Σ	Thallium	0.19	BNWJ	mg/kg	-
	Z	SURFACE SOIL SAMPLE	×	Zinc	124	3	mg/kg	2

RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

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42		•	ח מ	s ⁱ		- 6	0.3	0.1	•	0.5				8	3	מר ה ה	330	330	330	330		0.5	0.5	÷		2.5	0.3	0.1	•	>	
110/kg	100/kg	Day And	DA/DIII	Da/Dui	04/00	24/01	ma/ka	mg/kg	ma/ku	mg/kg	mg/kg	mg/kg	mg/kg	04/013	2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2	U0/kg	ua/ka	Ua/ka	ug/kg	ug/kg	mg/kg	mg/kg	mg/kg	пр/ка	mg/kg	mg/kg	mg/kg	mg/kg	Dx/DE	מא/מייי	מייטיייייייייייייייייייייייייייייייייי
AZ.	2	2 2	2	=	2	2 *		>		BW	כ	TWN0	3	-	•	-	7	7	7	7	CN2	BNWJ		8	7	•		5		5 5	•
C		7	2.5	2.5	, t	20.6	19.5	0.057	24.0	0.39	0.7	0.52	75.5	230	3	340	180	210	250	500	4.5	7.9	67.0	0.23	14.3	17	27.1	0,062	, c	7.0	;
	Antimony	Character	Section 2	The state of the s	Chromba	Connec	Load	Mercury	Nickel	Selenium	Silver	Thallium	Zlric	Phonanthrone	Finoreythone	Pyrana	Benzo(a) Anthracene	Chrysono	Benzo(b) Fluoranthone	· Bonzo(a)Pyrona	Antlmony	Arsonic	Barylllum	Cadmlum	Chromium	Coppor	Pand	Morcury	Soloolim	zoollo.	1
S	**	3	2	2	2	*	Σ	Σ	₹	Σ	Σ	Σ	Σ	S	>5	8	S	SV	SS	S	Σ	₹ :	Σ	∑ :	Σ	∑ :	Σ:	Σ 3	E Z	2	
SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SUBFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SUMPACE SOIL SAMPLE	SURFACE SOIL SAMPLE	1						
-	Z	Z	Z	2	Z	Z	Z	Z	z	Z	Z	Z	Z	Z	Z	Z	Z	Z	z	2	2	Ξ:	2 :	Ξ:	2 :	2 2	2 2	<u> </u>	Z	Z	:
0-2	0-2	t		- 1	t	0-2	0-2	0-2	0-2	1	1	0-5	0-2	0-2	0-2	- 1	0-2	1	1		1		0 0	1	ı	0 0		0-12	•	0-2	•
2-2	C-5	0-12	0-5	0 - 8	C-5	0-5	2-5	0 - 0	0-0	0-0	0-0	0-19	0 1 0	φ (9 0 0 0	ָרָ פְּ	ĵ	מ נ נ	<u></u>	י ט נ) (9 0	0-0	C-8	,						
SU-34	SU-34	SU-34	SU-34	SU-34	SU-34	SU-34	SU-34	1	SU-34	1	SU-34		SU-34	SU-35	SU-35	SU-35	SU-35	SU-35	SU-35	SU-35	SU-35	000	2 2 2	מרון אני	200	2017	SU-35	SU-35	SU-35	SU-35	20 113

RICKENBACKER ANG B HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

X X	9	0.5	0.5	-	-	2.5	0.3	1.0	*	0.5	-	-	~		¥ X	9	0.5	0.5	-	_	2.5	0.3	0.1	4	0.5	-	_	2	
ug/kg	mg/kg	mu/kg	mg/kg		ug/kg	mg/kg	מא/פניי	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg																	
AN	3	83	60	9		•	•	D .	•	BW	5	CWND	7.2	_	X X	3			•		•	•	ב	•	BW	כ	CWND	7.2	
QV	4.1	6.6	0.41	10.0	0.7	15	24.8	0.042	17.9	0.25	0.66	0.47	60.2		2	4.2	17.6	0.74	0.2	15.7	21.9	20.7	990.0	20.6	0.25	0.7	0.54	95.1	
ALL SEMI-VOLATILES	Antimony	Arsonic	Boryllum	Cadmlum	Chrombin	Coppor	Load	Morcury	Nickol	Salonium	Silvor	Thalllurn	Zinc		ALL SEMI-VOLATILES	Antimony	Arsonic	Boryllium	Cadmlum	Chromium	Coppor	Load	Mercury	Nickol	Salonium	Silvar	Thallium	Zinc	
S	×	\$	Z	×	Σ	¥	Σ	¥	Σ	×	2	Σ	¥		SV	×	¥	Σ	Σ	Σ	×	*	Σ	×	Σ	×	Σ	Σ	
SURFACE SOIL SAMPLE		SURFACE SOIL SAMPLE																											
TUO	DOUT	DOUT	TUO	DOUT	OUT	TUO	DOUT	OUT	DOUT	DOUT	TUO	DOUT	OUT		OUT	DOUT	DOUT	TUO	TUO	OUT	OUT	OUT							
0-2	0-5	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-5		0-5	0-2	0-5	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	
C-7		0-0	0-D	0-0	C-0	0-0	0-0	C-0	0-0	0-0	0-0	C-8	0-0	C-0	0-0														
SU-36	SU~30		SU-37	50-37	SU-37	SU-37	20-37	SU-37																					
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RICKENBACKERANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS – SOIL

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330	3	٥	0.5	0.5	-	_	2.5	0.3	0.1	*	0.5	-	-	2	0.00	3 6	000	000	3 6	930	000	9	0.5	0.5	_	_	2.5	0.3	0.1	4	0.5	_	_
na/ka		mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	rng/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	: 1	0.00	US/KU	US/KU	מלאט מילאט	ug/kg	ug/kg	mg/kg	ma/kg										
7		3	æ		•0		•	•	ס	•	BW	2	BNWJ	J.X	-	-	· ·	- د	, 7	7	7	CND			*B		•	•	8	•	BW	2	BNWJ
55		2.0	12.1	0.63	0.30	22.0	30.6	73.4	0.050	26.4	0.29	0.03	99.0	196	•	3 8	35	2 0	200	170	150	5.2	19.6	0.70	60.0	22.7	36.1	44.3	420.0	30.5	0.32	00.0	0.85
Pyrene		Animony	Arsenic	Baryllum	Cadmlum	Chrombm	Coppor	Load	Mercury	Nickol	Solonium	Silvor	Thalllum	Zinc	O Control of		Floorantione	Lylone	Banzo(li) Fluoranthenn	Bonzo(k) Fluoranthone	Benzo(a)Pyrene	Antimony	Arsenic	Baryllium	Cadmlum	Chrombin	Coppor	Load	Morcury	Nickol	Solonium	Silvar	Thaillum
\S.	; =	Σ	Σ	×	Σ	Σ	×	×	×	¥	Σ	\$	Σ	×	20	3 3	200	2	> >s	SS	SV	Σ	Σ	Σ	¥	Σ	Σ	Z	Σ	Σ	Σ	¥	≥
SURFACE SOIL SAMPLE		SURFACE SOIL SAMPLE	PIDEACE COLLEGATOR	STIDENCE COLL CANDLE	SURFACE SOIL SAMPLE	SOUTHER SOIL SAMPLE	SUBFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE												
2		Z	z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	2	3	2 2	2 2	Z	Z	Z	Z	Z	z	Z	Z	Z	Z	Z	Z	Z	Z	Z
0-0		7-0	0-2	0-2	0-2	0-2	0-2	0-5	0-5	0-5	0-2	0-2	0-2	0-2	5 - 0	- 1	1 1	2 0	0-2	0-2	0-2	02	0-5	0-5	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2
0-2		2-0	D-2	D-2	0-2	0-2	D-2	D-2	0-2	D-2	D-2	D-2	D-2	0-2	ć	- 1	2 6		2 -3	0-3	0-3	0-3	0-3	0-3	0-3	0-3	0-3	D-3	0-3	0-3			2-3
SU-39		55-06	SU-39	SU-38	SU-39	SU-39	SU-39	211-10	611-10	000	27 1 1 2	SU-40	SU-40	SU-40	SU-40	SU-40	SU-40	SU-40	SU-40	SU-40	SU-40	SU-40	SU-40	SU-40	SU-40	SU-40							

RICKENBACKERANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

UNITS LIMIT	000 000 ng/kg	00cc gx/gn		ug/kg		mg/kg	0		mg/kg	mg/kg	mg/kg 2.5	mg/kg 0.3	mg/kg 0.1	mg/kg	mg/kg 0.5		mg/kg	mg/kg			ug/kg 330	mg/kg 6	mg/kg 0.5	mg/kg 0.5	1 dy/kg	mg/kg 1	mg/kg 2.5	mg/kg 0.3	mg/kg 0.1	mg/kg 4	mg/kg 0.5	mg/kg 1	mg/kg 1
OVAUFIER UNITS LIMIT	۔	7	ح ح	7	7	TWO		6	5		•	•	ב	•	כ	כ	BNWJ	ר.א				- CND		80	ė		-	•	ב	•	80	ם	CWND
RESULTS	330	300	200	240	170	5.1	17.0	0.30	0.24	12.0	25.0	39.2	0.081	21.5	0.22	0.00	9.0	0.70		8	8	5.3	16.7	0.53	0.26	. 16	32.6	35	0.067	30.0	00.00	20.0	0.61
FOR	Fluoranthone	Pyrana	Chrysene	Bonzo(b) Fluoranthone	Bonzo(a)Pyrom	Antlmony	Arsenic	Beryllium	Cadmium	Chromkm	Соррег	Load	Morcury	Nickel	Solonlum	Silvar	Thallium	Zinc	i	Fluoranthono	Pyrana	Antimony	Arsonic	Boryllium	Cadmlum	Chromium	Copper	Load	Mercury	Nickel	Solenium	Silvar	Thaillurn
САТЕВОПУ	\s	SV	SV	SV	SV	×	≆	Σ	Σ	Σ	Σ	Σ	Σ	Σ	¥	Σ	Z	¥	i	> :)S	Σ	Σ	Σ	Σ	×	Σ	¥	Σ	Σ	\$	Σ	₹
DESCRIPTION	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	1 0 10 10 10 10 10 10 10 10 10 10 10 10	SURFACE SOIL SAMPLE													
IN/OUT	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	2	-	= :	2	Z	Z	Z	Z :	Z	Z	z	Z	Z.	Z	Z	Z
DEPTH	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-5	0-2	1	0-2	c	0 0	0-2	0-2	0-2	0-2	0-2	0-2	0-5	0-2	0-5	0-5	0-2	0-2	0-2
GRID .	D-4	0-4	D-4	D-4	D-4	0-4	D-4	D-4	0-4	0-4	0-4	D-4	0-4	0-4	0-4	0-4	0-4	0-4		ם מ	י כ	0-5	0-5	0-5	0-S	0-5	0-5	0-5	0-5	1	1	ï	0-2
SAMPLE .	SU-41	SU-41	SU-41	SU-41	SU-41	SU-41	SU-41	SU-41	SU-41	SU-41	SU-41	SU-41	SU-41	SU-41	SU-41	SU-41	SU-41	SU-41	27-110		200	20-42	SU-42										

. DEF	TH IN/OUT	DESCRIPTION	CATEGORY	ANALYSIS	RESULTS	QUAUFIER UNITS	11.5	DETECTION
	0-2	SURFACE SOIL SAMPLE	S	ALL SEMI-VOLATILES	9	AN	ug/kg	AN
	0-2	SURFACE SOIL SAMPLE	×	Anthrony	4.7	ZYS	mg/kg	60
	0-2	SURFACE SOIL SAMPLE	₹	Arsenic	0.3	8	mg/kg	0.5
	0-2 IN	SURFACE SOIL SAMPLE	\$	Boryllum	0.50		mg/kg	9.0
	0-2 IN	SURFACE SOIL SAMPLE	Σ	Cadmlum	0.22	5	mg/kg	-
	0-2	SURFACE SOIL SAMPLE	Σ	Chromhim	17.5		mg/kg	-
	0-2	SURFACE SOIL SAMPLE	Σ	Coppor	29.2	•	mg/kg	2.5
	0-2	SURFACE SOIL SAMPLE	Σ	Load	26.4	•	mg/kg	0.3
	0-2	SURFACE SOIL SAMPLE	Σ	Mercury	0.00	>	mg/kg	0.1
	0-2 IN	SURFACE SOIL SAMPLE	Σ	Nickel	31.0	•	mg/kg	•
	0-2	SURFACE SOIL SAMPLE	Σ	Solenlum	0.25	83	mg/kg	0.5
	0-2	SURFACE SOIL SAMPLE	Σ	Silvar	0.76	5	mg/kg	-
	0-2 IN	SURFACE SOIL SAMPLE	Σ	Thaillum	0.54	CWND	mg/kg	-
	0-2	A SURFACE SOIL SAMPLE	Σ	Zinc	110	?.x	mg/kg	2

RICKENBACKERANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

SAMPLE . GRID									-	
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SU-44	D-7	0-2	DOUT	SURFACE SOIL SAMPLE	SV	2 - Chlorophenol	0	7	ug/kg	330
SU-44	D-7	0-2	OUT	SURFACE SOIL SAMPLE	SV	3-Nitroaniline	24	7	ug/kg	1600
SU-44	D-7	0-2	TUO	SURFACE SOIL SAMPLE	S	Aconaphthene	8	7	ug/kg	330
SU-44	0-7	0-2	DO	SURFACE SOIL SAMPLE	S	4 - Nitroanilino	30	7	ug/kg	1600
SU-44	0-7	0-2	DOUT	SURFACE SOIL SAMPLE	SV	Phonanthrono	10	7	ug/kg	330
SU-44	D-7	0-2	DOUT	SURFACE SOIL SAMPLE	S	bis(2-Chlorethyl)ether	0	7	ug/kg	330
SU-44	0-7	0-2	OUT	SURFACE SOIL SAMPLE	SV	Anthracene	17	7	ug/kg	330
SU-44	0-7	0-2	TUO	SURFACE SOIL SAMPLE	SV	Fluoranthone	10	7	ug/kg	330
SU-44	0-7	0-2	TUO	SURFACE SOIL SAMPLE	S	Pyrana	39	7	ug/kg	230
SU-44	0-7	0-2	DOUT	SURFACE SOIL SAMPLE	SV	Benzo(a) Anthracene	36	2	ug/kg	330
SU-44	D-7	0-2	TUO	SURFACE SOIL SAMPLE	SV	Chrysane	36	7	Ug/kg	330
SU-44	07	0-2	OUT	SURFACE SOIL SAMPLE	SV	Bonzo(b) Fluoranthene	40	7	ug/kg	930
SU-44	0-7	0-2	OUT	SURFACE SOIL SAMPLE	SV	Bonzo(a)Pyrene	20	7	ug/kg	330
SU-44	0-7	0-2	DOUT	SURFACE SOIL SAMPLE	26	Indono(1,2,3-cd)Pyrane	20	7	ug/kg	220
SU-44	0-7	0-2	OUT	SURFACE SOIL SAMPLE	SV	Benzo(g,h,l)Perylene	20	7	ug/kg	330
SU-44	D-7	0-2	DOUT	SURFACE SOIL SAMPLE	*	Antimony	4.7	3	mg/kg	9
SU-44	2-0	0-2	DOUT	SURFACE SOIL SAMPLE	Σ	Arsenic	7.0	8	mg/kg	0.5
SU-44	D-7	0-2	DOCT	SURFACE SOIL SAMPLE	Σ	Beryllium	0.47	8	mg/kg	0.5
SU-44	D-7	0-2	TUO	SURFACE SOIL SAMPLE	Σ	Cadmlum	1.6	•	то/ка	-
SU-44	D-7	0-2	100	SURFACE SOIL SAMPLE	×	Chrombm	14		mg/kg	-
SU-44	2-0	0-2	DOUT	SURFACE SOIL SAMPLE	Σ	Coppor	45.3	•	mg/kg	2.5
SU-44	0-7	0-2	OUT	SURFACE SOIL SAMPLE	Z	Lead	77	•	mg/kg	0.3
SU-44	2-0	. 0-2	DO	SURFACE SOIL SAMPLE	×	Mercury	0.055	כ	mg/kg	0.1
SU-44	0-7	0-2	DOUT	SURFACE SOIL SAMPLE	×	Nickol	17.9	•	mg/kg	₹
SU-44	0-7	0-2	OUT	SURFACE SOIL SAMPLE	×	Solonlum	0.17	š	mg/kg	0.5
SU-44	7-0	0-2	DO	SURFACE SOIL SAMPLE	×	Silvar	0.77	2	mg/kg	-
SU-44	D-7	0-2	OUT	SURFACE SOIL SAMPLE	Σ	Thallium	0.45	CWNU	mg/kg	-
SU-44	0-7	0-2	OUT	SURFACE SOIL SAMPLE	×	Zinc	68.3	N.J	mg/kg	2

330	330	330	33	330	330	330	330	330	330	330	230	9	0.5	0.5	-	-	2.5	0.3	0.1	7	0.5	-	-	
	_			_		_	_	_		_	_							_	_					
ug/kg	ug/kg	US/KI	מט/אנו	טמ/אט	ng/kg	ug/kg	ug/ku	Ug/ku	ug/kg	ug/ku	טמ/גט	mg/kg	ma/ka	ma/ka	ma/ka	mg/kg	ma/kg	2/60						
	7											25	8	כ	•		•	•	2	•	35	ס	UNW	ž
11000	2200	23000	25000	15000	17000	20000	14000	15000	10000	3500	0000	4.3	4.0	0.25	1.9	12.5	11.9	32.4	0.053	13.0	0.19	0.71	0.5	43.9
Phonanthreno	Anthracono	Fluoranthene	Pyrono	Bonzo(a) Anthracono	Chrysana	Benzo(k) Fluoranthene	Banzo(b) Fluoranthana	Bonzo(a)Pyrone	Indeno(12,3-cd)Pyrene	Dibenz(a,h) Anthracene	Bonzo(g.h.i)Porylana	Antimony	Arsonic	Baryllium	Cadmium	Chrombin	Coppor	Lead	Mercury	Nickel	Salonlurn	Silvar	Thalllum	Zing
S	SV	SV	SV	S	S	SV	S	S	SV	SV	SV	Z	Σ	Σ	Σ	Σ	Σ	×	Σ	Σ	Σ	Σ	X	Σ
SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE						
DOUT	001	OUT	OUT	OUT		OUT	OUT		OUT	OUT	OUT	OUT	OUT	001	OUT	OUT	OOL	OUT						
0-5	0-5	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-5	0-2	0-2	0-5	0-2	0-2	0-5	0-2
E - 2	E-2	E-2	E-2	E-2	E-2	E-2	E-2	E-2	E-2	E-2	E-2	E-2	E-2	E-2	E-2	E-2	E-2	E-2	E-2	E-2	E-2	E-2	E-2	E-2
SU-45	SU-45	SU-45	SU-45	SU-45	SU-45	SU-45	SU-45	SU-45	50-45	SU-45														

RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

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TS LIMIT	330								ug/kg				mg/kg 6	mg/kg 0.5	mg/kg 0.5	mg/kg	mg/kg 1	mg/kg 2.5	mg/kg 0.3	mg/kg 0.1		mg/kg 0.5	mg/kg	mg/kg 1	,
GOAUTIEN UNITS		7						_	_	_	7	_	CND	E	6	•	E	•	•	ם	•	BW	E	- CWNU	_
	5100	970	7500	11000	2000	6700	0400	9000	0099	4000	1900	4700	4.2	7.2	0.40	1.0	13.0	10.4	54.9	0.058	19.7	0.23	1.0	0.51	
	Phonanthrone	Anthracono	Fluoranthone	Pyrano	Benzo(a) Anthracene	Chrysone	Bonzo(b) Fluoranthone	Benzo(k) Fluoranthone	Banzo(a)Pyrana	Indono(1,2,3-cd)Pyrana	Dibonz(a,h) Anthracona	Bonzo(g.h.i)Porylona	Antimony	Arsonic	Boryllum	Cadmlum	Chromhm	Coppor	Load	Morcury	Nickol	Solonium	Silver	Thallium	
	SV	SV	SV	AS .	SV	SV	SV	S	S	SS	S	SV	Σ	×	Z	Σ	×	Σ	Σ	Σ	Σ	Σ	×	×	-
SUBFACE SOIL SAMPLE		SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	TIGHT TO					
	100	TUO	TUO	DOUT	DOUT	DOUT	TUO	DOOT	TUO	TUO	TUO	OUT	OUT	OOT	OUT	OUT	OUT	100	OUT	OUT	OUT	DOOT	OUT	100	1
	0-2	0-2	0-2	0-2	0-2	0-2	0-5	0-2	0-5	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-5	0-5	0-2	0-2	0-2	0-2	0-5	0-2	
	E-3	E-3	E E	E-3	E-3	E-3	E-3	E-3	E-3	E-3	E-3	E-3	E-3	E-3	1 1 1	E - 3	E-3	C	E-3	E-3	ш С-	C .	E - 3	E-3	U
	SU-46	SU-46	SU-46	SU-40	SU-46	SU-46	SU-46	SU-46	SU-46	SU-46	SU-40	SU-46	SU-46	SU-46	SU-46	SU-46	SU-46	SU-46	SU-46	20-46	SU-46	SU-46	SU-46	SU-46	07

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2 E	000	330	8	0.5	0.5	-	-	2.5	0.3	1.0	4	0.5	-	-	2	330	330	330	330	330	9	0.5	0.5	-	-	2.5	0.0	0.1	*	0.5	-	-	8
01100	ug/kg	ug/kg	mg/kg	gy/gu	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	ng/kg	ug/kg	ug/kg	ug/kg	ug/ku	mg/kg	ma/ka																	
down ren	٦	7	3			•		•	•	ב	•	BW	>	CWWD	7.2	ז	7	7	7	7	3	<i>s</i>	8	•		•	•	כ	•	BW	>	CWND	ž
NESOLIS Q	160	130	4.7	14.4	0.59	0.59	10	29.3	41.2	0.056	29.6	0.44	0.77	0.49	111	170	300	270	170	130	4.9	10.0	0.49	0.05	15.4	32.7	43.7	90.0	31.6	0.59	0.0	0.56	5
201	Fluoranthono	Bonzo(b) Fluoranthono	Antimony	Arsenic	Boryllum	Cadinium	Chrombin	Coppor	Load	Mercury	Nickol	Salanlum	Silver	Thallium	Zinc	Phonanthrono	Fluoranthona	Pyrana	Chrysono	Banzo(a)Pyrene	Anlimony	Arsenic	Boryllium	Cadmlum	Chrombm	Copper	Lead	Mercury	Nickel	Selonlum	Silvar	Thalllum	Zluc
CALEGORY	S	SV	X	×	×	Σ	Σ	Σ	×	Z	Σ	Σ	Σ	×	Σ	 S	S	SV	s	\s.	Σ	×	Σ	Σ	Σ	Σ	Σ	Σ	Σ	×	Σ	X	77
DESCRIPTION	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	CHERNE CON CALIDIE
IN/OUT	DOUT	DOOT	TUO	TUO	TUO	DOUT	TUO	TUO	TUO	TUO	DOUT	DOUT	TUO	DOUT	DOUT	DOUT	TUO	OUT	DOUT	TUO	DOUT	DOUT	TUO	DOUT	DOUT	TUO	TUO	DOOT	DOOT	DOUT	OUT	OUT	1
DEPTH	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-5	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	4
GRID •	E-4	E-4	E-4	E-4	E-4	E-4	E-4	A-3	H-1	E-4	E-4	E-4	E-4	H - H	E-4	E-6	E-6	E-6	E-6	E-8	E-6	E-6	E-6	E-6	E-6	E-6					E-6		
SAMPLE	SU-47	SU-47	SU-47	SU-47	SU-47	SU-47	SU-47	SU-47	SU-47	SU-47	SU-47	SU-47	SU-47	SU-47	SU-47	SU-48	SU-48	SU-40	SU-40	SU-48	SU-48	SU-40	SU-40	SU-40	SU-40	SU-40	SU-48	SU-40	SU-48	SU-48	SU-48	SU-48	
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0-2	Z	SURFACE SOIL SAMPLE	SV	Phonanthrene	240	-	10/kg	5
0-2		SURFACE SOIL SAMPLE	S	Fluoranthong	450	,	nu/ka	000
0-2	Z	SURFACE SOIL SAMPLE	S	Pyrono	84		מט/גם	000
0-2	Z	SURFACE SOIL SAMPLE	S	Benzo(a) Anthracene	220	3	המ/אם	330
0-2	Z	SURFACE SOIL SAMPLE	SV	Chrysono	250	.7	ug/kg	330
0-2	Z	SURFACE SOIL SAMPLE	SV	Banzo(k) Fluoranthona	130	7	ua/ka	330
1	Z	SURFACE SOIL SAMPLE	SS	Banzo(b) Fluoranthene	320	7	ua/ko	330
1		SURFACE SOIL SAMPLE	SV	Bonzo(a)Pyrono	220	. 7	ug/kg	330
0-2	Z	SURFACE SOIL SAMPLE	SV	Indono(1,2,3-cd)Pyrana	140	7	ua/ka	33
0-2	Z	SURFACE SOIL SAMPLE	2	Antimony	0,4	ON O	ma/ka	9
0-2	Z	SURFACE SOIL SAMPLE	Σ	Arsonic	16.0	S	mo/kg	50
0-2		SURFACE SOIL SAMPLE	Σ	Boryllium	0.49	60	mo/ka	0.5
0-2		SURFACE SOIL SAMPLE	Σ	Cadmium	0.05	٠	ma/ka	-
1	<u>z</u>	SURFACE SOIL SAMPLE	×	Chrombm	15.4		mg/kg	-
0-2	_	SURFACE SOIL SAMPLE	×	Coppor	32.7	•	ma/kg	2.5
1		SURFACE SOIL SAMPLE	×	Load	43.7	•	mg/kg	0.3
0-2		SURFACE SOIL SAMPLE	×	Mercury	00.00	כ	ma/ka	0.1
1		SURFACE SOIL SAMPLE	×	Nickel	31.6	•	ma/kg	*
1		SURFACE SOIL SAMPLE	Σ	Solonlum	0.59	MB	mg/kg	0,5
1		SURFACE SOIL SAMPLE	Σ	Silvar	0.0	ס	mg/kg	-
t		SURFACE SOIL SAMPLE	Σ	Thallium	0.58	CWNO	mg/kg	-
0-5	Z	SURFACE SOIL SAMPLE	Σ	Zinc	190	L.N	mg/kg	2
3-5	Z	AUGER BORING	>	ALL VOLATILES	QX		ua/ka	2
3-5	Z	AUGER BORING	SV	ALL SEMI-VOLATILES	QX	AN	na/ka	AN
1	Z	AUGER BORING	×	Anlimony	4.0	UNB	mo/kg	
1	Z	AUGER BORING	Σ	Arsonic	22.9	2	mo/kg	, -
1	Z	AUGER BORING	×	Boryllum	0.59		ma/ka	
3-5		AUGER BORING	×	Cadmlum	0.47	8	mo/kg	0.5
1		AUGER BORING	×	Chrombin	17.3		mg/kg	-
3-5		AUGER BORING	Σ	Coppor	34.7		mg/kg	2.5
1		AUGER BORING	Σ	Load	22		ma/kg	0.3
3-5	Z	AUGER BORING	Σ	Mercury	0.059	5	mo/ka	0
3-5		AUGER BORING	×	Nickol	35.3		ma/ka	*
1	Z	AUGER BORING	X	Solonlum	0.2	CWND	mo/ka	
ı		AUGER BORING	X	Silvor	0.70	2	ma/ka	?
ï		AUGER BORING	Σ	Thalllum	1.2	DAB.	ma/kg	-
3-2	Z	AUGER BORING	×	Zinc	101		ma/ka	

RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

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	AB1-552	B-2	0-10	Z	AUGER BORING	>	Ethylbanzona	6700		ug/kg	2
	AB1-552	B-2	0-10	Z	AUGER BORING	>	m/p-Xylano	0009		ug/kg	S
	AB1-552	B-2	0-10	Z	AUGER BORING	>	o-Xylana	12000		ug/kg	S
	AB1-552	B-2	0-10	Z	AUGER BORING	SV	Naphthalone	130	2	ug/kg	330
	AB1-552	B-2	0-10	Z	AUGER BORING	Σ	Antimony	4.7	UNA	mg/kg	9
	AB1-552	B-2	0-10	Z	AUGER BORING	Σ	Arsonlc	16.1	2	. mg/kg	-
	AB1-552	B-2	0-10	Z	AUGER BORING	Z	Baryllium	0.35	В	mg/kg	0.5
	AB1-552	B-2	0-10	Z	AUGER BORING	Z	Cadmlum	0.35	В	mg/kg	0.5
	AB1-552	8-2	0-10	Z	AUGER BORING	Σ	Chromium	10.0		mg/kg	
	AB1-552	B-2	8-10	Z	AUGER BORING	×	Copper	25,6		mg/kg	2,5
	A81-552	B-2	8-10	Z	AUGER BORING	×	Load	22.0	•	mg/kg	0.3
	AB1-552	8-2	0-10	Z	AUGER BORING	Σ	Mercury	0.050	2	mg/kg	0.1
	AB1-552	8-2	01-0	Z	AUGER BORING	Σ	Nickel	29		mg/kg	•
	A81-552	8-2	01-0	Z	AUGER BORING	Σ	Salantum	0.52	BNWJ	mg/kg	0.5
	AB1-552	8-2	0-10	z	AUGER BORING	Σ	Silver	0.77	2	mg/kg	-
	AB1-552	8-2	01-0	Z	AUGER BORING	Σ	Thallium	0.47	BNU	mg/kg	-
- 1	AB1-552	B-2	0-10	Z	AUGER BORING	Σ	Zinc	92.1		mg/kg	2
	1	ı									
	A82-551	8-4	3-2	Z	AUGER BORING	>	Bonzone	-	7	ug/kg	S
	AB2-551	9-4	3-5	Z	AUGER BORING	SV	Naphthalano	1200		ug/kg	330
	AB2-551	8-4	3-5	Z	AUGER BORING	Σ	Antimony	4.6	UNR	mg/kg	9
	AB2-551	8-4	3-5	z	AUGER BORING	Σ	Arsonic	10.0	3	mg/kg	-
	AB2-551	8-4	3-5	Z	AUGER BORING	Σ	Boryllum	0.57	8	mg/kg	0.5
	AB2-551	B-4	3-5	Z	AUGER BORING	Σ	Cadmlum	0.34	83	mg/kg	0.5
	AB2-551	9-4	3-5	z	AUGER BORING	Σ	Chromium	14.9		па/ка	-
	AB2-581	8-4	3-6	Z	AUGER BORING	Σ	Copper	20.3		mg/kg	2.5
	AB2-581	8-4	3-5	Z	AUGER BORING	Σ	Load	17.9		mg/kg	0.3
	AB2-551	8-4	3-5	Z	AUGER BORING	Σ	Mercury	0.057	כ	mg/kg	0.1
	AB2-581	8-4	1	Z	AUGER BORING	Σ	Nickel	97.6		mg/kg	4
	A82-551	8-4	1	Z	AUGER BORING	Σ	Solenium	0.19	CWND	mg/kg	0.5
	AB2-551	8-4	1	Z	AUGER BORING	Σ	Silver	0.75	כ	mg/kg	-
	AB2-551	8-4	3-2	Z	AUGER BORING	Σ	Thallium	=	BN	mg/kg	-
	AB2-581	B-4	3-2	Z	AUGER BORING	Σ	Zluc	91.1		ma/kn	•

RICKENBACKERANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

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A84-SS2	B-8	8-10	Z	AUGER BORING	>	Acetone	250	C	na/ka	Ē
AB4-552	B-6	0-10	Z	AUGER BORING	>	Ellivibenzone	20	3	10 /kg	3 "
AB4-552	8-8	0-10	Z	AUGER BORING	>	m/p-Xylono	96		ug/ka	
AB4-SS2	B-6	0-10	Z	AUGER BORING	>	o-Xylana	15		מט/גם	
AB4-552	9-8	0-10	Z	AUGER BORING	SV	Naphilialone	000		Da/kg	330
AB4-552	B-8	0-10	Z	AUGER BORING	SV	Phonanthrene	18	7	un/ku	200
AB4-552	B-6	0-10	Z	AUGER BORING	SV	Fluoranthene	8	7	un/ka	000
AB4-552	0-0	0-10	Z	AUGER BORING	S	Pyrana	8	7	un/ka	86
AB4-882	B-6	0-10	Z	AUGER BORING	Σ	Antimony	10	ON D	ma/kg	3
AB4-552	8-6	0-10	Z	AUGER BORING	Σ	Arsonic	16,3	2	mo/kg	, –
AB4-SS2	9-8	0-10	Z	AUGER BORING	Σ	Boryllium	0.79		ma/ka	. 50
AB4-SS2	8-6	0-10	z	AUGER BORING	×	Cadmlum	0.45	63	ma/ka	5.0
AB4-552	0-0	0-10	Z	AUGER BORING	Σ	Chrombm	10.0	l	ma/ka	-
AB4-552	B-0	0-10	Z	AUGER BORING	Σ	Coppor	20.4		ma/kg	2.5
AB4-SS2	B-0	01-0	z	AUGER BORING	Σ	Load	20,3	•	ma/ka	0.0
AB4-SS2	B-6	0-10	Z	AUGER BORING	Σ	Morcury.	0,050	2	mo/ko	1.0
A84-SS2	ī	0-10	z ,	AUGER BORING	Σ	Nickol	27.0		ma/kg	4
AB4 - 552	9-6	0-10	Z	AUGER BORING	2	Solonium	0.35	BNWJ	mg/kg	0.5
AB4-882	01.0	9-10	z	AUGER BORING	Σ	Silvar	0.74	ס	ma/ka	_
A84-SS2		0-10	z	AUGER BORING	Σ	Thaillum	0.46	Z B	ma/ka	-
-552	8-8	9-10	Z	AUGER BORING	¥	Zinc	90.3		mg/kg	2
AB5-581	C-2	3-5	Z	AUGER BORING	>	SE ITA IOV 114	Ş			
AB5-551	C-2	3-5	Z	AUGER BORING	S	ALL SEMI-VOLATILES	2 5	2 2	00/80	¥ ×
AB5-551	C-2	3-5	Z	AUGER BORING	×	Antlmony	4	UNB	מא/אמ	¥ 4
AB5-551	C-2	3-5	Z.	AUGER BORING	Σ	Arsonic	13.9	2	ma/ka	
AB5-551	C-2	3-5	z	AUGER BORING	×	Baryllium	0.67	!	a /kg	· v
AB5-551	C-2	3-2	Z	AUGER BORING	\$	Cadmium	0.45	æ	ma/ka	500
AB5-SS1	C-2	3-5	Z	AUGER BORING	×	Chrombm	16.5)	ma/ka	; -
AB5-551	C-2	3-5	z	AUGER BORING	×	Coppor	29.4		mo/ka	- 25
AB5-551	C-2	3-5	Z	AUGER BORING	Σ	Lead	16.1		mo/ka	
AB5-551	C-2	1	Z	AUGER BORING	Σ	Morcury	0.050	ס	ma/ka	0.0
AB5-551	C-2	t	Z	AUGER BORING	Σ.	Nickol	31.1		ma/ka	~
AB5-551	C-2	1	Z	AUGER BORING	Σ	Solonlum	0.2	BNWJ	mg/kg	0.5
AB5-551	C-2	1	z	AUGER BORING	Σ	Silvar	0.74	ס	mg/kg	-
-881	1	1	Z	AUGER BORING	Σ	Thaillum	0.53			•
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ретеспон имт	¥.₹	×	9	-	0.5	0.5	_	2.5	0.3	0.1	4	0.5	_	-	2	×	X	ø	-	0.5	0.5	-	2.5	0.3	0.1	4	0.5	-	_	- 2
DETECTION												_		_		_	_		_							_		_		_
UNITS	ua/ka	ug/kg	mg/kg	ug/kg	ua/ka	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mo/kg	mg/kg	mg/kg	ma/ko	mg/kg	mg/kg	mo/kg												
	Ϋ́Z	¥ Z	UNN	3	80.	89		-		2		LWNB	ס	BNU		¥ Z	¥Z	RNU	3	89				•	כ		BNWJ	כ	BNC	
RESULTS QUALIFIER	Q	QN	4.5	14.4	0.33	0.22	11.6	27.1	15.4	0.057	24.0	0,55	0.74	0.38	02.0	Q.	2	4.5	22.2	0.33	3.3	16.5	16.9	19.5	0.050	25.2	0.29	0.73	00.0	03.1
ANALYSIS	ALLVOLATILES	ALL SEMI-VOLATILES	Antimony	Arsonic	Boryllum	Cadmium	Chromium	Copper	Load	Marcury	Nickel	Salonium	Silvar	Thaillum	Zinc	ALLVOLATILES	ALL SEMI-VOLATILES	Antimony	Arsonic	Baryllium	Cadmium	Chrombm	Copper	Load	Morcury	Nickel	Solonlum	Silver	Thallium	Zinc
CATEGORY	>	SV	Σ	Σ	Σ	Σ	Σ	×	Σ	Σ	Σ	Σ	Σ	Σ	Σ	>	SV	Σ	×	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	×	¥
DESCRIPTION	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING
IN/OUT	Z	Z	Z	Z	Z	z	Z	z	Z	Z	z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z.	Z	Z	Z	z	Z
ОЕРТН	0-10	0-10	8-10	9-10	0-10	0-10	8-10	8-10	8-10	0-10	0-10	8-10	0-10	8-10	8-10	3-2	3-2	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-2	1	3-5
GRID #	C-2	C-2	C-2	C-2	C-2	C-2	C-2	C-2	C-2	C-2	C-2	C-2	C-2	C-2	C-2	C-4	0 -4	0-1	0	C-4	V-7	0-1	710	7 -0	C-4	C-4	0 - V	7-0	V-0	C-4
SAMPLE	AB5-552	A85-552	ABS-552	AB5-SS2	AB5-552	AB5-552	AB5-SS2	ABS-SS2	AB5-552	AB5-552	AB5-552	ABS-552	AB5-552	AB5-552	AB5-SS2	AB6-551	AB6-551	A86-551	AB6-551	A86-551	AB6-551	AB6-551	ABG-521	AB6-551	AB6-551	AB6-551	AB6-551	A86-551	AB6-551	A86-551

RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

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UNITS	ug/kg	ug/kg	mg/kg	mg/kg	աշ/kg	mg/kg	mg/kg	mg/kg	mg/kg	ug/kg	ng/kg	mg/kg																		
4 Marie 1-	7	A'N	CNR	3	>					כ		BNWJ	>	200		6	¥ Z	RND	2		æ				5		CNO CNO	5	BAL	
RESULTS QUAUFIER	_	ON.	9.6	20.3	0.22	0.49	12.4	25.7	14.0	90'0	20.3	0.53	0.64	0.39	05.7	8	Q	5.1	10.2	0.63	0.30	17.0	30.0	17.6	0.061	35.0	0.5	0.03	0.73	7.66
FOR	Benzone	ALL SEMI-VOLATILES	Antimony	Arsonic	Berylllum	Cadmium	Chrombin	Coppor	Load	Morcury	Nickol	Solonlum	Silvar	Thaillurn	Zinc	Methylene Chloride	ALL SEMI-VOLATILES	Antimony	Arsonic	Boryllum	Cadmlum	Clirombin	Copper	Lead	Morcury	Nickel	Solenlum	Silvar	Thalllum	Zinc
сатевову	>	S	Σ	Z	Σ	Z	Σ	×	×	Σ	¥	×	Σ	Σ	¥	>	S	Σ	×	×	×	Σ	Σ	×	Σ	×	×	Σ	×	M
DESCRIPTION C	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	- AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING
IN/OUT	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	z	Z	TUO	DUO	TUO	TUO	TUO	TUO	TUO	TUO	DOUT	DOUT	DOUT	DOUT	TUO	DOUT	OUT
DEPTH	0-10	0-10	8-10	8-10	0-10	0-10	0-10	8-10	8-10	0 - 10	0-10	8-10	0-10	0-10	0-10	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5
GRID .	0-4	0-14	C-4	C-4	4-0	C-4	0-1	0-1	C-4	C-4	C-4	C-4	4-0	Q-7	C-4	C-7	C-7	C-7	C-7	C-7	C-7	C-7	C-7	C-7	C-7	C-7	C-7	C-7	C-7	C-7
SAMPLE #	A86-552	AB6-552	AB6-552	AB6-552	AB6-552	AB6-552	AB6-552	A86-552	AB6-SS2	AB6-552	AB6-552	AB6-552	AB6-552	AB6-552	AB6-552	AB7-551	AB7-551	AB7-551	AB7-551	AB7-551	AB7-551	AB7-551	AB7-551	AB7-551	AB7-551	AB7-551	AB7-551	AB7-551	AB7-551	AB7-551

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DETECTION	¥ Z	VV V	9	-	0.5	0.5	-	2.5	0.3	0.1	*	0.5	-	-	2		ž	Ϋ́	9	-	0.5	0.5	-	2.5	0,3	0.1	4	0,5	-	-	7
UNITS	ug/kg	ug/kg	mg/kg	DA/DE	mg/kg		ug/kg	ng/kg	mg/kg	mű/kg	mg/kg																				
quAUFIER	NA	٧X	RNO	2	100					>		CWND	>	TNB BVT			Y Y	Y Y	SNO	3	E					כ		SWNB	>	282	
RESULTS Q	QV	QN.	4.4	21	0.33	0.55	11.5	23.0	16.7	0.057	25.7	0.34	0.72	0.54	65.9		QN	ON.	4.5	16.0	0.45	0.68	13.6	20	15.7	0.057	30	0.25	0.74	0.63	02.2
ANALYSIS FOR R	ALLVOLATILES	ALL SEMI-VOLATILES	Antimony	Arsonic	Boryllium	Cadmlum	Chromhm	Coppor	Prog	Mercury	Nickol	Solonlum	Silvor	Thallium	Zinc		ALLVOLATILES	ALL SEMI-VOLATILES	Antimony	Arsonic	Boryllium	Cadmlum	Chromium	Copper	Prog	Mercury	Nickol	Selonlum	Silver	Thallium	Zinc
CATEGORY	>	SV	×	Σ	×	¥	Σ	Σ	Σ	Σ	Σ	Σ	\$	Σ	×		>	SV	×	Σ	2	≊	Σ	≆	Σ	Σ	2	2	Σ	Σ	M
DESCRIPTION	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING		AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING
IN/OUT:	OUT	TUO	TUO	TUO	DOUT	DOUT	TUO	OUT	OUT	DOUT	TUO	TUO	DOOT	DOUT	DOOT		Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	N
рертн	8-10	9-10	0-10	0-10	8-10	0-10	0-10	01-0	0-10	0-10	0-10	0-10	0-10	8-10	0-10		3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5
.GND.	C-7	C-7	C-7	C-7	C-7	C-7	C-7	C-7	C-7	C-7		C-7	C-7	C-7	C-7		D-2	0-2	D-2	0-2	D-2	0-2	D-2	0-2	0-2	0-2	D-2	D-2	D-2	D-2	0-2
SAMPLE	AB7-552	AB7-552	AB7-552	AB7-552	AB7-552	AB7-552	AB7-552	AB7-552	AB7-552	AB7-552	AB7-SS2	AB7-552	AB7-552	AB7-552	AB7-552		AB0-551	AB0-551	ABG-551	AB0-551	A80-551	AB8-551	AB0-551	A80-551	AB0-551	AB8-551	AB8-551	AB0-SS1	AB0-551	AB0-551	AB8-551

RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

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UNITS	ug/kg	ug/kg	mg/kg	mg/kg.	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	DX/BH	mg/kg	mg/kg	DX/DE	mg/kg	mg/kg		ug/kg	UQ/KQ	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mű/kg	па/ка	та/ка	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg
- Carlon L	X X	٧X	HNO	3	89	60				ס		CWNG	5	2			₹:	ď Z	E N	3	8	B			ທ	5		3	D	28	
RESULTS QUALIFIER	Q	QN	25	17.4	0.37	0.37	15.9	30.2	14.9	0.059	22.2	0.35	0.02	0.40	93.7		2	Q N	4.9	20.7	0.49	0.49	16.9	27.7	27.7	0.00	30.6	0.21	0.01	0.40	97.2
ANALYSIS FOR RE	ALL VOLATILES	ALL SEMI-VOLATILES	Antimony	Arsenic	Beryllium	Cadınlum	Chrombm	Coppor	Lead	Mercury	Nickel	Solonium	Silver	Thalllent	Zinc		ALL VOLATILES	ALL SEMI-VOLATILES	Anilmony	Arsenic	Baryllum	Cadmlum	Chromium	Copper	Lead	Marcury	Nicket	Solonlum	Silver	Thalllum	Zing
CATEGORY	>	38	Σ	Σ	Σ	Σ	≱	Σ	×	Z	Σ	Σ	×	Σ	Z		>	SS.	Σ	Σ.	Σ	Σ	Σ	×	×	X	Σ	Z	Z	Σ	×
DESCHIPTION CA	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING		AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	SNIGOR GROUN
IN/OUT	2	2	2	<u> </u>	. Z	Z	<u>z</u>	<u>z</u>	Z	Z	Z	2	Z	<u>z</u>	Z		z	z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	14
DEPTH II	1	1 1		2 5	1 0	01-0	0-10	0-10	0-10	0-10	0-10	0-10	0-10	0-10	0-10		3-6	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	35	3-5	3-5	
•	c	7 6	2 0	7 6	2 6	2 -0	0-2	0-2	0-2	0-2	D-2	0-2	0-2	- 1	0-2		0-4	0-4	0-4	0-4	0-4	0-4	0-4	0-4	0-4	0-4	D-4	0-4	0-4	0-4	
SAMPLE . GRID	200	ABU-552	AB0-334	A60 - 554	A00 - 00 A	AB0-552	AB0-552	AB0 - SS2	ABG-SS2	AB0-552	AB8-552	AB6-552	AB0-552	AB0-552	AB0-552	1	AB9-551	AB9-551	AB9-551	AB9-551	AB9-551	AB9-551	AB9-551	AB9-551	AB9-551	- 1	AB9-551	AB9-551	A89-551	A89-551	

RICKENBACKERANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS – SOIL

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	սն/kg	սն/kg	mg/kg	mű/kű		ug/kg	ug/kg	mg/kg																							
	ž	¥ X	2	₹		E				<u> </u>		- CNO	D	ONB BNC		-	¥Z	¥ Z	RNO	2		E)				5		N N	>	BNS	
	Q	QN	90	16.1	0.64	0.51	17.3	24.9	19.5	0.00	30.7	0.21	0.04	0.74	6.06		2	0	4.7	21	0.50	0.35	17.0	29.4	17.6	90.0	28.6	0.19	0.76	0.58	90.5
	ALL VOLATILES	ALL SEMI-VOLATILES	Antimony	Arsonic	Boryllum	Cadmium	Chromium	Coppor	Load	Marcury	Nickel	Salonlum	Silvar	Thalifum	Zinc		ALL VOLATILES	ALL SEMI-VOLATILES	Antimony	Arsonic	Boryllum	Cadmium	Chromium	Coppor	Lead	Morcury	Nickol	Salonium	Silver	Thalllum	Zinc
_	>	SV	Σ	Σ	Σ	×	Z	Σ	Σ	Σ	×	Σ	X	Σ	Σ		>	S	Σ	Σ	×	Σ	Σ	Σ	×	×	Σ	Σ	Σ	¥	Σ
	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING		AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING
	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	z	Z	Z	Z	Z		Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	2
	0-10	0-10	0-10	0-10	0-10	01-0	0-10	0-10	0-10	01-0	0-10	0-10	9-10	0-10	0-10	:	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5
	0-4	0-4	4-0	0-4	D-4	0-4	0-4	0-4	0-4	4-0	4-0	0-4	D-4	D-4	0-4		0-5	0-5	0-5	0-5	0-5	0-5	0-5	0-5	0-5	0-5	0-5	0-5	0-5	0-5	0-5
	AB9-552	AB9-552	AB9-552	A89-SS2	AB9-SS2	AB9-552	AB9-552	AB9-552	AB9-552	AB9-552	AB9-SS2	AB9-552	AB9-552	AB9-552	A89-882		AB10-SS1	AB10-SS1	AB10-551	AB10-551	AB10-551	AB10-551	AB10-SS1	AB10-551	AB10-SS1	AB10-SS1	AB10-SS1	AB10-551	AB10-SS1	AB10-551	AB10-551

RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

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SAMPLE #	GRID ≰	ОЕРТН	IN/OUT	DESCRIPTION	САТЕВОВУ	FOR	RESULTS QUALIFIER UNITS	UAUFIER	UNITS	UNITS DETECTION
AB12-SS7		21-23	NI	AUGER BORING	>	enelyX-o	9	٦	ug/kg	5
AB12-557	C-5		z	AUGER BORING	>	1,3-Dichlorobonzone	9	٦	ug/kg	S
AB12-557	C-5		Z	AUGER BORING	>	1,2/1,4 - Dichlorobanzone	9		ug/kg	20
AB12-SS7	C-5	21-23	Z	AUGER BORING	SV	ALL SEMI-VOLATILES	Q	¥X	ug/kg	AN
AB12-SS7	C-5	21-23	Z	AUGER BORING	Σ	Antimony	3.9	NS.	mg/kg	9
AB12-SS7	C-5		z	AUGER BORING	Σ	Arsonic	0.0	8	mg/kg	-
AB12-557	C-5		Z	AUGER BORING	×	Boryllium	0.38	8	mg/kg	0.5
AB12-SS7		_	Z	AUGER BORING	¥	Cadmlum	0.10	כ	mg/kg	0.5
AB12-557	C-5		z	AUGER BORING	×	Chrombin	0.5		mg/kg	-
AB12-557	C-5		z	AUGER BORING	¥	Copper	21.7	3	mg/kg	2.5
AB12-SS7	C-5	21-23	Z	AUGER BORING	×	Load	7.7	?	mg/kg	0.3
AB12-557	C-5		Z	AUGER BORING	Z	Marcury	0.064	כ	mg/kg	0.1
AB12-557	C-5		Z	AUGER BORING	Σ	Nickel	14.5		nıg/kg	*
AB12-557	C-5		Z	AUGER BORING	₹	Solonlum	0.28	BW	mg/kg	0.5
AB12-SS7	C-5		Z	AUGER BORING	Σ	Silver	0.63	כ	mg/kg	-
AB12-557		21-23	Z	AUGER BORING	Σ	Thalllum	0.12	BNWJ	mg/kg	-
AB12-557	C-5	21-23	Z	AUGER BORING	Σ	Zinc	32.0	2	ma/ka	2

RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

	_								
C-8	0-10	z	AUGER BORING	>	Bonzene	15000		ug/kg	ĸ
Q-6	8-10	Z	AUGER BORING	>	Ethylbonzone	15000		ng/kg	S
Q-8	0-10	Z	AUGER BORING	>	m/p-Xylene	15000		ug/kg	ľ
0-0	0-10	Z	AUGER BORING	>	o-Xylana	27000		ug/kg	S
C-8	0-10	Z	AUGER BORING	S	Naphthaleno	8		ug/kg	330
0-0	0-10	Z	AUGER BORING	S	2 - Mothylnaphthalone	200		ug/kg	330
C-8	0-10	Z	AUGER BORING	SV	Di-n-Butyl Phthalate	130	7	ug/kg	330
0-6	0-10	z	AUGER BORING	¥	Antimony	3.3	CND	mg/kg	9
0-C	0-10	Z	AUGER BORING	Σ	Arsonic	16.9		mg/kg	-
C-8	0-10	Z	AUGER BORING	Σ	Boryllium	09'0		mg/kg	0.5
0-0	0-10	Z	AUGER BORING	Σ	Cadmlum	0.25	æ	mg/kg	0.5
0-0	0-10	Z	AUGER BORING	×	Chromium	13.9		mg/kg	-
0-6	0-10	Z	AUGER BORING	Σ	Copper	33.5	3	mg/kg	2.5
C - 6	0-10	Z	AUGER BORING	Σ	Load	19.6	2	mg/kg	0.3
0-6	01-0	Z	AUGER BORING	Σ	Morcury	0.057	D	mg/kg	0.1
0-0	0-10	Z	AUGER BORING	Σ	Nickel	31.0		mg/kg	*
D-0	0-10	2	AUGER BORING	ĭ	Salonlum	0.3	BW	mg/kg	0.5
0-8	0-10	2	AUGER BORING	Σ	Silvor	0.55	כ	mg/kg	-
0-0	0-10	Z	AUGER BORING	Σ	Thallium	0.29	BNWJ	mg/kg	-
0-0	0-10	Z	AUGER BORING	¥	Zinc	02.7	₹	mg/kg	2
C-6	21-25	Z	AUGEN BORING	>	Вопзепо	0		ug/kg	¥.
C-0	21-25	Z	AUGER BORING	SV	ALL SEMI-VOLATILES	ON	¥	ug/kg	¥
0-8	21-25	Z	AUGER BORING	Σ	Antimony	3.2	CNU	mg/kg	9
0-0	21-25	Z	AUGER BORING	Σ	Arsenic	61.2		mg/kg	-
0-0	21-25	Z	AUGER BORING	Σ	Beryllium	0.32	Θ.	mg/kg	0.5
0-0	21-25	Z	AUGER BORING	Σ	Cadınlum	0.16	83	mg/kg	0.5
C-8	21-25	Z	AUGER BORING	Σ	Chromium	9.4		mg/kg	-
C-6	21-25	Z	AUGER BORING	×	Copper	46	3	mg/kg	2.5
C-8	21-25	Z	AUGER BORING	Σ	Lead	22.9	?.2	mg/kg	0.3
C-6	21-25	Z	AUGER BORING	X	Morcury	0,055	ס	mg/kg	0.1
0-C	21-25	Z	AUGER BORING	Σ	Nickol	15,3		mg/kg	4
C-6	21-25	Z	AUGER BORING	×	Solonlum	9.0	≯	mg/kg	0.5
C-6	21-25	Z	AUGER BORING	X	Silvar	0.53	ס	mg/kg	_
C-6	21-25	Z	AUGER BORING	×	Thalllum	60.0	MND	mg/kg	_
0-D	21-25	Z	AUGER BORING	₹	Zinc	73.6	2	ma/ka	

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DETECTION	100	S	AN	9	-	0.5	0.5	_	2.5	0.0	0.1	•	0.5					•	٧X	9		0.5	0.5		2.5	0.3	0.1		0.5			
UNITS	טמ/אט	ממ/אמ	ug/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	па/ка	mg/kg		ug/kg	ug/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	ma/ka	ma/kg	ma/ka	ma/ka	ma/kg	ma/ka	ma/kg	mg/kg						
QUAUFIER			٧X	35			כ		3	ž	8		35	>	SON	Z		7	ž	NO.			83		2	-z			BW	2	ANO	3
RESULTS Q	640	250	ON	4.0	14.0	0.73	0.23	10.2	23.7	15.0	0.090	27.2	0.21	0.0	0.11	76.0		*	2	2.6	14.5	0.72	0.13	10.4	19.9	13.2	0.16	30,3	0.31	0,43	0.094	60.0
ANALYSIS FOR	Acetone	Ethylbonzono	ALL SEMI-VOLATILES	Antimony	Arsonic	Beryllium	Cadmium	Chrombin	Coppor	Load	Marcury	Nickol	Solonlum	Silvar	Thaillum	Zinc		Trichloroethone	ALL SEMI-VOLATILES	Antimony	Arsonic	Boryllium	Cadmium	Chromium	Coppor	Load	Marcury	Nickel	Safonium	Silver	Thalllum	Zinc
CATEGORY	>	>	SV	Σ	¥	₹	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	¥		>	S	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	×	×	Σ	Σ	W
DESCRIPTION	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING		AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING	AUGER BORING
IN/OUT	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	z	Z	Z		z	Z	Z	Z	Z	Z	z	Z	z	Z	Z	Z	z	Z	Z
ОЕРТН	13-15	13-15	13-15	13-15	13-15	13-15	13-15	13-15	13-15	13-15	13-15	13-15	13-15	13-15	13-15	13-15		25-27	25-27	25-27	25-27	25-27	25-27	25-27	25-27	25-27	25-27	25-27	25-27	25-27	25-27	25-27
GRID .	0 - 5	0-5	0-5	05	0-5	0-5	0-5	0-5	0-5	0-5	0-5	0-5	0-5	0-5		0-5		0-5	0-5	0-5	0-5	0-5	0-5	0-5	0-5	0-5	05	0-5	D-5	0-5	0-5	0-5
≱ BTAMVS	AB15~553	AB15-553	AB15-553	AB15-553	AB15-553	AB15-553	AB15-553	AB15-553	AB15-553	AB15-SS3	AB15-SS3	AB15-553	AB15-SS3	AB15-553	AB15-553	AB15-553		AB15-550	AB15-550	AB15-550	AB15-550	AB15-550	AB15-550	AB15-550	AB15-550	AB15-550	AB15-550	AB15-550	AB15-550	AB15-550	AB15-550	AB15-550

RICKENBACKER ANG B HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

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MW4-852	8-1	0-10	TUO	MONITORING WELL	>	ALL VOLATILES	QN	٧	ug/kg	NA
MW4-552	B-1	9-10	TUO	MONITORING WELL	S	ALL SEMI-VOLATILES	QN	Y X	ug/kg	A'N
MW4-552	B-1	0-10	DOUT	MONITORING WELL	Σ	Antimony	3.2	CND	mg/kg	9
MW4-552	B-1	0-10	DOUT	MONITORING WELL	Σ	Arsonic	17.4		mg/kg	-
MW4-552	8-1	0-10	DOUT	MONITORNG WELL	Σ	Beryllium	0.39		mg/kg	0.5
MW4-552	B-1	8-10	TUO	MONITORING WELL	Σ	Cadmlum	0.24	m	mg/kg	0.5
MW4-552	8-1	0-10	TUO	MONITORNG WELL	Σ	Chrombm	1.0		mg/kg	-
MW4-552	8-1	0-10	TUO	MONITORING WELL	Σ	Coppor	20.7	₹	mg/kg	2.5
MW4-552	8-1	0-10	TUO	MONITORNG WELL	2	Load	15.7	ž	mg/kg	0.0
MW4-552	8-1	0-10	TUO	MONITORNG WELL	Σ	Mercury	0.057	5	mg/kg	0.1
MW4-552	8-1	0-10	TUO	MONITORNG WELL	Σ	Nickel	24		mg/kg	4
MW4-552	8-1	0-10	DOUT	MONITORNG WELL	Σ	Solonlum	0.10	BW	mg/kg	0.5
MW4-552	B1	01-9	TUO	MONITORNG WELL	×	Silvor	0.52	D	mg/kg	-
MW4-552	8-1	0-10	TUO	MONITORNG WELL	×	Thallium	0.34	DWWJ	mg/kg	_
MW4-552	8-1	0-10	OUT	MONITORNG WELL	×	Zinc	10	2	mg/kg	2
MW4-553	0-1	13-15	DOUT	MONITORNG WELL	>	ALL VOLATILES	ON O	¥ Z	ug/kg	Y Y
MW4-553	8-1	13-15	TUO	MONITORING WELL	S	ALL SEMI-VOLATILES	QN	YZ.	ug/kg	Y.
MW4-553	B-1	13-15	OUT	MONITORNG WELL	Σ	Antimony	3.4	25	mg/kg	9
MW4-553	B-1	13-15	OUT	MONITORING WELL	Σ	Arsenic	9.1		mg/kg	_
MW4-553	8-1	13-15	TUO	MONITORING WELL	Σ	Boryllium	0.43		mg/kg	0.5
MW4-553	8-1	13-15	OUT	MONITORING WELL	Σ	Cadmlum	0.60		mg/kg	0.5
MW4-553	8-1	13-15	OUT	MONITORING WELL	×	Chromium	11.2		mg/kg	_
MW4-553	B-1	13-15	TUO	MONITORING WELL	Σ	Coppor	19.7	7	mg/kg	2.5
MW4-553	8-1	13-15	DOUT	MONITORING WELL	Σ	Lead	15.1	7.2	mg/kg	0.3
MW4-883	8-1	13-15	DOCT	MONITORING WELL	Σ	Mercury	0.057	5	mg/kg	0.1
MW4-553	8-1	13-15	DOOT	MONITORING WELL	Σ	Nicko	25.2		mg/kg	•
MW4-553	0-1	13-15	TUO	MONITORING WELL	Σ	Salonlum	0.57	≥	mg/kg	0.5
MW4-553	8-1	13-15	TUO	MONITORING WELL	Σ	Silvor	0.56	D	mg/kg	•
MW4-553	8-1	13-15	TUO	MONITORING WELL	Σ	Thallium	0.27	CWNB	mg/kg	
MW4-553	8-1	13-15	DOC	MONITORING WELL	*	Zine	- 10	2	17/Va	. •

MONITORING WELL	ם אפרר ^	ALL VOLATILES	Q.	¥	ממ/אם	ď
MONITORING WELL	IG WELL SV	ALL SEMI-VOLATILES	QN	AN AN	UQ/kg	Y.
MONITORING WELL	NG WELL	Antimony	4.5	UNB	ing/kg	9
MONITOR		Arsonic	26	3	mg/kg	-
MONITOR		Beryllium	0.57		mg/kg	0.5
MONITOR		Cadmlum	0.23	Ð	mg/kg	0.5
O NOW	MONITORING WELL	Ciromum	16.5	•	ing/kg	- 6
MONITO		Load	10.0	SS	Ing/kg	0.0
MONITO	MONITORING WELL	Morcury	0.055	D	ing/kg	0.1
MONITO		Nickol	20	•	mg/kg	4
MONITO		Solonlum	0.51	7+2	mg/kg.	0.5
MONTO		Silvar	0.75	ס	m0/kg	-
MONITO		Thalllum	0.31	Ð	mg/kg	-
MONITOR	MONITORING WELL M	Zinc	05		mg/kg	2
MONTO		Ethylbonzona	7400		ug/kg	ß
MONITO		m/p-Xylana	1900		ug/kg	S
		o-Xylana	2000	•	βη/βη	S
	MONITORING WELL SV	Naphilialone	8 9	7	ug/kg	330
MONITO		Antimony	2 4	BNI	Da/gn	3
MONITO		Arsonic	0.3	2	Dayka Dayka	-
MONITO	MONITORING WELL	Boryllium	0.32	69	mg/kg	0.5
MONITO		Cadmium	0.21	8	mg/kg	0.5
MONITO		Chramkm	0.0		па/ка	
MOMITO		Сорры	57.4	•	mg/kg	2.5
MONITO		Load	17.6	₹	mg/kg	0.0
MONITO		Morcury	0,050	2	mg/kg	0.1
MONITO	MONITORING WELL	Nickol	13.7		mg/kg	4
MONITO		Salonlum	0.18	BNSJ	mg/kg	0.5
MONIT	MONITORING WELL	Silvar	0.7	כ	mg/kg	-
MONITO		Thalllum	0.050	כ	mg/kg	
MONIT	MONITORING WELL	Zinc	65.0		mg/kg	2

	A X	¥Z	9	-	0.5	0.5	-	2.5	0.3	<u>.</u>	•	0.5	-	-	~	•	2	S	2	2	3	ž	9	-	0.5	0.5	-	2.5	0.3	٥.	*	0.5	_	_	R
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	ug/kg	ug/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	מא/מתו	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	1	Dx/Bn	ug/kg	ug/kg	ug/kg	ug/kg	ug/kg	mg/kg	mg/kg	mg/kg	mg/kg	ng/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	та/ка	mg/kg
	¥ X	¥ X	UNR	2		80		•	NSV	>		7+2	כ	8				7	٥		7	¥ Z	RND	7		8		•	NS.	כ		NS	>	3	
	ON	2	4.4	19.6	0.54	0.33	10	29.9	10.7	0.057	20.0	62.0	0.72	0.29	02.0		Sc.	N	1000	40	-	ON	3.9	11.3	0.49	0.5	14.5	21.6	15.1	0.057	20.5	06.0	0.65	0.070	72.0
	ALL VOLATILES	ALL SEMI-VOLATILES	Antimony	Arsenic	Boryllium	Cadınlum	Chrombm	Coppor	Load	Morcury	Nickol	Solonium	Silver	Thallfurn	Zinc		Vinyi Chlorida	1,1 - Dichloroethene	trans-12-Dichloroethene	Trichloroationa	Toluene	ALL SEMI-VOLATILES	Antlmony	Arsenic	Boryllium	Cadmium	Chromkm	Copper	Load	Mercury	Nickel	Selonlum	Silver	Thattium	Zinc
	>	S	Σ	Σ	×	×	×	Z	Σ	Σ	Σ	Σ	Σ	Σ	¥		>			>	>	SV	Σ	X	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	*
_	MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITOPING WELL	MONITORING WELL	MOMITORING WELL	MONITORING WELL	MONITORING WELL		MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITORING WELL	MONTORING WELL	MONITORING WELL							
_	TUO	TUO	DOUT	TUO	TUO	TUO	DOOT	DOOT	OUT	OUT	DOUT	100	DOCT	DOUT	OUT	1	100	OUT	OUT	OUT	DOUT	DOUT	OUT	OUT	OUT	DOCT	DOUT	DOUT	OUT	OUT	OUT	OUT	OUT	DO	TUO
	0-10	01-0	0 - 10	8-10	0-10	0-10	01-0-	01-0	0-10	0-10	0-10	01-0	0-10	0-10	0 - 10	;	13-15	13-15	13-15	13-15	13-15	13-15	13-15	13-15	13-15	13-15	13-15	. 13-15	13-15	13-15	13-15	13-15	13-15	13-15	13-15
_	E-5	E-5	E-5	E-5	E - 5	E-5	R I	E-5	1	0 1 1	R-5	E-5	E-5	E-5	E - 5	E-5	E-5	E-5	R-5	R 1 S	S 1	E-5													
	MW6-552	MW6-552	MW6-552	MW6-552	MW6-552	MW6-552	MW6-552	MW6-552	MW6-552	MW6-552	MW6-552	MW6-552	MW6-552	MW6-552	MW6-552		MW6-553	MW6-553	WW6-553	MW6-553	MW6-553	MW6-553	MW6-553	MW6-553	MW6-553	MW6-553	MW6-553	MW6-553	MW6-553	MW6-553	MW6-553	MW6-553	MW6-553	MW6-553	WW6-553

RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

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S DETECTION	ĸ	50	22	vo	S	ž	9	-	0.5	0.5	-	2.5	0.3	0.1	•	0.5	-	-	2		a	r.	× '	۰ م	- 6	2.0	-	2.5	0,3	0.1	*	0.5	-	-	2
UNITS	ua/ka	un/ka	UQ/kg	ug/kg	UQ/KG	ug/kg	ma/ka	mg/kg	ma/ka	mg/kg	mg/kg	ma/kg	mg/kg	ma/kg	mg/kg	ma/ka	ma/ka	ma/kg	mg/kg	-	ממ/אם	na/ka	ug/kg	Bx/gm	mo/kg	ma/ka	ma/ka	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg
QUALIFIER UNITS	٦					¥2	CNO	2		80		•	282	כ		CNU	5	BW		·	•	7	Y S	2 2	2	5	1	•	NSZ	÷		BN	כ	8	
RESULTS C	98	2100	006	1000	1200	QN	4.3	15.7	0.65	0.22	19	26.4	19.2	0.059	25.7	0.19	0.71	0.27	79.9		9	4	0 v	0 6	0.75	0.24	22.4	25.3	25.2	0.063	31.6	0.32	0.03	0.23	9.60
ANALYSIS FOR	1,1,1-Trichloroethane	Bonzono	Elliylbenzono	m/p-Xylono	o-Xylana	ALL SEMI-VOLATILES	Antimony	Arsonic	Beryllium	Cadmlum	Chromium	Copper	Load	Morcury	Nickol	Selentum	Silvor	Thallium	Zinc		011021100	ouonio1	ALL SEMI - VOLATILES	Accord	Boryllium	Cadinlum	Chromkum	Copper	Lead	Marcury	Nickel	Solonium	Silvor	Thalilum	Zinc
САТЕВОПУ	>	>	>	>	>	SV	×	×	Σ	Σ	Σ	×	Σ	×	Σ	Σ	×	Σ	×	>	> >	> ?	> ×	₹ ≥	Σ	×	×	×	2	×	Σ	¥	Σ	Z	Σ
DESCRIPTION	MONITORING WELL	MONITORNG WELL	MONITORNG WELL	MONITORNG WELL	MONITORNG WELL	MONITORNG WELL	MOMTORNG WELL	MONITORNG WELL	MONITORNG WELL	MONITORNG WELL	MONITORNG WELL	MONITORING WELL	MONITORING WELL	MONITORNO WELL	MONITORNG WELL	MONITORNG WELL	MONITORING WELL	MONITORNG WELL	MONITORING WELL	LEW GNECTION	MONITORING WELL	MONTONIO WELL	MONITORNG WELL	MONITORNG WELL	MONITORING WELL	MONITORNG WELL	MONITORING WELL	MONITOPING WELL	MONITORING WELL	MONITORING WELL					
IN/OUT	TUO	DOOT	OUT	DOUT	TUO	DOUT	TUO	TUO	DOUT	DOUT	TUO	OUT	TUO	OUT	DOUT	DOUT	DOUT	TUO	OUT	170	1	5 5	100	TOO	OUT	DOUT	OUT	DO	DOUT	TUO	DOO	OCT.	OUT	100	100
DEPTH IN/OU	8-10	8-10	8-10	8-10	8-10	8-10	8-10	8-10	8-10	8-10	8-10	0-10	01-0	0-10	0-10	0 - 10	0-10	0-10	0-10	13-15	13-15		13-15	13-15	13-15	13-15	13-15	13-15	13-15	13-15	13-15	13-15	13-15	13-15	13-15
GRID	D-8	D-8	D-8	0-0	0-0	0-0	0 - 0	0-0	D-0	0-0	0-8	0-0	D-8	0-0	0-0	0-0	0-0	D-0	0-0	0-0	4	0 6	8-0	9-0	0-8	0-0	0-0	9-0	0-8	0-8	0-0	1	1	1	0-0
SAMPLE /	MW7-882	MW7-552	MW7-552	MW7-552	MW7-552	MW7-552	MW7-552	MW7-552	MW7-552	MW7-552	MW7-552	MW7-552	MW7-552	MW7-552	MW7-552	MW7-552	MW7-552	MW7-552	MW7-552	MW7-553	MW7-SS	MW7-563	MW7-553	MW7-553	MW7-553	MW7-553	MW7-553	MW7-553	WW7-553	MW7-553	MW7-553	WW7-553	CSS-/MW	Coo-/ww	COO- LAW
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RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

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SAMPLE / GRID /										
MW0-552	C-10	0-10	TUO	MONITORING WELL	>	ALLVOLATILES	S	×	ug/kg	×
MW0-552	C-10	01-0	TUO	MOMTORING WELL	SV	ALL SEMI-VOLATILES	Q	Y Y	ug/kg	¥Z
MW0-552	C-10	0-10	TUO	MONITORNG WELL	Σ	Antimony	4.5	UNB	mg/kg	9
MW0-552	C-10	0-10	TUO	MONITORING WELL	Σ	Arsenic	23	3	mg/kg	-
MW0-552	C-10	0-10	DOUT	MONITORING WELL	Σ	Boryllum	0.50		mg/kg	0.5
MW0-552	0-10	0-10	TUO	MONITORNO WELL	Σ	Cadmluin	0.22	83	mg/kg	0.5
MW8-552	C-10	01-0	OUT	MONITORING WELL	×	Chromhm	15.0		mg/kg	-
MW0-552	C-10	0-10	OUT	MONITORING WELL	Σ	Copper	25.7	•	то/ка	2.5
MW0-552	C-10	0-10	DOUT	MONITORING WELL	Σ	Load	15.4	7SN	mg/kg	0.0
MW0-552	C-10	0-10	TUO	MOMITORING WELL	Σ	Morcury	0.057	כ	mg/kg	0.1
MW8-552	C-10	0-10	OUT	MONITORNG WELL	Σ	Nickel	32.1		mg/kg	~
MW0-552	C- 10	0-10	DOUT	MONITORING WELL	Σ	Solonium	1.7	7SN	mg/kg	0.5
MW0-552	0-10	0-10	OUT	MONITORING WELL	Σ	Silvar	0.73	5	mg/kg	-
MW0-552	C-10	0-10	DOUT	MONITORING WELL	Σ	Thallium	0.090	MΩ	mg/kg	-
MW6-552	C-10	0-10	OUT	MONITORING WELL	×	Zinc	02.6		mg/kg	2
MW0-553	C-10	13-15	OUT	MONITORNG WELL	>	Bonzono	8	7	ug/kg	5
MW0-SS3	C-10	13-15	OUT	MONITORNG WELL	S	ALL SEMI-VOLATILES	ON	Y X	սց/kg	¥ X
MW0-553	C-10	13-15	OUT	MONITORING WELL	×	Antimony	S	UNB	mg/kg	9
MW6-SS3	C-10	13-15	DOUT	MONITORING WELL	×	Arsonic	5.2	Zig.	mg/kg	-
MW0-553	C-10	13-15	DOUT	MONITORNG WELL	Σ	Boryllium	0.29	כ	mg/kg	0.5
MW0-553	C-10	13-15	DOCT	MONITORING WELL	×	Cadmlum	0.24	כ	mg/kg	0.5
MW0-553	C-10	13-15	OUT	MONITORING WELL	Σ	Chrombin	6.1		mg/kg	-
MW8-553	C-10	13-15	DOUT	MONITORING WELL	Σ	Copper	12	•	mg/kg	2.5
MW0-553	C-10	13-15	DOUT	MONITORING WELL	Σ	Lead	12.1	NSC	mg/kg	0.3
MW0-553	C-10	13-15	DOUT	MONITORNG WELL	Σ	Morcury	0.063	ס	mg/kg	0.1
MW8-SS3	C-10	13-15	TUO	MONITORING WELL	¥	Nickol	13.1		mg/kg	4
MW8-553	C-10	13-15	DOUT	MONITORING WELL	Σ	Salonlum	0.0	NST	mg/kg	0.5
MW8-553	C-10	13-15	DOUT	MONITORING WELL	Σ	Silver	0.03	ס	mg/kg	-
MW0-553	C-10	13-15	TUO	MONITORING WELL	×	Thalllum	0.003)	mg/kg	-
MW6-553	C-10	13-15	OUT	MONITORING WELL	×	Zinc	59.5		mg/kg	2

RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

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MW9-SS2	A-4	0-10	OUT	MONITORING WELL	>	ALL VOLATILES	S	NA	ug/kg	AN.
MW9-552	A-4	8-10	DOUT	MONITORING WELL	SV	ALL SEMI-VOLATILES	QN	A'N	ug/kg	A Z
MW9-552	A-4	0-10	OUT	MONITORING WELL	Σ	Antlmony	4.2	RND	mg/kg	9
MW9-552	A-4	01-0	DOOT	MONITORING WELL	×	Arsenic	16.2	3	mg/kg	-
MW9-552	A-4	0-10	DOUT	MONITORING WELL	Σ	Boryllium	0.62		mg/kg	0.5
MW9-552	A-4	0-10	TUO	MONITORING WELL	×	Cadmlum	0.31	83	mg/kg	0.5
MW9-552	A-4	0-10	DOUT	MONITORING WELL	×	Chromkim	13.1		mg/kg	-
MW9-552	A-4	0-10	DOUT	MONITORING WELL	Σ	Copper	24.6		mg/kg	2.5
MW9-552	A-4	0-10	TUO	MONITORING WELL	×	Load	13.0	3	mg/kg	0.3
MW9-552	A-4	0-10	DOUT	MOMTORING WELL	Σ	Morcury	0.050	>	mg/kg	0.1
MW9-552	A-4	8-10	OUT	MONITORING WELL	Σ	Nickel	31		mg/kg	4
MW9-552	A-4	0-10	DOOT	MONITORING WELL	×	Solonlum	0.40	BS	mg/kg	0.5
MW9-552	A-4	8-10	DOUT	MONITORING WELL	Σ	Silver	0.60	>	mg/kg	-
MW9-552	A-4	0-10	OUT	MONTORING WELL	Σ	Thalilum	4.0	8	mg/kg	-
MW9-552	Y-4	0-10	OUT	MONITORING WELL	¥	Zinc	04.7		mg/kg	2
MW9-553	A-4	13-15	TOO	MONITORING WELL	>	ALLVOLATILES	Q.	₹ Z	ug/kg	Y.
MW9-553	A-4	13-15	DOUT	MONITORING WELL	SV	ALL SEMI - VOLATILES	Q	¥Z	ug/kg	4 Z
MW9-553	A-4	13-15	DOUT	MONITORING WELL	Σ	Antlmony	4.2	RND	mg/kg	9
MW9-553	A-4	13-15	DOUT	MONITORING WELL	Σ	Arsenic	22.2	7	та/ка	
MW9-553	A-4	13-15	DOOT	MONITORING WELL	Σ	Beryllium	0.72		mg/kg	0.5
MW9-553	A-4	13-15	OUT	MONITORING WELL	Σ	Cadmlum	0.21	80	mg/kg	0.5
MW9-553	A-4	13-15	DOUT	MONITORNG WELL	Z	Chrombin	13.6		mg/kg	-
MW9-5S3	Y-4	13-15	DOUT	MONITORING WELL	Σ	Copper	24.9		mg/kg	2.5
MW9-553	A-4	13-15	TUO	MONITORING WELL	2	Lead .	18.4	2	mg/kg	0.3
MW9-SS3	A-4	13-15	TUO	MONITORING WELL	Σ	Morcury	0.062	ס	mg/kg	0.1
MW9-553	AA	13-15	TUO	MONITORING WELL	Σ	Nickel	33		mg/kg	4
KW9-553	A-4	13-15	DOUT	MONITORING WELL	×	Selonium	0.21	3	mg/kg	0.5
MW9-553	A-4	13-15	TUO	MONITORING WELL	Σ	Sliver	09.0	2	mg/kg	-
CSS-6MW	A-4	13-15	OUT	MONITORING WELL	Σ	Thalllum	0.49	m	mg/kg	-
MW9-553	A-4	13-15	OUT	MONITORING WELL	W	Zinc	07.1		mg/kg	2

RICKENBACKERANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

MW10-SS1 A-0 J0-5 OUT MONITORING WELL M MW10-SS1 A-0 3-5 OUT MONITORING WELL M MW10-SS2 A-0 3-5 OUT MONITORING WE				_	-	
A-0		V ALL VOLATILES	8	¥ X	ug/kg	¥ Z
A-0 3-5 OUT MONTORNG WELL M A-0 3-15 OUT MONTORNG WELL M A-0 13-15 OUT MONTORNG WELL M A-0 13-	MONITORING WELL	V ALL SEMI-VOLATILES	QN	٧X	UQ/kg	¥ Z
A-0 3-5 OUT MONTORNG WELL M A-0 13-15 OUT MONTORNG WELL M A-		M	1.6	7	mg/kg	0
A-8		M	19.0	7	mg/kg	-
A-0 3-5 OUT MONTORNG WELL M A-13-15 OUT MONTORNG WELL M A-		M Boryllum	-		mg/kg	0.5
A-0 3-5 OUT MONTORNG WELL M A-0 13-15 OUT MONTORNG WELL M A-1 13-15 OUT MONTORNG		M Cadmlum	0.37		mg/kg	0.5
A-0 3-5 OUT MONTORNO WELL M A-0 13-15 OUT MONTORNO M A-0		Сһгошһш	20.6	7	mg/kg	-
A-0 3-5 OUT MONTORNO WELL M A-0 13-15 OUT MONTORNO M A		Соррог	49.3		mg/kg	2.5
A-0 3-5 OUT MONTORNG WELL M A-0 13-15 OUT MONTORNG M A-0 13-15 OUT MONTO	77	Load	20.7		mg/kg	0.0
A-0 3-5 OUT MONTORNO WELL M A-0 13-15 OUT MONTORNO WELL SV A-0 13-15 OUT MONTORNO WELL M A-0 13-15 OUT MONTORNO M A-0 13-15 OUT MONT		Mercury	90.0	5	mg/kg	0.1
A-0 3-5 OUT MONTORNG WELL M A-0 3-5 OUT MONTORNG WELL M A-0 3-5 OUT MONTORNG WELL M A-0 13-15 OUT MONTORNG WELL M		MICKO	32.3		ing/kg	7
A-0 3-5 OUT MONTORNG WELL M A-0 3-5 OUT MONTORNG WELL M A-0 13-15 OUT MONTORNG WELL M A-0 13-15 OUT MONTORNG WELL SV ALL SEMI-V A-0 13-15 OUT MONTORNG WELL M A-0 13-15 OUT MONTORNG M A-0 13-15 OUT MONT		M Salonlum	0.51	7	mg/kg	0.5
A-0 3-5 OUT MONTORNG WELL M A-0 13-15 OUT MONTORNG WELL M A-0 13-15 OUT MONTORNG WELL SV ALL SEMI-V A-0 13-15 OUT MONTORNG WELL M A-0 13-15 OUT MONTORNG M A		M	0.44	כ	mg/kg	-
A-0 13-15 OUT MONITORING WELL W ALL V A-0 13-15 OUT MONITORING WELL SV ALL SEMI-V A-0 13-15 OUT MONITORING WELL M A-1 13-15 OUT MONITORING WELL M A-2 13-15 OUT MONITORING WELL M A-3 13-15 OUT MONITORING WELL M A-4 13-15 OUT MONITORING WELL M A-6 13-15 OUT MONITORING WELL M A-7 13-15 OUT MONITORING WELL M A-8 13-15 OUT MONITORING WELL M A-9 13-15 OUT MONITORING WELL M A-1 13-15 OUT MONITORING WELL M A-1 13-15 OUT MONITORING WELL M A-2 13-15 OUT MONITORING WELL M A-3 13-15 OUT MONITORING WELL M A-6 13-15 OUT MONITORING WELL M A-7 13-15 OUT MONITORING WELL M A-8 13-15 OUT MONITORING WELL M A-9 13-15 OUT MONITORING WELL M A-1 13-15 OUT MONITORING M A-1 1		M	0.51	7	mg/kg	-
A-0 13-15 OUT MONITORING WELL SV ALL SEMI-V A-0 13-15 OUT MONITORING WELL SV ALL SEMI-V A-0 13-15 OUT MONITORING WELL M A-1 13-15 OUT MONITORING WELL M A-2 13-15 OUT MONITORING WELL M A-3 13-15 OUT MONITORING WELL M A-1 13-15 OUT MONITORING WELL M A-2 13-15 OUT MONITORING WELL M A-3 13-15 OUT MONITORING WELL M A-4 13-15 OUT MONITORING WELL M A-6 13-15 OUT MONITORING WELL M A-7 13-15 OUT MONITORING WELL M A-8 13-15 OUT MONITORING WELL M A-9 13-15 OUT MONITORING WELL M A-1 13-15 OUT MONITORING WELL M A-1 13-15 OUT MONITORING WELL M A-2 13-15 OUT MONITORING WELL M A-3 13-15 OUT MONITORING WELL M A-6 13-15 OUT MONITORING WELL M A-7 13-15 OUT MONITORING WELL M A-8 13-15 OUT MONITORING WELL M A-9 13-15 OUT MONITORING WELL M A-1 13-15 OUT MONITORING M		M Zinc	109	J.	mg/kg	2
A-0 13-15 OUT MONITORING WELL V ALL V A-0 13-15 OUT MONITORING WELL M A-1 13-15 OUT MONITORING WELL M A-2 13-15 OUT MONITORING WELL M A-1 13-15 OUT MONITORING WELL M A-2 13-15 OUT MONITORING WELL M A-3 13-15 OUT MONITORING WELL M A-4 13-15 OUT MONITORING WELL <t< td=""><td></td><td></td><td></td><td></td><td></td><td></td></t<>						
A-0 13-15 OUT MONTORNG WELL M A-1 13-15 OUT MONTORNG WELL M A-2 13-15 OUT MONTORNG WELL M A-1 13-15 OUT MONTORNG WELL M A-2 13-15 OUT MONTORNG WELL M A-3 13-15 OUT MONTORNG WELL M A-0 13-15 OUT MONTORNG WELL M A-0 13-15 OUT MONTORNG WELL M A-0		V ALL VOLATILES	QN	٧X	ug/ku	¥
A-0 13-15 OUT MONITORING WELL M	MONITORING WELL	SV ALL SEMI-VOLATILES	QN	¥Z	ug/kg	¥Z
A-0 13-15 OUT MONITORNG WELL M		M	2.2	7	mg/kg	9
A-0 13-15 OUT MONTOPING WELL M		M Arsonic	10.0	7	mg/kg	-
A-0 13-15 OUT MONTORNG WELL M A-8 13-15 OUT MONTORNG WELL M A-0 13-15 OUT MONTORNG WELL M		M Boryllium	0.25	_	mg/kg	0.5
A-8 13-15 OUT MONTORNG WELL M A-0 13-15 OUT MONTORNG WELL M		M Cadmlum	0.72		mg/kg	0.5
A-0 13-15 OUT MONITOPING WELL M		M	7.4	7	mg/kg	-
A-0 13-15 OUT MONITOPING WELL M		М	22.6		mg/kg	2.5
A-0 13-15 OUT MONITORING WELL M		M	12.1	7	mg/kg	0.3
A-0 13-15 OUT MONITORNG WELL M		M	90.0	D	mg/kg	0.1
A-0 13-15 OUT MONITOPING WELL M A-0 13-15 OUT MONITOPING WELL M A-0 13-15 OUT MONITOPING WELL M		Mickol	19.7		mg/kg	4
A-0 13-15 OUT MONITORING WELL		Saf	0.54	כ	ш9/ка	0.5
A-8 13-15 OUT MONTORING WELL		M	0.75	5	mg/kg	-
		E.T.	0.54	7	mg/kg	-
MW10-553 A-8 13-15 OUT MONITORING WELL M		M	6.00	7	т9/кд	2

RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

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SAMPLE & GRID .									_	
MW11-SS1	F-7	3-5	TUO	MONITORNG WELL	>	ALL VOLATILES	QN	¥ X	ug/kg	¥ z
MW11-551	F-7	3-5	DOUT	MONITORNG WELL	SV	ALL SEMI-VOLATILES	QN	¥ Z	ug/kg	٧Z
MW11-551	F-7	3-5	TUO	MONITORING WELL	Σ	Antlmony	3.3	7	mg/kg	9
MW11-551	F-7	3-5	DOOT	MONITORING WELL	Σ	Arsonic	17.9	7	mg/kg	-
MW11-551	F-7	3-5	TUO	MONITORING WELL	Σ	Boryllium	0.41		mg/kg	0.5
MW11-551	F-7	3-5	DOUT	MONITORING WELL	¥	Cadmlum	0.83		שמ/אמ	0.5
MW11-551	F-7	3-5	DOOT	MONITORING WELL	≆	Chroinium	8.5	2	mg/kg	-
MW11-551	F-7	3-5	DOUT	MONITORING WELL	Σ	Copper	33.4		mg/kg	2.5
MW11-551	F-7	3-5	TUO	MONITORNO WELL	Σ	Load	16.2	7	mg/kg	0.3
MW11-551	F-7	1	TUO	MONITORING WELL	Σ	Marcury	0.00	ס	mg/kg	0.1
MW11-551	F-7	3-5	TUO	MONITORING WELL	Σ.	Nickol	20.2		mg/kg	*
MW11-551	F-7	3-5	DOUT	MONITORING WELL	Z	Salanlum	0.30	>	mg/kg	0.5
MW11-551	F-7	3-5	DOUT	MONITORING WELL	Z	Silvar	99.0	>	mg/kg	-
MW11-551	F-7	3-5	TUO	MONITORING WELL	Σ	Thallium	0.43	7	mg/kg	-
MW11-551	F-7	3-5	TUO	MONITORING WELL	Z	Zinc	79.0	2	т9/кд	2
CSS-11WM	F-7	13-15	TUO	MONITORING WELL	>	ALL VOLATILES	ON	¥	ug/kg	¥
MW11-553	F-7	13-15	TUO	MONITORING WELL	SS	ALL SEMI-VOLATILES	2	¥ X	UU/kg	¥ Z
MW11-553	F-7	13-15	TUO	MONITORING WELL	Σ	Antimony	2.7	7	mg/kg	9
MW11-553	F-7	13-15	TUO	MONITORING WELL	×	Arsonic	13.4	7	mg/kg	-
MW11-553	F-7	13-15	DOUT	MONITORING WELL	Z	Boryllium	0.41		mg/kg	0.5
MW11-553	F-7	13-15	TUO	MONITORING WELL	Z	Cadmlum	0.59		mg/kg	0.5
MW11-553	F-7	13-15	TUO	MONITORING WELL	Σ	Chromium	11.1	7	mg/kg	-
MW11-553	F-7	13-15	TUO	MONITORING WELL	×	Coppe	22		mg/kg	2.5
. MW11-553	F-7	13-15	TUO	MONITORING WELL	Σ	Load	15.2	7	та/ка	0.3
MW11-553	F-7	13-15	TUO	MONITORING WELL	Σ	Morcury	90.0	5	mg/kg	0.1
MW11-553	F-7	13-15	TUO	MONITORNG WELL	×	Nickel	16.0		mg/kg	₹
MW11-553	F-7	13-15	TUO	MONITORING WELL	X	Sol	0.5	7	mg/kg	0.5
MW11-553	F-7	13-15	DOUT	MONITORNG WELL	Z		0.63	>	mg/kg	-
MW11-553	F-7	13-15	DOOT	MONITORING WELL	≆	Thallium	0.5	7	mg/kg	-
MW11-553	F-7	13-15	OUT	MONITORING WELL	≆	Zinc	70.3	7	mg/kg	2

RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

DETECTION UNITS LIMIT	A N	Y X	9	-	0.5	0.5	-	2.5	0.3	0.1	*	0.5	•	-	2
UNITS	o ajour	24/20	Bay/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	ma/ka						
VAUFIER	42	(A	7	7			7			D		7	כ	2	7
nesults aualfien units	C	2 2	-	20.7	0.64	0.52	15.0	26.7	17.71	00'0	32.1	0.30	0.40	00.00	98.2
N FOR	ATT VOLATIES	ALL SEMI_VOLATILES	Antimony	Arsonic	Baryllum	Cadmium	Chromlim	Copper	Load	Mercury	Nickel	Selenium	Silvar	Thalllum	Zinc
CATEGORY	>	\ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \	∑ ≥	Σ	×	×	×	Σ	×	Σ	×	×	Σ	×	Σ
DESCRIPTION	TEM CNECK MON	MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITORNG WELL	MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITORING WELL
וא/סטד	110	5 5	OUT	OUT	TUO	DOUT	DOO	TUO	OUT	DOUT	TUO	OUT	DOUT	DOUT	OUT
DEPTH	1	ו ו	0 0	3-5	3-5	3 - 5	3-6	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5
GRID #	i L	o o	1 III	E-9	E-9	E-9	B = 8	0 - U	E-0	E-9	E-9	E-9	0 - W	E-9	81 H
SAMPLE #	MW12_cc1	WW.D.F.S.	MW12-551	MW12-551	MW12-551	MW12-551	MW12-551	WW12-551							
		_				_	_		_		_		_	_	_

RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

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SAMPLE #	GRID # DEPTH	GRID € DEPTH IN/OUT	IN/OUT	DESCRIPTION	CATEBORY	ANALYSIS	RESULTS C	QUALIFIER UNITS		DETECTION
MW12-553	E I	13-15	TUO	MONITORING WELL	>	ALLVOLATILES	Q	¥ Z	ug/kg	¥ Z
MW12-553	E-9	13-15	TUO	MONITORNG WELL	SV	Phonanthrono	300	7	ug/kg	300
MW12-553	E-9	13-15	DOUT	MONITORING WELL	SV	Anthracono	02	7	ug/kg	300
MW12-553	E-9	13-15	DOUT	MONITORING WELL	SV	Carbazolo	190	7	ug/kg	380
MW12-553	8-H	13-15	TUO	MONITORING WELL	SV	Fluoranthone	310	7	ng/kg	380
MW12-553	8 1 2	13-15	DOUT	MONITORNG WELL	SV	Pyrana	250	ה	ug/kg	300
MW12-553	E-3	13-15	DOUT	MONITORING WELL	SV	Bonzo(a) Anthracene	110	7	ug/kg	300
MW12-553	B 山	13-15	DOUT	MONITORING WELL	S	Chrysono	130	7	ug/kg	300
MW12-553	о 1 Ш	13-15	DOUT	MONITORNG WELL	SV	Bonzo(b) Fluoranthono	110	7	ug/kg	200
MW12-553	61 11	13-15	DOCT	MONITORNG WELL	SV	Banzo(a)Pyrana	02	7	ug/kg	380
MW12-553	E-9	13-15	TUO	MONITORING WELL	×	Antlinony	2.9	7	mg/kg	9
WW12-553	E-9	13-15	DOUT	MONITORING WELL	×	Arsonic	4	7	mg/kg	-
MW12-553	E-9	13-15	DOUT	MONITORNG WELL	Σ	Berylllum	0.22		mg/kg	0.5
MW12-553	E-9	13-15	TUO	MONITORNG WELL	Σ	Cadmium	76.0		mg/kg	0.5
MW12-553	E-3	13-15	OUT	MONITORING WELL	₹	Chromhm	5.0	2	mg/kg	-
MW12-553	3 1 2	13-15	OUT	MONITORNG WELL	Σ	Coppor	30.5		mg/kg	2.5
MW12-553	E-9	13-15	DOUT	MONITORNO WELL	Σ	Load	=	7	mg/kg	0.3
MW12-853	0 1 1	13-15	DOUT	MONITORNG WELL	X	Mercury	50'0	5	mg/kg	0.1
MW12-553	E-3	13-15	DOUT	MONITORNG WELL	Σ	Nickol	9.7		mg/kg	4
MW12-553	日-8	13-15	DOUT	MONITORNG WELL	Σ	Selonium	0.26	7	mg/kg	0.5
MW12-553	E-9	13-15	DOUT	MONITORNG WELL	×	Silver	0.47	ס	mg/kg	-
MW12-553	E-9	13-15	TUO	MONITORNG WELL	×	Thalllum	0.26	7	mg/kg	-
MW12-553	6 П	13-15	DOOT	MONITORING WELL	×	Zinc	50.1	r	mg/kg	2

RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

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		ng/kg	mg/kg	mű/kű	mg/kg	mg/kg	mg/kg	mg/kg		ug/kg	mg/kg	mg/kg	mg/kg	ту/ка	mg/kg															
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		ON.	2.5	12.6	0.5	0.63	12.4	19.2	23.1	0.00	10.6	00.0	c	0.40	179		Q.	1.7	10.0	0.41	9.0	10.3	27.2	20.8	0.05	15.5	0.5	0.50	0.5	109
	NOT ANALIZED	ALL SEMI-VOLATILES	Antimony	Arsonic	Boryllium	Cadmium	Chrombm	Coppor	Load	Mercury	Nickol	Solonlum	Silvar	Thalllum	Zinc	NOT ANALYZED	ALL SEMI-VOLATILES	Antimony	Arsonic	Boryllium	Cadmlum	Chromium	Copper	Lead	Mercury	Nickol	Solonium	Silver	Thallium	Zinc
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| NOT ANALYZED | ALL SEMI-VOLATILES | Antimony | Arsenic | Beryllium | Cadmlum | Chromhm | Coppor | Load | Mercury

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| | E-5 0-2 OUT SURFACE SOIL SAMPLE V | E-5 0-2 OUT SURFACE SOIL SAMPLE SV ALL SEMI-VOLATILES ND NA UG/KG | E-S 0-2 OUT SURFACE SOIL SAMPLE SV ALL SEMI-VOLATILES ND NA Ug/kg N ALL SEMI-VOLATILES ND NA Ug/kg N ANIMONY 1,9 J mg/kg | E-5 0-2 OUT SURFACE SOIL SAMPLE V NOT ANALYZED ND NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Antimony 1.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Arsonic 10.9 J mg/kg | E-5 0-2 OUT SURFACE SOIL SAMPLE V NOTANALYZED ND NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE SV ALL SEMI-VOLATILES ND NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Arsonic 1,9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Arsonic 10,9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Baryllium 0,39 mg/kg | E-5 0-2 OUT SURFACE SOIL SAMPLE V NOTANALYZED ND NA ug/kg E-5 0-0 OUT SURFACE SOIL SAMPLE SV ALL SEMI-VOLATILES ND NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Arsonic 1,9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Arsonic 10,9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Baryllium 0,39 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cadmlum 0,51 mg/kg | E-5 0-2 OUT SURFACE SOIL SAMPLE V NOT ANALYZED ND NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Antimony 1.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Arsonic 10.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Boryllium 0.39 mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cadmium 0.39 mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cadmium 0.51 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cadmium 0.51 J mg/kg | E-5 0-2 OUT SURFACE SOIL SAMPLE V NOT ANALYZED ND NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Antimony 1.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Arsonlc 10.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Boryllium 0.39 mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cadmlum 0.51 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Chrombum 0.61 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Chrombum 0.61 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Chrombum 0.61 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Coppor 26.6 J mg/kg | E-5 0-2 OUT SURFACE SOIL SAMPLE V NOT ANALYZED ND NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Antimony 1.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Arsonlo 10.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Boryllium 0.39 mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cadmlum 0.51 mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Chrom/Lm 0.51 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Coppor Coppor 20.6 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Coppor Coppor 20.6 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Coppor Co.6 J mg/kg <th>E-5 0-2 OUT SURFACE SOIL SAMPLE V NOT ANALYZED ND NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Antimony 1.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Arsonlo 10.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cadmlum 0.39 mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cadmlum 0.51 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Coppor 20.6 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Coppor 20.6 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Coppor 20.6 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Morcury 0.05 U mg/kg <t< th=""><th>E-5 0-2 0UT SURFACE SOIL SAMPLE V NOTANALYZED ND NA ug/kg E-5 0-2 0UT SURFACE SOIL SAMPLE NA Antimony 1.9 J mg/kg E-5 0-2 0UT SURFACE SOIL SAMPLE M Arsonic 10.9 J mg/kg E-5 0-2 0UT SURFACE SOIL SAMPLE M Arsonic 10.9 J mg/kg E-5 0-2 0UT SURFACE SOIL SAMPLE M Cadmium 0.39 mg/kg E-5 0-2 0UT SURFACE SOIL SAMPLE M Coppor 20.6 J mg/kg E-5 0-2 0UT SURFACE SOIL SAMPLE M Coppor 20.6 J mg/kg E-5 0-2 0UT SURFACE SOIL SAMPLE M Coppor 20.6 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Morcury 0.05 U mg/kg <t< th=""><th>E-5 0-2 0UT SURFACE SOIL SAMPLE V ALL SEMI-VOLATILES ND NA ug/kg E-5 0-2 0UT SURFACE SOIL SAMPLE M Antimony 1.9 J mg/kg E-5 0-2 0UT SURFACE SOIL SAMPLE M Arsenic 10.9 J mg/kg E-5 0-2 0UT SURFACE SOIL SAMPLE M Arsenic 0.39 mg/kg E-5 0-2 0UT SURFACE SOIL SAMPLE M Chrombun 0.51 mg/kg E-5 0-2 0UT SURFACE SOIL SAMPLE M Coppor 28.6 J mg/kg E-5 0-2 0UT SURFACE SOIL SAMPLE M Coppor 22.4 mg/kg E-5 0-2 0UT SURFACE SOIL SAMPLE M Morcury 0.05 U mg/kg E-5 0-2 0UT SURFACE SOIL SAMPLE M Morcury 0.05 U mg/kg E-5 0-2<th>E-S 0-2 0UT SURFACE SOIL SAMPLE V NOTANALYZED ND NA ug/kg E-S 0-2 0UT SURFACE SOIL SAMPLE M Antimony 1.8 J mg/kg E-S 0-2 0UT SURFACE SOIL SAMPLE M Antimony 1.8 J mg/kg E-S 0-2 0UT SURFACE SOIL SAMPLE M Antimony 1.9 J mg/kg E-S 0-2 0UT SURFACE SOIL SAMPLE M Cadmlum 0.51 mg/kg E-S 0-2 0UT SURFACE SOIL SAMPLE M Coppor 22.4 mg/kg E-S 0-2 0UT SURFACE SOIL SAMPLE M Mercury 0.05 J mg/kg E-S 0-2 0UT SURFACE SOIL SAMPLE M Mercury 0.05 U mg/kg E-S 0-2 0UT SURFACE SOIL SAMPLE M Mercury 0.05 U mg/kg E-S</th><th>E-5 0-2 OUT SURFACE SOIL SAMPLE V NOTANALYZED NA
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 Mercury 0.05 U mg/kg E-S</th> <th>E-5 0-2 OUT SURFACE SOIL SAMPLE V NOTANALYZED NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M ALL SEM-VOLATILES NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Arsenic 1.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Arsenic 10.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cadmlum 0.51 mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cadmlum 0.6 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Morcury 0.6 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Morcury 0.05 U mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Morcury 0.05 U mg/kg E-5 0-2</th> <th>E-5 0-2 OUT SURFACE SOIL SAMPLE V ALL SEMI-VOLATILES ND NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Arsonic 1.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Arsonic 10.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Arsonic 0.39 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Chrombum 0.51 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Chrombum 0.61 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Morcury 0.6 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Morcury 0.05 U mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Morcury 0.05 U mg/kg</th> <th>E-5 0-2 OUT SURFACE SOIL SAMPLE V NOTANALYZED NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M ALL SEMI-VOLATILES ND NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Arsenic 1.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cadmium 0.39 mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cappor 22.4 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cappor 22.4 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cappor 22.4 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Morcury 0.05 U mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Morcury 0.05 U mg/kg E-5</th> <th>E-5 0-2 OUT SURFACE SOIL SAMPLE V NOTANALYZED NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Antimony 1.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Assanle 10.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Assanle 10.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Chrombum 0.51 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Chrombum 0.51 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M More Coppor 20.6 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M More More Coppor 20.6 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M More More Coppor 20.0 J mg/kg</th> <th>E-5 0-2 OUT SURFACE SOIL SAMPLE V ALL SEM-VOLATILES ND NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Alsania 1,9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Assonia 10,9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Assonia 0,39 mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cadmium 0,39 mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cappor Cappor 20,6 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cappor 20,6 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M N Load 22,4 mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M N N N N N N</th> | E-S 0-2 0UT SURFACE SOIL SAMPLE V NOTANALYZED ND NA ug/kg E-S 0-2 0UT SURFACE SOIL SAMPLE M Antimony 1.8 J mg/kg E-S 0-2 0UT SURFACE SOIL SAMPLE M Antimony 1.8 J mg/kg E-S 0-2 0UT SURFACE SOIL SAMPLE M Antimony 1.9 J mg/kg E-S 0-2 0UT SURFACE SOIL SAMPLE M Cadmlum 0.51 mg/kg E-S 0-2 0UT SURFACE SOIL SAMPLE M Coppor 22.4 mg/kg E-S 0-2 0UT SURFACE SOIL SAMPLE M Mercury 0.05 J mg/kg E-S 0-2 0UT SURFACE SOIL SAMPLE M Mercury 0.05 U mg/kg E-S 0-2 0UT SURFACE SOIL SAMPLE M Mercury 0.05 U mg/kg E-S | E-5 0-2 OUT SURFACE SOIL SAMPLE V NOTANALYZED NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M ALL SEM-VOLATILES NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Arsenic 1.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Arsenic 10.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cadmlum 0.51 mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cadmlum 0.6 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Morcury 0.6 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Morcury 0.05 U mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Morcury 0.05 U mg/kg E-5 0-2 | E-5 0-2 OUT SURFACE SOIL SAMPLE V ALL SEMI-VOLATILES ND NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Arsonic 1.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Arsonic 10.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Arsonic 0.39 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Chrombum 0.51 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Chrombum 0.61 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Morcury 0.6 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Morcury 0.05 U mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Morcury 0.05 U mg/kg | E-5 0-2 OUT SURFACE SOIL SAMPLE V NOTANALYZED NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M ALL SEMI-VOLATILES ND NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Arsenic 1.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cadmium 0.39 mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cappor 22.4 J mg/kg
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 OUT SURFACE SOIL SAMPLE M Atlantiony 1.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Atlantiony 1.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Atlantiony 1.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Atlantiony 2.3 mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Atlantiony 2.3 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Atlantiony 0.3 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Atlantion 0.43 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Atlantion 0.43 J mg/kg < | E-5 0-2 OUT SURFACE SOIL SAMPLE V ALL SEM-VOLATILES ND NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M ALL SEM-VOLATILES ND NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Assenic 1:9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Assenic 1:9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Chambum 0.51 Mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Mmcruy 0.51 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Mmcruy 0.05 U mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Mmcruy 0.05 U mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Mmcruy 0.05 U mg/kg | E-5 0-2 OUT SURFACE SOIL SAMPLE V ALL SEM-VOLATILES ND NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M ALL SEM-VOLATILES ND NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Assanle 1.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Chrombum 0.39 mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Chrombum 0.31 mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Chrombum 0.31 mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Morcury 0.05 U mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Morcury 0.05 U mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Morcury 0.05 U Mg/kg E-5 | E-5 0-2 OUT SURFACE SOIL SAMPLE V ALL SEMA-VOLATILES ND NA ug/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M ALL SEMA-VOLATILES ND Jmg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M ALL SEMA-VOLATILES ND Jmg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Acsonic 10.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cadrimun 0.51 Jmg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Marcuny 0.05 Jmg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Marcuny 0.05 Jmg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Marcuny 0.05 Jmg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Marcuny 0.05 Jmg/kg E-5 0-2 OUT SURFACE SOIL S | E-5 0-02 OUT SURFACE SOIL SAMPLE V ALL SEMAVIZED ND NA Up/kg E-5 0-02 OUT SURFACE SOIL SAMPLE M ALL SEMAVOLATILES ND NA Up/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Assarid 1.9 J mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cadmium 0.53 mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cadmium 0.53 mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cadmium 0.53 mg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Cappin 0.53 Umg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Mmg/kg 0.53 Umg/kg E-5 0-2 OUT SURFACE SOIL SAMPLE M Mmg/kg 0.54 Umg/kg E-5 0-2 OUT SURFACE SOIL SAM | E-5 0-01 SURFACE SOIL SAMPLE V ALL SEMAVYZED ND NA UIPKB E-5 0-02 00T SURFACE SOIL SAMPLE M ALL SEMAVOLATILES ND NA UIPKB E-5 0-02 00T SURFACE SOIL SAMPLE M Assanle 1.9 J mg/kg E-5 0-02 00T SURFACE SOIL SAMPLE M Cadrium 0.59 J mg/kg E-5 0-02 00T SURFACE SOIL SAMPLE M Cadrium 0.51 J mg/kg E-5 0-02 00T SURFACE SOIL SAMPLE M Cadrium 0.51 J mg/kg E-5 0-02 00T SURFACE SOIL SAMPLE M Moreury 0.05 U mg/kg E-5 0-02 00T SURFACE SOIL SAMPLE M Moreury 0.43 J mg/kg E-5 0-02 00T SURFACE SOIL SAMPLE M NOTANALYZED M Mg/kg < |

RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

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DETECTION LIMIT		350	350	350	350	350	350	9	-	9.0	0.5	-	2.5	0.3	0.1	4	0.5	-	-	2
1		ug/kg	ug/kg	UQ/kg	ug/kg	ug/kg	ug/kg	mg/kg												
VAUFIER		7	2	7	7	7	2	7	7			7			D		2)	7	7
RESULTS QUAUFIER UNITS		310	300	430	200	330	520	1.0	15.0	09.0	1.7	14.9	32.3	7.66	0.05	27.8	0.43	9.0	0.43	441
ANALYSIS FOR	NOT ANALYZED	Phonanthrono	Fluoranthene	Pyrane	Benzo(a) Anthracene	Chrysono	Benzo(b) Fluoranthene	Antimony	Arsonic	Boryllum	Cadmlum	Chromhm	Coppor	Lead	Mercury	Nickol	Salonlum	Silver	Thaillum	Zinc
CATEGORY	>	SV	SV	SV	S	SV	SV	Σ	Σ	Σ	Σ	Z	×	Σ	Σ	Σ	Σ	Σ	Σ	Σ
DESCRIPTION C	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE						
IN/OUT	OUT	TUO	DOUT	OUT	OUT	DOUT	DOUT	DOUT	TUO	OUT	DOUT	OUT	DOUT	DOUT	OUT	TUO	OUT	TUO	TUO	TUO
DEPTH	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2
GNID. ≰	F-3	F-3	F-3	F-3	F-3	H-3	F-3	F-3	F-3	F-3	F-3	F-3	F-3	F-3	F-3	F - 3	F-3	F-3	F-3	F-3
SAMPLE	888	888	828	888	559	888	888	828	888	888	888	889	828	828	888	888	888	888	828	889

21.7 N D N D N D N D N D N D N D N D D D N D D D N D D D N D D D N D D D	Sold T	******	HAND BORING HAND BORING HAND BORING HAND BORING HAND BORING	Z Z Z Z Z Z	0 0 0 0 0 0		1 1 1 1
21.2 7.12 ND NO	2 0 1	: X		ZZZZ	1111		1 1 1
21.7 ON D	∡ ος	ZZZ		ZZZ	0 0 0	7 7 7	
ND 21.7	2	X X		ΣZ	0 - 2	1 1	- 1
Q.	2	Σ		Z	0-2	1	
-							- 1
157	Lead	×	HAND BORING	Z	0-2	C-3	HB1-551
19.7		X		Z	0-2	C-3	HB1-551
16.3	ธ์	¥		Z	0-2	C-3	HB1-551
7		Σ		Z	0-2	C-3	HB1-551
0.70		Σ		Z	0-2	C-3	HB1-551
15		2		Z	1	C-3	HB1-551
Q	₹	Σ		Z		C-3	HB1-551
1100	2-Mallylna	SV		Z	-	C-3	HB1-551
440000		>		Z	0-5	C-3	HB1-551
92000	Mathylane Chloride	>	HAND BORING	Z	0-2	C-3	HB1-551
274	Zinc	₹	SURFACE SOIL SAMPLE	OUT	0-2	E-4	5510
0.47	Thallfum	×	SURFACE SOIL SAMPLE	TUO	0-2	ш	5510
0.64	Slivor	M	SURFACE SOIL SAMPLE	DO	0-2	ш	5510
0.47	Salonium	×	SURFACE SOIL SAMPLE	OUT	0-2	Ш	5510
31.0	Nickol	¥	SURFACE SOIL SAMPLE	OUT	ı	ш	5510
0.05	Marcury	Σ	SURFACE SOIL SAMPLE	DOC	- 1		3510
50.2	Load	Σ	SURFACE SOIL SAMPLE	DOC	1	E-4	5510
34.1	Copper	Σ	SURFACE SOIL SAMPLE	OUT	0-2	H-4-	5510
10.5	Chromium	Σ	SURFACE SOIL SAMPLE	OUT	0-2	E-4	5510
=	Cadmlum	Z	SURFACE SOIL SAMPLE	OUT	0-2	E-4	\$510
0.71	Borylllum	Σ	SURFACE SOIL SAMPLE	DOUT	0-2	E-4	5510
17.4	Arsanic	Σ	SURFACE SOIL SAMPLE	DOUT	1	E-4	5510
1.9	Antimony	Σ	SURFACE SOIL SAMPLE	TUO	1	E-4	2510
10	Indono(12,3-cd)Pyrono	SV	SURFACE SOIL SAMPLE	DOUT	- 1	E-4	5510
150	Bonzo(b) Fluoranthono	SV	SURFACE SOIL SAMPLE	DOUT	- 1	E-4	5510
140	Chrysone	SV	SURFACE SOIL SAMPLE	OUT	ı	E-4	2510
90	Bonzo(a) Anthracono	SV	SURFACE SOIL SAMPLE	DOCT	- 1	E-4	5510
92	Pyrana	SV	SURFACE SOIL SAMPLE	OUT		E-4	5510
220	Fluoranthene	25	SURFACE SOIL SAMPLE	OUT	0-2	E-4	5510
8	Phonanthrono	S	SURFACE SOIL SAMPLE	DOUT	0-2	E-4	2510
	NOT ANALYZED	>	SURFACE SOIL SAMPLE	DOUT	0-2	,	
-						F 4	2510
i	120 220 100 66 140 150 81 17.4 0.71 1.1 18.5 34.1 50.2 0.05 31.0 0.47 0.47 274 0.47 274 0.05 1100 1100 1100 1100 1100 1100 1100	NOT ANALYZED Phonanthrona Fluoranthene Fluoranthene Fluoranthene Fluoranthene Chrysone Bonzo(b) Fluoranthone Indono(12,3-cd) Pyrone Antimony Arsanic Solonium Slivor Thaillium Chromium O Solonium Slivor Thaillium O Thaillium Zinc Antimony Arsanic Arsanic Boryllium O Cadmium O Coppor	NOT ANALYZED Phonanthrona Fluoranthrona Fluoranthrona Bonzo(a)Anthracene Chrysona Bonzo(b)Fluoranthona Indono(12,3-c-c)Pyrona Antimony Arsenic Boryillum Cluod Mercuy Nickol Solonium Solonium Cluod Antimony Arsenic Antimony Arsenic Antimony Arsenic Antimony Arsenic Cadmium Cluod Solonium Cluod Solonium Cluod Solonium Cluod Antimony Arsenic Antimony Arsenic Antimony Arsenic Cadmium Cluomium 11 Cluod Antimony Arsenic Cadmium Cluomium 11 Cluomium Cluomium Cluomium Coppor	SURFACE SOIL SAMPLE SV Bonzo(a)Anthracono SURFACE SOIL SAMPLE SV Bonzo(b)Fluoranthono SURFACE SOIL SAMPLE SV SURFACE SOIL SAMPLE SV SURFACE SOIL SAMPLE M M Arsanlc HAND BORING M Cadmium Cadmium HAND BORING M CACIMIUM C	SURFACE SOIL SAMPLE M M MORIVION HAND BORING M Cadmium T TABILIUM COPPOR TABILIUM Arsanlc CADPOR TABILIUM CAGINIUM M CAMMINIUM M C	SURFACE SOIL SAMPLE SV Phonanthrono 15	OUT SURFACE SOIL SAMPLE SV Fluoraniliania 20

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	0-2	Z	HAND BORING	>	Mathylana Chlorida	30000	4	սց/kը	S
		Z	HAND BORING	>	o-Xylona	300000		ug/ku	3
	0-2	Z	HAND BORING	SV	Naphilhalono	390		ug/kg	330
C-3	0-2	Z	HAND BORING	S	2 - Methylnaphthalene	098		ug/kg	33
C-3	0-2	Z	HAND BORING	*	Antimony	ON	z	mg/kg	9
-2	0-2	Z	HAND BORING	Σ	Arsonic	11,5	z	mg/kg	0.5
<u>د</u>	0-2	Z	HAND BORING	Σ	Boryllium	2	Z	mg/kg	0.1
-3	0-2	Z	HAND BORING	×	Cadmlum	Q	z	mg/kg	-
C-3	0-2	z	HAND BORING	Σ	Chrombin	16.2	z	mg/kg	- •
-3	0-2	Z	HAND BORING	×	Copper	21.9	z	mg/kg	- (
C-3	0-2	Z	HAND BORING	Σ	Lead	10	•	mg/kg	n (
C-3	0-2	Z	HAND BORING	Σ	Morcury	QN		mg/kg	0.1
C-3	0-2	ž	HAND BORING	Σ	Nickol	27.0	z	mg/kg	- 1
C-3	0-2	z	HAND BORING	Σ	Salonlum	Q N		mg/kg	6.0
C-3	0-2	Z	HAND BORING	Σ	Silvar	2		mg/kg	-
C-3	0-2	Z	HAND BORING	×	Thallium	Q		mg/kg	2
- 1		Z	HAND BORING	2	Zinc	02.1	z	mg/kg	
-3	3-5	Z	HAND BORING	>	Elhylbonzono	120000	D	מא/מח	1 0
C-3	3-5	Z	HAND BORING	>	Mathylana Chiorida	130000	<	00/kg	ימי
C-3	3-5	Z	HAND BORING	>	o-Xylana	1900000	D	ng/kg	S.
C-3	3-5	Z	HAND BORING	S	Di-n-butyi Phihalato	9		ng/kg	330
C-3	3-5	Z	HAND BORING	SV	Naphihalono	1600		ng/kg	330
C-3	3-5	Z	HAND BORING	SV	2-Mothylnaphthalono	2000		ng/kg	င္က
C-3	3-5	Z	HAND BORING	Σ	Antimony	QN	z	mg/kg	o
C-3	3-5	Z	HAND BORING	Σ		17.6		mg/kg	0.5
C-3	3-6	Z	HAND BORING	Σ	Baryllium	0.04		mg/kg	0.1
6-3	3-5	Z	HAND BORING	Σ		6.9	g	па/ка	
C-3	35	Z	HAND BORING	×	Chromium	10.4		mg/kg	
C-3	3-5	Z	HAND BORING	¥	Copper	22.7		mg/kg	
C-3	3-5	Z	HAND BORING	Σ	Load	0		mg/kg	
C-3	3-5	Z	HAND BORING	×	~			mg/kg	
C-3	3-5	Z	HAND BORING	Z		~		mg/kg	
C-3	3-5	Z	HAND BORING	Σ	Sol	2		mg/kg	0.5
C-3	3-5	Z	HAND BORING	¥		2		mg/kg	•
C-3	3-5	Z	HAND BORING	Z	Tha	Q :		mg/kg	0.
610	4	141	011000	•	111				

RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS – SOIL

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HB2+3-SS1	0 - 0	0-2	Z	HAND BORING	>	ALL VOLATILES	Q	NA	ug/kg	X X
HB2+3-551	0-c	0-2	Z	HAND BORING	S	Naphthalene	970		ug/kg	330
HB2+3-SS1	0-c	0-2	Z	HAND BORING	S	2-Malhylnaphthalone	1900		og/kg	000
	0-0	0-2	Z	HAND BORING	Σ	Anilmony	QN	z	mg/kg	9
HB2+3-SS1	0-0	0-2	Z	HAND BORING	×	Arsonic	12	z	mg/kg	0.5
HB2+3-551	0-0	0-2	Z	HAND BORING	¥	Boryllium	Q	z	mg/kg	0.1
HB2+3-551	0-0	0-2	Z	HAND BORING	Z	Cadmlum	Q.	z	mg/kg	-
HB2+3-551	0 - 0	0-2	Z	HAND BORING	Σ	Chromium	1.1	z	mg/kg	-
HB2+3-551	0-0	0-2	N	HAND BORING	Σ	Coppor	19.5	z	mg/kg	-
HB2+3-551	0-0	0-2	Z	HAND BORING	Σ	Lead	27		mg/kg	vn
HB2+3-551	0-0	0-2	Z	HAND BORING	Σ	Morcury	Q.		mg/kg	0.1
HB2+3-551	0-0	0-2	Z	HAND BORING	Σ	Nickol	21.4	z	mg/kg	-
HB2+3-551	0-C	0-5	Z	HAND BORING	Σ	Salonium	2	`	mg/kg	0.5
HB2+3-SS1	0 - 0	0-2	Z	HAND BORING	Σ	Silvar	QN		mg/kg	-
HB2+3-551	0-0	0-2	Z	HAND BORING	×	Thallium	QN		mg/kg	10
HB2+3-551	0-0	0-2	N	HAND BORING	×	Zinc	1.98	z	mg/kg	-
HB2-SS2	0-6	0-2	Z	HAND BORING	>	o-Xylene	43	0	ng/kg	20
HB2-552	0-0	0-2	Z	HAND BORING	S	Di-n-Butyl Phthalate	0060		ug/kg	330
HB2-552	0-0	0-2	Z	HAND BORING	S	Naphthalono	2400		ug/kg	330
HB2-552	0-0	0-2	Z	HAND BORING	SS	Phonanthreno	2600		ug/kg	330
HB2-552	0-8	0-2	Z	HAND BORING	S	2-Mathylnaphilhalano	23000		ng/kg	330
HB2-552	0-8	0-2	Z	HAND BORING	*	Antimony	QN	z	mg/kg	9
HB2-552	0-0	0-2	Z	HAND BORING	×	Arsonic	10.5		mg/kg	0.5
HB2-552	9-0	0-2	Z	HAND BORING	×	Baryllium	10.0		mg/kg	0.1
HB2-552	9-0	0-2	Z	HAND BORING	×	Cadmlum	0.2	g	mg/kg	-
H02-552	C-6	0-2	Z	HAND BORING	×	Chrombm	17.7		mg/kg	-
HB2-552	0O	0-2	Z	HAND BORING	×	Copper	24.4		mg/kg	-
H82-552	0-8	0-2	Z	HAND BORING	×	Load	9.0	z	mg/kg	S.
HB2-552	0-0	0-2	Z	HAND BORING	2	Marcury	Q.		mg/kg	0.1
HB2-552	0-8	0-2	Z	HAND BORING	×	Nickel	20.0		mg/kg	-
HB2-552	9-0	0-2	Z	HAND BORING	Σ	Solonlum	2	*	mg/kg	0.5
HB2-552	9-0	0-5	Z	HAND BORING	Σ	Silvar	2		mg/kg	- ;
HB2-552	9-0	0-5	Z	HAND BORING	2	Thallium	Q 2		mg/kg	0
HB2-552	2	0	-	CHECK CHAIL		1				

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HB2-553	9-D	3-5	Z	HAND BORING	>	ALL VOLATILES	02	AN AN	ug/kg	¥
H82-553	C-8	3-5	Z	HAND BORING	S	Di-n-butyi Phihatato	2900		ug/kg	000
HB2-553	9-O	3-5	Z	HAND BORING	SV	Fluorena	8		ug/kg	000
HB2-553	0-0	3-5	Z	HAND BORING	SV	Naphthalono	570		ug/kg	330
HB2-553	0-0	3-5	Z	HAND BORING	S	Phonanthrono	700		ug/kg	000
HB2-583	0-8	3-5	Z	HAND BORING	S	2-Melhylnaphthalene	2000		ug/kg	330
HB2-553	0-0	3-5	Z	HAND BORING	×	Anlimony	Q	z	mg/kg	9
HB2-553	0-6	1	Z	HAND BORING	Σ	Arsonic	7::	z	mg/kg	0.5
HB2-553	0-0	3-5	Z	HAND BORING	Σ	Baryllum	QN	z	mg/kg	0.1
HB2-553	0-0		Z	HAND BORING	Z	Cadmlum	QX	z	mg/kg	-
HB2-553	0-0	3-5	z	HAND BORING	Σ	Chromium	10.1	z	mg/kg	-
HB2-553	0-0	3-5	z	HAND BORING	\$	Coppor	10.2	z	mg/kg	-
HB2-553	0-0	3-5	Z	HAND BORING	Σ	Load	0.2	8	mg/kg	S
HB2-553	0-0	3-5	Z	HAND BORING	Σ	Marcury	Q		mg/kg	0.1
HB2-553	0-6	3-5	Z	HAND BORING	Σ	Nickol	25	z	mg/kg	-
HB2-553	C-6	3-5	Z	HAND BORING	×	Satentum	QN N		mg/kg	9.0
HB2~553	0-0	3-5	Z	HAND BORING	Σ	Silvar	QN		mg/kg	-
HB2-583	0-0	3-5	Z	HAND BORING	Σ	Thalllum	Q		mg/kg	10
HB2-SS3	0-0	3-5	ᆂ	HAND BORING	≥	Zinc	73.4	z	mg/kg	-
HB4-881	6-9	0-2	Z	HAND BORING	>	ALL VOLATILES	QN	¥Z	ug/kg	₹ Z
1	8-3	0-2	Z	HAND BORING	SV	Pyrana	230		ng/kg	330
HB4-881	1	0-2	Z	HAND BORING	Σ	Antimony	ON	z	mg/kg	9
	8-3	0-2	Z	HAND BORING	Σ	Arsonic	7.8	z	mg/kg	0.5
HB4-SS1	8-3	0-5	Z	HAND BORING	×	Boryllium	4.0	8	mg/kg	0.1
HB4-SS1	B-3	0-2	Z	HAND BORING	Σ	Cadmium	OZ.		mg/kg	-
HB4-551	8-3	0-2	Z	HAND BORING	Σ	Chrombm	10	g	mg/kg	-
HB4-SS1	8-3	0-2	z	HAND BORING	Σ	Coppor	56	•	mg/kg	-
HB4-551	8-3	0-2	Z	HAND BORING	Σ	Lead	27	•	mg/kg	S
HB4-551	B-3	0-2	Z	HAND BORING	Σ	Morcury	QN	ס	mg/kg	0.1
HB4-551	B-3	0-2	Z	HAND BORING	Σ	Nickol	23	• .	mg/kg	1
HB4-551	B-3	0-2	Z	HAND BORING	Σ	Selonlum	ON.	z	mg/kg	0.5
HB4-551	8-3	0-2	Z	HAND BORING	Σ	Silvar	Q		mg/kg	-
HB4-551	B-3	0-5	Z	HAND BORING	Σ	Thalllum	ON.	z	mg/kg	10
HB4-551	2	6-0	2	CINIO CINAL	**	7122	2		;	

RICKENBACKERANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

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SAMPLE .	GRID .	DEPTH	IN/OUT	DESCRIPTION CA	сатевопу	-ANALYSIS FOR P	RESULTS Q	QUAUFIER UNITS LIMIT	UNITS	DETECTION
HB4-552	B-3	0-2	Z	HAND BORING	>	ALL VOLATILES	Q.	N V	ug/kg	¥ Z
HB4-552	8-3	0-2	Z	HAND BORING	S	ALL SEMI -VOLATILES	ON	A N	ug/kg	YY Y
HB4-552	B3	0-2	Z	HAND BORING	×	Antimony	Q	z	mg/kg	Ð
HB4-552	8-3	0-2	Z	HAND BORING	Σ	Arsenic	45	z	mg/kg	0.5
HB4-552	B-3	0-2	<u>z</u>	HAND BORING	2	Boryllium	6.0	•	mg/kg	0.1
HB4-552	B-3	0-2	Z	HAND BORING	×	Cadmium	QN		mg/kg	-
HB4-552	8-3	0-2	Z	HAND BORING	×	Chromhm	15	O	mg/kg	-
HB4-552	B-3	0-2	Z	HAND BORING	×	Copper	30		mg/kg	-
HB4-552	8-3	0-2	Z	HAND BORING	ž	Lead	20	•	mg/kg	'n
HB4-SS2	8-3	0-2	Z	HAND BORING	Σ	Mercury	2	כ	mg/kg	0.0
H84-SS2	B-3	0-2	Z	HAND BORING	X	Nickel	09	•	mg/kg	-
HB4-552	8-3	0-2	Z	HAND BORING	Σ	Salanlum	Q	z	mg/kg	0.5
HB4-552	8-3	0-2	Z	HAND BORING	Z	Silvar	Q.		mg/kg	-
HB4-552	8-3	0-2	ž	HAND BORING	×	Thallium	ON.	z	mg/kg	0
HB4-552	B-3	0-2	N.	HAND BORING	M	Zinc	131		mg/kg	-
HB4-SS3	B-3	3-5	Z	HAND BORING	>	ALL VOLATILES	2	¥	ug/kg	₹ Z
HB4-5S3	8-3	3-5	Z	HAND BORING	SV	ALL SEMI-VOLATILES	2	Y Y	ug/kg	۲ ۲
HB4-553	B-3	3-5	Z	HAND BORING	Σ	Antimony	QN	z	mg/kg	9
HB4-SS3	B-3	3-5	Z	HAND BORING	×	Arsonic	58	z	mg/kg	0.5
HB4-5S3	B-3	3-5	Z	HAND BORING	Σ	Baryllium	9.0	•.	mg/kg	0.1
HB4-553	8-3	3-5	Z	HAND BORING	Σ	Cadmium	Q		mg/kg	-
HB4-553	8-3	3-5	Z	HAND BORING	×	Chrombin	12	O	mg/kg	-
HB4-553	8-3	3-5	Z	HAND BORING	Σ	Copper	35	•	mg/kg	-
HB4-553	8-3	3-5	Z	HAND BORING	\$	Pead	10	•	mg/kg	ĸ
HB4-553	B-3	3-5	Z.	HAND BORING	×	Marcury	2	כ	mg/kg	0.1
HB4-553	8-3	3-5	Z	HAND BORING	Σ	Nickol	4	•	mg/kg	-
HB4-SS3	8-3	3-5	z	HAND BORING	×	Sol	2	Z	mg/kg	0.5
HB4-SS3	B-3	3-5	Z	HAND BORING	×	Silvor	QN		mg/kg	-
HB4-SS3	B-3	3-5	z	HAND BORING	×	Tha	Q	z	mg/kg	10
HB4-SS3	B-3	3-5	Z	HAND BORING	Σ	Zinc	100		mg/kg	-

RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

₹	- 44	<	ø	0.5	0.1	-	_	-	S	<u>.</u>	_	0.5	-	0	-		¥	¥	0	0.5	0.1	-	_	-	S	0.1	-	0.5	-	2	-
Z	2	Z		0	0					0		•					_	_		•	_										
ug/kg	07/01	Bx/Bn	mg/kg	ш9/ка	mg/kg	UX/QE	mg/kg	ша/ка	mg/kg		ug/kg	ng/kg	mg/kg	тя/ка																	
٧X	44	2	z	. "					•	ם		z			z		¥ Z	ď Z	z						•	ב		z			z
2	2	2	Q N	9.4	0.5	Q	0	23	10	Q Q	25	QN	QN	Q	03	-	Q	Q.	Q.	4.1	0.5	9	01	22	=	QN	10	ON.	0 2	Q	79
ALL VOLATILES	01 114 1011 1140	ALL SEMI-VOLATILES	Antimony	Arsenic	Berylllum	Cadmlum	Chrombin	Coppor	Load	Marcury	Nickol	Selonium	Silvar	Thallium	Zlnc		ALL VOLATILES	ALL SEMI-VOLATILES	Antimony	Arsonic	Borylllum	Cadmlum	Chromium	Copper	Lead	Marcury	Nickol	Selonlum	Silvor	Thallium	Zinc
>		20	Σ	Σ	Σ	×	Σ	Σ	Σ	Σ	ĭ	Σ	¥	Σ	Σ		>	S	Σ	Σ	Σ	Σ	Σ	×	Σ	Σ	Σ	×	æ	Σ	Z
HAND BORING		HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING		HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING
2		z	Z	Z	z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z		Z	Z	Z	Z	z	Z	Z	Z	Z	Z	Z	z	Z	Z	Z
0-0	>	0-2	02	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2		0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-5	0-2	0-2	0-2	0-2
7.			Q-1	C-4	C-4	C-4	C-4	014	7-0	¥-0	Q-4	C-4	0-14	Q-4	0-14		0-5	D-5	0-5	0-5	0-5	0-5	0-5	0-5	0-5	0-5	0-5	0-5	D-5	0-5	0-5
אטפייטטו	2	HB5-551	HB5-551	HB5-551	HBS-551	HB5-551	H05-551		HB6-551	HB6-551	HB6-551	HB6-551	HBG-551	HB6-551	HB6-551	HB6-551	HB6-551	HB6-551	HB0-551	HB6-551	HB6-551	HB6-551	HB6-551								
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RICKENBACKERANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

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DETECTION	X A	AN AN	9	0.5	0.1	_	-	-	w	0.1	-	0.5	-	01		¥z	¥ Z	•	0.5	0.1					0.0		0.5		_	
UNITS	ug/kg	ug/kg	mg/kg	ug/kg	ng/kg	mg/kg	mg/kg	mg/kg	mg/kg	та/ка	mg/kg																			
QUAUFIER	Y Z	Y X	z						•	J.		z			2	ž	¥	z						•	3		z			Z
RESULTS O	QN	Q	2	15	0.5	Q	12	32	10	Q.	25	ON.	Q	QN	98	QN	Q	Q	12.	9.0	Q.	4	52	15	ON	25	2	Q	Q	92
ANALYSIS FOR	ALL VOLATILES	ALL SEMI-VOLATILES	Antimony	Arsonic	Beryllium	Cadmlum	Chromhm	Copper	Load	Mercury	Nickol	Solonlum	Silvar	Thallium	Zinc	ALLVOLATILES	ALL SEMI-VOLATILES	Antimony	Arsenic	Boryllium	Cadmlun	Chromium	Copper	Load	Morcury	Nickol	Solonium	Silvor	Thalilum	Zing
сатевопу	>	SV	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	\$	Σ	>	S	Σ	×	Σ	Σ	Σ	Σ	×	Σ	Σ	Σ	\$	≨	×
) ревспіртом (HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING	HAND BORING
ΙΝ/ΟυΤ	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	z	Z	Z	Z	Z	Z	2	Z	Z	Z
ОЕРТН	0-2	0-2	0-2	0-2	0-2	0-2	0-5	0-2	0-2	0-2	0-2	0-2	. 0-2	0-2	0-2	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5	3-5
GRID∵#	1 1	0-5	0-5	D-5	0-5		0-5	0-5	D-5	0-5	05	0-5	0-5	D-5	0-5	0-5	0-5	0-5	0-5	D-5	0-5	D-5	0-5	D-5	9-Q	D-5	0-5	0-5	0-5	
SAMPLE .	HB6-888	HB6-552	HB6-552	HB6-552	HB6-552	HB6-552	HB6-552	HB0-552	HB6-552	HB6-552	HB6-552	HB6-552	HB6-552	HB6-552	HB8-552	HB6-5S3	HB6-553	HB6-553	CSS-98H	HB6-553	HB6-553	HB6-553	HB6-553	HB6-553	HB6-553	H86-553	HB6-553	HB6-553	HB6-553	HB6-553

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SAMPLE # C	апо €	ОЕРТН	IN/OUT	DESCRIPTION CA	САТЕВОВУ	ANALYSIS FOR	RESULTS Q	QUALIFIER UNITS		DETECTION
MW1-551	0	13-15	Z	MONITORING WELL	>	Elhylbenzono	1700		ug/kg	v
MW1-551	0	13-15	Z	MONITORING WELL	>	o-Xylana	1600		ug/kg	<u>د</u>
MW1-551	0 -0	13-15	z	MONITORING WELL	S	ALL SEMI-VOLATILES	2	Y X	ug/kg	Y Y
MW1-551	0-0	13-15	Z	MONITORING WELL	Σ	Antimony	ON.	z	mg/kg	9
MW1-551	0-8	13-15	Z	MONITORNG WELL	Σ	Arsonic	7.0	z :	mg/kg	5.0
MW1-551	0-0	13-15	Z	MONITORING WELL	Σ	Boryllium	0	z	mg/kg	
MW1-551	010	13-15	Z	MONITORING WELL	¥	Cadmium	QN	z	mg/kg	
MW1-551	9-0	13-15	Z	MONITORING WELL	Z	Chrombin	2.5	E	mg/kg	-
MW1-551	0-6	13-15	Z	MONITORING WELL	Σ	Copper	12.7	z :	mg/kg	- 1
MW1-551	0-0	13-15	Z	MONITORING WELL	Σ	Load	ON I	Z	mg/kg	n (
MW1-551	0-0	13-15	Z	MONITORING WELL	Σ	Morcury	ONG	:	mg/kg	
MW1-551	0-0	13-15	Z	MONITORNG WELL	Σ	Nickol	13.3	z	тр/кр	- 1
MW1-551	0-0	13-15	Z	MONITORNG WELL	Σ	Solonlum	2	≥	mg/kg	0.5
MW1-551	0-0	13-15	Z	MONITORING WELL	¥	Silvar	Q	z	mg/kg	-
MW1-551	0-0	13-15	Z	MONITORNG WELL	Σ	Thalllum	Q	z	mg/kg	01
MW1-551	8-D	13-15	Z	MONITORING WELL	×	Zinc	50.1	z	mg/kg	-
MW1-552	0-0	14-18	Z	MONITORING WELL	>	Benzene	1900		ug/kg	S.
MW1-552	0-0	14-16	Z	MONITORING WELL	>	Ethylbanzona	1 1000		ממ/אמ	ימ
MW1-552	9-0	14-18	Z	MONITORING WELL	>	o-Xylana	20000		ug/kg	S.
MW1-552	0-0	14-18	Z	MONITORING WELL	28	Naphthalone	630		ug/kg	330
MW1-552	0-0	14-10	Z	MONITORING WELL	S	2 - Mathylnaphthalana	1200		ug/kg	330
MW1-552	0-0	14-16	Z	MONITORING WELL	Σ	Antimony	2	z	mg/kg	9
MW1-552	9-0	14-18	Z	MONITORING WELL	×	Arsanic	10.9	z	mg/kg	0.5
MW1-552	0-0	14-16	Z	MONITORING WELL	×	Boryllium	2	z	mg/kg	0.0
MW1-552	9-O	14-18	Z.	MONITORING WELL	Σ	Cadmlum	Q Z	z	mg/kg	-
MW1-552	0-0	14-16	z	MONITORING WELL	×	Chrombm	4.0	z	mg/kg	
MW1-552	9-D	14-18	Z	MONITORING WELL	Σ	Copper	17.2	z	mg/kg	-
MW1-552	9-O	14-18	Z	MONITORING WELL	Σ	Load	9.0	8	mg/kg	20
MW1-552	0-0	14-16	Z	MONITOPING WELL	Σ	Mercury	DNG		mg/kg	0.1
MW1-552	0-0	14-16	Z	MONITORING WELL	Σ	Nickel	22.2	z	па/ка	-
MW1-552	0-0	14-18	Z	MONITORNG WELL	×	Sol	0	≥ :	mg/kg	0.5
MW1-552	0-0	14-18	Z	MONITORING WELL	Σ		2	z	ma/kg	- 1
MW1-552	0-0	14-16	z	MONITORING WELL	X	Tha	QX	z	ng/kg	0,
MWI-SS2	0-0	14-18	Z	MONITORING WELL	Σ	Zluc	63.0		mg/kg	

RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

L MIT	¥ Z	¥	•	0.5	0.1	_		- v	, -	; -		2	- 0			-	7 4		5 4	3 6				- 0	G ;		- 0	ů.	- ç	2 •	-
L	ug/kg	חמ/גם	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	0×0	UN/011	ma/kg	100/kg	mg/kg	Da/kg	Du Allen	тид/кд		04/0n	200	DX/DE	וויייייייייייייייייייייייייייייייייייי	Da/gm	מא/מעו	mg/kg	By/Bu	mg/kg	mg/kg	Dx/DLL	mg/kg	mg/kg	ga/gm	mg/kg
RESULTS QUAUFIER UNITS	A X	¥ Z	z	z	_				•		- 2	Z		-	Z		411	2	z	2 6	n						:	z		. :	z
ESULTS QU	QV	ON.	QN	15	Q.	Q	15	27	2 9	2 2	S S	Q S	2 5	2	99	-	o (2 !	ON I	0.7	0.2	2	12	21	5	Q	51	Q !	Q :	O T	74
FOR	ALL VOLATILES	ALL SEMI-VOLATILES	Antlmony	Arsonic	Boryllum	Cadmlum	Chromhm	Coppor	Lead	Morcury	Nicker	Salonlun	Silvor	- Inalium	Zinc		Metrylene Chloride	ALL SEMI-VOLATILES	Antimony	Arsenic	Baryllum	Cadmium	Chromum	Coppor	Lead	Mercury	Nicko	Solonium	Silvar	Thallium	Zinc
сатевояу	>	S	×	×	×	2	¥	\$	Σ	X :	Σ	Σ	Σ	Σ	×		>	S.	*	ž	Σ	Σ	Σ	¥	Σ	×	\$	Z	×	¥	
DESCRIPTION CA	MONTORNG WELL	MONTORNG WELL	MONTORING WELL	MONITORING WELL	MOMTORING WELL	MONTORING WELL	MOMTORNG WELL	MONITORNG WELL	MONTOPING WELL	MONITORING WELL	MONITORNG WELL	MONTORNO WELL	MONITORING WELL	MONITORING WELL	MONITORNG WELL		MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITORING WELL	MONITORNG WELL	MONITORNG WELL	MONTORNG WELL	MONTORNG WELL	MONITORNG WELL	MONITORING WELL	MONITORNO WELL	MONITORING WELL	MOMTORNG WELL	MOMTORING WELL	THE CHARGE THE CALL
IN/OUT	TITO	150	110	100	TUO	OUT	DOUT	TUO	OUT	DOCT	DOUT	OUT	OUT	DOO	TUO		TUO	DOUT	TUO	TUO	DOUT	DOUT	TUO	DOUT	TUO	TUO	DOUT	TUO	DOOT	OUT	1
DEPTH	- 6-0	7 0	1 0	1 6	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2		0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	(
		7 - 0	1 1	2 0	B - 7	B-7	8-7	8-7	B-7	B-7	B-7	8-7	8-7	0-7	8-7		8-7	B-7	B-7	B-7	B-7	8-7	8-7	B-7	8-7	B-7	B-7	B-7	8-7	B-7	
SAMPLE # GRID #		MW2-551	MW2-331	MWZ I SSI	MW2-595	WW2-5551	MW2-551	MW2-551	MW2-551	MW2-551	MW2-551	MW2-551	MW2-551	MW2-551	MW2-551	ł	MW2-552	MW2-552	MW2-552	MW2-552	MW2-552	MW2-552	MW2-552	MW2-552	WW2-552	MW2-552	MW2-552	MW2-552	MW2-552	MW2-552	

RICKENBACKER ANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

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8-7	3-5	TUO	MONITORING WELL	>	Methylene Chloride	S		ug/kg	Ŋ
B-7	3-5	DOUT	MONITORING WELL	SV	ALL SEMI-VOLATILES	ON.	¥	ug/kg	¥ X
8-7	3-5	TUO	MONITORING WELL	Z	Antimony	ON	z	mg/kg	8
B-7	3-5	DOUT	MONITORING WELL	Σ	Arsenic	9.8	z	mg/kg	0.5
8-7	3-5	DOUT	MONITORING WELL	×	Baryillum	QX		mg/kg	0.1
B-7	3-5	TUO	MONITORING WELL	Σ	Cadmlum	QN	•	mg/kg	-
B-7	3-5	TUO	MONITORING WELL	Z	Chrombim	0		mg/kg	-
B-7	3-5	TUO	MONITORNG WELL	Σ	Coppor	24		mg/kg	-
B-7	3-5	OUT	MONITORING WELL	Σ	Lead	28		mg/kg	S
8-7	3-5	DOUT	MONITORING WELL	×	Mercury	9		mg/kg	0.1
9-7	3-5	DOUT	MONITORNG WELL	Σ	Nickof	9.5	63	mg/kg	-
0-7	3-5	TUO	MONITORING WELL	Σ	Solonium	9	z	mg/kg	0.5
8-7	3-5	DOUT	MONITORING WELL	Σ	Silvar	QN		mg/kg	-
B-7	3-5	TUO	MONITORING WELL	Σ	Thalllun	2		mg/kg	0
B-7	3-5	OUT	MONITORING WELL	Σ	Zinc	7.6	z	mg/kg	
						:		;	;
V-0	0-2	Z	MONITORING WELL	25	ALL SEMI-VOLATILES	Q	₹ Z	ug/kg	₹ Z
C	02	Z	MONITORING WELL	Σ	Antimony	2	z	mg/kg	9
0-1V	0-2	Z	MONITORING WELL	×	Arsenic	15.3		mg/kg	0.5
C-4	0-5	Z	MONITORING WELL	X	Berylllum	2	z	mg/kg	0.1
C-1	0-2	Z	MONITORING WELL	2	Cadmlum	Q		mg/kg	-
7-5 	0-2	Z	MONITORING WELL	Z	Chromkim	13.1		mg/kg	-
0-1	0-2	z	MONITORING WELL	Z	Copper	20.5		mg/kg	-
7-0	0-2	Z	MONITORING WELL	Σ	Lead	15.6		mg/kg	S
7-0	0-5	Z	MONITORING WELL	Σ	Mercury	ON		mg/kg	0.1
0-1	0-2	Z	MONITORING WELL	Σ	Nickel	24.1		mg/kg	-
70	0-2	Z	MONITORING WELL	Σ	Solonium	2	≯	mg/kg	0.5
20	0-2	Z	MONITORING WELL	Σ	Silvar	Q	z	mg/kg	-
2-2	0-2	Z	MONITORING WELL	Σ	Thallium	QN		mg/kg	10
2-5	0-2	N	MONITORING WELL	×	Zinc	1.70		mg/kg	1

RICKENBACKERANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - SOIL

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SAMPLE CGHID								_	
C-4	0-10	Z	MONITORNG WELL	28	ALL SEMI-VOLATILES	QV	A X	ug/kg	Y Y
0 4-0	8-10	Z	MONITORNG WELL	Z	Antimony	QV	z	mg/kg	6
C-4	0-10	Z	MONITORING WELL	Σ	Arsonic	16.1		mg/kg	9.0
0-14	0-10	Z	MONITORING WELL	2	Boryillum	Q	z	mg/kg	0.1
C-4	0-10	Z	MONITORING WELL	Σ	Cadmlum	Q		mg/kg	-
C-4	0-10	Z	MONITORING WELL	Σ	. Chromium	14.1	-	mg/kg	-
C-4	0-10	Z	MONITORNG WELL	Σ	Coppor	19.1		mg/kg	-
C-4	9-10	Z	MONITORING WELL	×	Pood	13.7		mg/kg	50
C-4	6-10	Z	MONITORING WELL	¥	Mercury	ON.		mg/kg	1.0
C-4	0-10	Z	MOMTORING WELL	Σ	Nickol	24.9		mg/kg	
C-4	0-10	Z	MONITORING WELL	Σ	Solonlum	Q.		mg/kg	0.5
C-4	0-10	Z	MONITORING WELL	Σ	Silvor	Q	z	mg/kg	-
C-4	8-10	Z	MONITORING WELL	Σ	Thallium	10.5		пу/ка	01
C-4	0-10	Z	MOMITORING WELL	Σ	Zinc	63.7		та/ка	-
2 - 2 - 2	2	2	SUBFACE SOIL SAMPLE	>	Banzo(k) Fluoranthena	200		ua/ka	330
2 2 2 2 2 2	1 2	Z	SURFACE SOIL SAMPLE	25	Fluoranthana	430		UQ/kg	330
8-2.C-2	0-2	Z	SURFACE SOIL SAMPLE	Σ	Antimony	QN	z	mg/kg	9
B-2,C-2	0-2	Z	SURFACE SOIL SAMPLE	Σ	Arsonic	14.1		mg/kg	0.5
B-2,C-2	0-2	Z	SURFACE SOIL SAMPLE	Σ	Beryllum	0.79		mg/kg	0.1
B-2,C-2	0-2	Z	SURFACE SOIL SAMPLE	Σ	Cadmlum	6.3	g	mg/kg	-
B-2,C-2	0-2	Z	SURFACE SOIL SAMPLE	Σ	Chrombin	22.1		mg/kg	-
B-2,C-2	0-2	Z	SURFACE SOIL SAMPLE	Σ	Coppor	29.3		mg/kg	-
B-2,C-2	0-5	Z	SURFACE SOIL SAMPLE	Σ	Lead	41.9		mg/kg	ιn
B-2,C-2	0-2	Z	SURFACE SOIL SAMPLE	Σ	Morcury	QN		mg/kg	0.1
0-2,C-2	0-2	Z	SURFACE SOIL SAMPLE	×	Nickol	20.2		mg/kg	-
B-2,C-2	0-2	Z	SURFACE SOIL SAMPLE	Σ	Solonium	Q	*	mg/kg	0.5
B-2,C-2	0-2	Z	SURFACE SOIL SAMPLE	¥	Silvar	QN		mg/kg	-
B-2,C-2	0-2	Z	SURFACE SOIL SAMPLE	×	Thallum	QN		mg/kg	10
B-2,C-2	0-2	Z	SURFACE SOIL SAMPLE	M	Zinc	422		mg/kg	-

ECTION	UMIT	000	330	9	0.5	0.1	-	-	-	S	0	-	0.5	-	10
DETECTION	UNITS UMIT	ug/kg	ug/kg	mg/kg	то/ко	mg/kg									
	JAUFIER			z			Ø								
	RESULTS QUALIFIER	200	570	QN	19.7	0.0	0.3	22.4	20.1	35.3	ON.	20.3	Q	0 2	ON N
ANALYSIS		Benzo(k) Fluoranthone	Fluoranthone	Antimony	Arsonic	Boryllum	Cadmlum	Chrombm	Copper	Lead	Morcury	Nickol	Solonium	Silvar	Thalllum
	САТЕВОЯУ	SS	SV	Σ	¥	\$	×	×	×	Σ	×	Σ	Σ	Σ	×
	DESCRIPTION CA	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE
000000000000000000000000000000000000000	IN/OUT	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	2
	DEPTH	0-10	0-2	0-5	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2
\$65.500 (Jan. 196.)	GRID .	0-00-0	0-20-3	0-20-3	D-2.D-3	0-2,0-3	D-2.D-3	D-2.D-3	D-2.0-3	D-2.D-3	D-2,D-3	0-2,0-3	0-2,0-3	0-2,0-3	0-20-3
	SAMPLE #	0-0-0-0-0-0-0-0-0-0-0-0-0-0-0-0-0-0-0-	SU3+4-GS1 D-2D-	SU3+4-GS1 D-2.D-	SU3+4-GS1 D-2,D-	SU3+4-GS1 D-2,D-3	SU3+4-GS1 D-2.D-3	SU3+4-GS1 D-2.D-3	SU3+4-GS1 D-2.0-3	SU3+4-GS1 D-2.D-3	SU3+4-GS1 D-2,D-3	SU3+4-GS1 D-2.D-3	SU3+4-GS1 D-2,D-3	SU3+4-GS1 D-2,D-3	SU3+4-GS1 D-2.D-3

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AUFIER SUNIS	0CC 03/60	0CC 330	00CC 03/kg		occ na/gu	00cc 04/8n	occ By/8n	ug/kg 330	02/kg 330	00CC 03/kg	0CC 23/0n		ug/kg 330	N mg/kg 6	S mg/kg 0.5	mg/kg 0.1	G mg/kg 1	mg/kg 1	mg/kg 1	mg/kg 5	mg/kg 0.1	mg/kg 1	W mg/kg 0.5	mg/kg 1	mg/kg 10	- 1
HESULIS QUADFIER	200	2100	2600	3200	1700	2000	4100	2000	360	4100	1700	4200	2600	QN	17.71	0.75	6.0	25.0	8.73	107	QV	19.1	QN	Q	Q	
FOIR	Anthracono	Banzo(n) Anthracene	Benzo(a)Pyrene	Benzo(b) Fluoranthone	Benzo(g,h,l)Porylone	Bonzo(k) Fluoranthone	Bis(2-othythoxyt)Phthalato	Chrysona	Dibonz(a,h)Anthracono	Fluoranthono	Indano(12,3-cd)Pyrana	Phonanthrono	Pyrono	Antimony	Arsonic	Boryillum	Cadmlum	Chrombm	Coppor	Lead	Marcury	Nickel	Solonlum	Silvar	Thallium	1
CATEGORY	SV	SV	SV	SV	S	SV	SV	S	SV	SV	SV	S	SS	Σ	Σ	×	Σ	×	Σ	×	₹.	×	Σ	2	×	
DESCRIPTION	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	
IN/OUT	Z	2	Z	ž	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	
DEPTH IN/OUT	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	
GMID	B-3,C-3	B-3,C-3	B-3,C-3	B-3,C-3	B-3,C-3	B-3,C-3	B-3,C-3	B-3,C-3	B-3,C-3	B-3,C-3	B-3,C-3	B-3,C-3	B-3,C-3	B-3,C-3	B-3,C-3	B-3,C-3	B-3,C-3	B-3,C-3	B-3,C-3	B-3,C-3	B-3,C-3	B-3,C-3	B-3,C-3	B-3,C-3	B-3,C-3	
	SU5+6-GS1	SU5+6-GS1	SU5+6-GS1	SUS+6-GS1	SU5+6-GS1	SUS+6-GS1	SUS+6-GS1	SU5+6-GS1	SU5+6-GS1	SUS+6-GS1	SUS+6-GS1	SUS+6-GS1	SUS+6-GS1	SUS+6-GS1	SU5+6-GS1	SUS+6-GS1	SU5+6-GS1	SUS+6-GS1	SUS+6-GS1							

0-2	Z	SURFACE SOIL SAMPLE	SV	Banzo(b) Fluoranthone	230		ug/kg	330
0-2	Z	SURFACE SOIL SAMPLE	SV	Di-n-butyl Phihafale	940		ug/kg	330
0-2	z	SURFACE SOIL SAMPLE	SV	Fluoranthane	520		ug/kg	33
0-2	Z	SURFACE SOIL SAMPLE	Σ	Antimony	Q	z	mg/kg	89
0-2	Z	SURFACE SOIL SAMPLE	2	Arsonic	15.2		mg/kg	9'0
0-2	Z	SURFACE SOIL SAMPLE	Σ	Boryllum	6.0		mg/kg	0.1
0-2		SURFACE SOIL SAMPLE	Σ	Cadmium	9	Ø	mg/kg	-
0-2	Z	SURFACE SOIL SAMPLE	Σ	Chrombin	22.1		mg/kg	-
0-2	Z	SURFACE SOIL SAMPLE	Σ	Coppor	23.1		mg/kg	-
0-2	Z	SURFACE SOIL SAMPLE	Σ	Load	55.7		mg/kg	S
0-2	Z	SURFACE SOIL SAMPLE	×	Morcury	QN		mg/kg	0.1
0-5	Z	SURFACE SOIL SAMPLE	×	Nickol	20.0		mg/kg	-
1-4,0-4	Z	SURFACE SOIL SAMPLE	Σ	Solonium	QN		mg/kg	5,0
B-4,C-4 0-2	Z	SURFACE SOIL SAMPLE	2	Silvor	Q		mg/kg	-
1-4,C-4 0-2	Z	SURFACE SOIL SAMPLE	Z	Thallium	QN		mg/kg	10
4 0-2	Z	SURFACE SOIL SAMPLE	×	Zinc	298		mg/kg	-
-4	Z	SUBFACE SOIL SAMPLE	ΛS	Benzo(a)Pyrene	360		uo/ka	330
-4 0-2		SURFACE SOIL SAMPLE	SV	Banzo(b) Fluoranthana	570		ug/kg	330
-4 0-2	Z	SURFACE SOIL SAMPLE	SV	Bonzo(k) Fluoranthona	380		ug/kg	330
-4 0-2	Z	SURFACE SOIL SAMPLE	SV	Chrysono	420		UU/KO	330
-4 0-2	Z	SURFACE SOIL SAMPLE	28	Fluoranthene	840		ug/kg	930
-4 0-2		SURFACE SOIL SAMPLE	SV	Phenanthreno	410		ng/kg	930
-4 0-2	Z .	SURFACE SOIL SAMPLE	SV	Pyreno	009		ug/kg	330
-4 0-2	Z Z	SURFACE SOIL SAMPLE	×	Antimony	ON	z	mg/kg	9
-4 0-2	Z	SURFACE SOIL SAMPLE	Σ	Arsonic	17.9		mg/kg	0.5
-4	Z	SURFACE SOIL SAMPLE	×	Boryllium	0.95		mg/kg	0.1
D-4,E-4 0-2	Z	SURFACE SOIL SAMPLE	¥	Cadmlum	9.1	g	mg/kg	-
		SURFACE SOIL SAMPLE	×	Chrombin	20.1		mg/kg	-
D-4,E-4 0-2		SURFACE SOIL SAMPLE	×	Coppor	73.1		mg/kg	-
D-4,E-4 0-2	Z	SURFACE SOIL SAMPLE	X	Lead	72.0		mg/kg	S
-4 0-2		_	×	Mercury	Q		mg/kg	0.1
4	N	SURFACE SOIL SAMPLE	Σ	Nickol	26.6		mg/kg	-
-4 0-2	Z	SURFACE SOIL SAMPLE	Σ	Selonium	Q	٠	mg/kg	0.5
-4 0-5			Σ	Silver	2		та/ка	_
-4	Z		Σ	Thalllum	Q		mg/kg	10
0-5	Z	SURFACE SOIL SAMPLE	2	Zinc	280		270	_

SAMPLE & GRID & DEPTH	EPTH IN/OUT	DESCRIPTION	слтевопу		×		-	
SU11+12-GS C-5,D-5	0-2 IN	SURFACE SOIL SAMPLE	S	Banzo(a)Pyrene	230		ug/kg	000
C-5,D-		SURFACE SOIL SAMPLE	SV	Banzo(b) Fluoranthana	050		UQ/kg	330
SU11+12-GS C-5,D-5	0-2	SURFACE SOIL SAMPLE	SV	Banzo(k)Fluoranthena	9		ug/kg	330
SU11+12-GS C-5,D-5	0-2	SURFACE SOIL SAMPLE	SV	Bis(2-ethythoxyl)Phthalato	4 80		ug/kg	330
SU11+12-GS C-5,D-5	0-2	SURFACE SOIL SAMPLE	SV	Chrysene	540		ug/kg	330
SU11+12-GS C-5,D-5	0-2	SURFACE SOIL SAMPLE	SV	Di-n-butyl Phihalato	6500		ng/kg	000
SU11+12-GS C-5,D-5	0-2 IN	SURFACE SOIL SAMPLE	SV	Fluorena	1400		UQ/kg	330
SU11+12-GS C-5,D-5	0-2 IN	SURFACE SOIL SAMPLE	SV	Phenonthrono	1100		ug/kg	330
SU11+12-GS C-5,D-5	0-2	SURFACE SOIL SAMPLE	۸s	Pyrene	1000		ug/kg	330
SU11+12-GS C-5,0-5	0-2 IN	SURFACE SOIL SAMPLE	Σ	Antimony	Q N	z	mg/kg	9
SU11+12-GS C-5,D-5	0-2 IN		¥	Arsenic	17.7		mg/kg	0.5
SU11+12-GS C-5,D-5	0-2 IN		Σ	Boryllium	0.70		mg/kg	0.1
SU11+12-GS C-5,D-5	0-2		Σ	Cadmlum	60	Ø	mg/kg	-
SU11+12-GS C-5.0-5	0-2 IN		Σ	Chromium	19.5		mg/kg	-
SU11+12-GS C-5,D-5	0-2	SURFACE SOIL SAMPLE	Σ	Coppor	29.7		mg/kg	~
SU11+12-GS C-5,D-5	0-2 IN	SURFACE SOIL SAMPLE	Σ	Load	35		mg/kg	S
SU11+12-GS C-5,D-5	0-2	SURFACE SOIL SAMPLE	≆	Marcury	QN		mg/kg	0.1
SU11+12-GS C-5,D-5	0-2	SURFACE SOIL SAMPLE	Σ	Nickel	26.5		mg/kg	_
SU11+12-GS C-5,D-5	0-2	SURFACE SOIL SAMPLE	×	Sel	ON.	S	mg/kg	0.5
SU11+12-GS C-5,D-5	0-2 IN	SURFACE SOIL SAMPLE	Σ	Silvar	Q.		mg/kg	-
SU11+12-GS C-5,D-5	0-2 IN	SURFACE SOIL SAMPLE	Σ	Tha	2		mg/kg	01
SU11+12-GS C-5,D-5	0-2 IN	SURFACE SOIL SAMPLE	×	Zinc	113		mg/kg	-
		7 10140 100 1041010	è		5		27/01	c
	V 1		ò	מווכחומו	3 5		DA/SO	3 8
	2 .) : 	Pluc	9	•	ng/kg	000
10.6	2 .		Σ:	₹	2 (ž	gx/gm	0 (
B-6,C-	2 .		Σ:		12.8		gx/gm	6.0
B-6,C-	~		Σ:		0.89	(Da/kg	0.1
			Σ:		1.7	9	DX/XD	- ,
			Σ :	5	9.0		DX/DH	
	2		Σ	ŭ 	18.7		mg/kg	-
B-6,C-	2		≆		41.5		mg/kg	LS.
			X	2	Q.		mg/kg	0.1
	2	-	X		14.8		mg/kg	-
SU13+14-GS B-6,C-5	2		Σ	Sol	Q		mg/kg	0.5
SU13+14-GS B-6,C-5	0-2	SURFACE SOIL SAMPLE	Σ	Sliver	ON.		mg/kg	-
	0-2 IN		Σ	Tha	2		mg/kg	10
SU13+14-GS B-6,C-5	0-2 IN	SURFACE SOIL SAMPLE	Σ	Zinc	120		mg/kg	-

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DETECTION CUMIT	000	3	230	000	9	0.5	0.1	-	-	-	S	0.1	-	9.0	-	10	-
		US/KD	ng/kg	ug/kg	mg/kg	mu/ka	mg/kg	mg/kg	mg/kg								
QUAUFIER UNITS					Z			O						}			
RESULTS QUALIFIER UNITS	0,0	250	740	520	Q	17.2	0.75	7.7	23.9	20.7	52.1	ON N	26.4	2	Q.	Q	135
ANALYSIS FOR	10.47	pauzo(p) rinotaninana	Fluoranthono	Pyrene	Aritimony	Arsonic	Boryllium	Cadmium	Chrombm	Соррог	Load	Mercury	Nickel	Salonium	Silvar	Thallium	Zinc
CATEGORY	7	20	SV	SV	Σ	2	Σ	Σ	×	Σ	Σ	Σ	Σ	Σ	Σ	Σ	¥
DESCRIPTION	2 10142 102 20420112	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE	SURFACE SOIL SAMPLE
IN/OUT	1	2	z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	Z	z	N
DEPTH IN/OI	c	710	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2	0-2
GRID # DEPTH IN/O	6	010010	C-6,0-6	C-6,D-6	C-6,D-6	C-6,D-8	C-6,D-8	C-6,D-8	C-6,D-6	C-6,D-6	C-6,D-8	C-6,D-6	C-6,D-8	C-6,D-6	C-6,D-8	C-6,D-6	C-6.D-6
SAMPLE #	0 0 0	00-01-000	SU15+16-GS C-6,D-6	SU15+18-GS C-6,D-6	SU15+10-GS C-6,D-6	SU15+16-GS C-6,D-6	SU15+18-GS C-6.D-6										

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SAMPLE #	GRID # FILTERED	VO TED DATE	DESCRIPTION	CATEGORY	ANALYSIS FOR	RESULTS C	QÜALIFIER	UNITS	DETECTION
MWI-GWI	9	19-Sep-88	Groundwater Sample	>	Bonzone	• 46.		Ug/L	S
MW1-GW1	C-6	19-Sap-88	Groundwater Sample	>	Acelono	. 120	5	UG/L	8
MW1-GW1	9-0	19-Sop-80	Groundwater Sample	SV	ALL SEMI-VOLATILES	Q.	Y X	ng/L	AZ.
MW1-GW1	0-e	19-Sep-88	Groundwater Sample	>	o-Xylene	20		ng/L	S.
MWI-GWI	9-0	19-Sop-88	Groundwater Sample	Σ	Antimony	2		mg/L	0.06
MWI-GWI	0-0	19-Sep-88	Groundwater Sample	×	Arsonic	0.34		ng/L	0.005
MWI-GWI	9-0	19-Sep-88	Groundwater Sample	Σ	Beryllium	ON	z	mg/L	0.001
MW1-GW1	G-6	19-Sep-88	Groundwater Samplo	Σ	Cadmlum	0.15		mg/L	0.01
MW1-GW1	9-0	19-Sep-88	Groundwater Sample	Σ	Chromlum	0.52		mg/L	0.01
MW1-GW1	0-0	19-Sep-88	Groundwaler Sample	Σ	Copper	0.88		mg/L	0.01
MW1-GW1	9-0	19-Sep-88	Groundwater Sample	Σ	Load	0.82		mg/L	0.005
MW1-GW1	0-e	19-Sop-88	Groundwater Sample	Σ	Moreury	0.0003		mg/L	0.0002
MWI-GWI	9-O	19-Sop-80	Groundwater Sample	Σ	Nickol	0.84		mg/L	10.0
MW1-GW1	9-C	19-Sop-88	Groundwater Sample	Σ	Solonium	2	F	mg/L	0.005
MW1-GW1	9-0	19-Sop-88	Groundwater Sample	Σ	Shor	2	z	mg/L	0.01
MWI-GWI	0-e	19-Sop-88	Groundwater Sample	Σ	Thallium	Q		mg/L	0.1
MW1-GW1	9-0	19-Sop-88	Groundwater Sample	Σ	Zluc	3.6		mg/L	10.0
MWI-GWID	9-0	19 - Sep - 88	Groundwater Sample	>	Bonzaga	34		700	u
MW1-GW1D	9-0	19-50-88	Groundwater Sample	\ <u>\</u>	2 - Methylogothan	. 5		1/85	9 5
MWI-GWID	9-0	19-Sop-88	Groundwater Sample	Σ	Antimony	2 2		1/0E	90.0
MW1-GW1D	9-0	19-Sop-88	Groundwater Sample	Σ	Arsonic	4.0	н	mg/L	0.005
MW1-GW1D	C-6	19-Sop-88	Groundwater Sample	Z	Boryllum	ON.	z	mg/L	0.001
MW1-GW1D	9-0	19-Sop-88	Groundwater Samplo	Σ	Cadmlum	0.18	•	mg/L	0.01
MWI-GWID	9-0	19-Sep-88	Groundwater Sample	Σ	Chromlum	99.0		mg/L	10.0
MW1-GW1D	9-0	19-Sep-88	Groundwater Sample	Σ	Copper	1.1		mg/L	10.0
MW1-GW1D	9-0	19-Sop-88	Groundwater Sample	Σ	Load	0.99		mg/L	0.005
MW1-GW1D	9-0	19-Sop-88	Groundwater Samplo	Σ	Morcury	0.0003		mg/L	0.0002
MW1-GW1D	9-0	19-Sop-88	Groundwater Sample	Σ	Nicket	-		mg/L	0.01
MW1-GW1D	9-0	19-Sop-88	Groundwater Sample	Σ	Solonium	Q.	F	mg/L	0.005
MW1-GW1D	9-0	19-Sep-88	Groundwater Sample	Σ	SNer	2	z	mg/L	0.01
MWI-GWID	9-0	19-Sep-88	Groundwater Sample	Σ	Thallium	ON.		mg/L	0.1
MW1-GW1D	06	19-Sep-80	Groundwater Sample	Σ	Zinc	4.3		mg/L	0.01

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SAMPLE # 0	YE GRID # FIL	YES/NO FILTERED	ОЛТЕ	DESCRIPTION	CATEGORY	FOR	RESULTS	QUALIFIER. UNITS	ER. UNITS	LIMIT
	2		8 - 3	olu Hay relaylorises	>	ALL VOLATILES	Q.	¥ Z	ug/L	NA
MW2-GW1	1-9		00-dpc-6	oldman tolembring	λυ:	ALL SEMI-VOLATILES	ON .	YZ	ug/L	Y X
MW2-GW1	B-7	-	80-dos-6	Glouidwatel campio	> >	Antimony	ON		mg/L	90.0
MW2-GW1	8-7	_	99-des-6	Grownian Sample	2	Arsonic	0.184		mg/L	0.005
MW2-GW1	8-7	≃ —	19-Sop-88	Groundwater Sample	Ξ.	a illino	2	z	ma/L	0.001
MW2-GW1	B-7	=	19-Sop-88	Groundwater Sample	Σ:	Delyman	5	2	1/000	0.01
MW2-GW1	0-7	=	19-Sop-88	Groundwater Sample	Σ	Caamium	21.0	-	1/04	100
MW2-GW1	B-7	<u>-</u>	19-Sop-88	Groundwater Sample	Σ	Chromium	07.0		1/0 E	0.0
MW2-GW1	E-7	=	19-Sop-88	Groundwater Sample	Σ:	Jackdon .	20.0	-	1/0 11	0.005
MWZ-GW1	8-7	_	19-Sop-68	Groundwater Sample	Σ:	רפסק	6000		1/0	0.0002
MW2-GW1	B-7	_	19-Sop-08	Groundwater Sample	Σ	Malcul	2000,0		1/04	0.01
MW2-GW1	B-7	_	19-Sop-88	Groundwater Sample	Σ	Nexo	2 2	>	1/00	0.005
MW2-GW1	B-7	_	19-Sop-88	Groundwater Sample	Σ:	Enluga Enluga	2 2	- 2	1/00	0.01
MW2-GW1	B-7	_	19-Sep-88	Groundwater Sample	Σ	SIN OF	2 5	2	6	
MW2 - GW1	8-7	_	19-Sop-88	Groundwater Sample	Σ	Thailum	2 3		יייטער ר	5 6
WWD-GWM	B-7	_	19-Sep-88	Groundwater Sample	Σ	Zinc	2.4		1/6	
					^	Telchorothylon	44		טמ/ר	ĸ
MW3-GW1	0-14		22-Sop-88	Groundwaler Sample	> 2	SHILD IOV-MAD IIA	CZ	¥X	מס/ר	¥Z
MW3-GW1	C-4	~	22-Sop-88	Groundwater Sample	<u> </u>	ALL SEMI-YOUTHERS	2 5		mo/L	90,0
MW3-GW1	Q-4		22-Sop-88	Groundwater Sample	Σ	Anomary .	41.0		ma/L	0.005
MW3-GWI	C-4		22-Sop-88	Groundwater Sample	Σ:			2	1/04	0.001
MW3-GW1	0-14		22-Sop-80	Groundwater Sample	Σ	Baryllam		-	1/0	0.01
MW3-GWI	0-14		22-Sap-88	Groundwater Sample	Σ	Cadmium			1/6.00	000
MW3-GW1	C-4		22-Sop-80	Groundwater Sample	∑ :	Chromium	0.0		1/8.0	0.01
MW3-GWN	C-4		22-Sop-88	Groundwater Sample	Σ	Inddoo) / C	0 005
MW3-GW1	C-4		22-Sop-88	Groundwater Sample	Σ	Lead			1/611	0000
MW3-GW1	C-4		22-Sop-88	Groundwater Sample	Σ	Mercury			1,01	1000
MW3-GWM	C-4		22-Sep-88	Groundwater Sample	Σ	Nickol		,	1/800	20.0
MW3-GW1	0-14		22-Sap-88	Groundwater Sample	Σ	Solonium	ö.	ח	mg/L	20.0
MW3-GW1	0		22-Sep-88	Groundwater Sample	Σ	STOCK		Z.	1/A	5.0
MW3-GW1	0-4		22-Sep-88	Groundwater Sample	Σ	mullan!			1/A E	0.0
MW3-GW1	0-4		22-Sop-88	Groundwater Sample	Σ	7inc	50.0	c	1/6:	2
MW3-GW1	C-4		17-Oct-88	Groundwater Sample	SV	Bis (2 - of tylhoxyl) phihala			7	

RICKENBACKERANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - GROUNDWATER

SVMI LE F GUID			Control of the Contro							
				diam's	2	Antimomy	35.3	Э	ng/L	9
MWI-GW2	9-O	0	06-400-30	Groundwardt Sample		Arsonic	9,5	B	ug/L	9
MW1-GW2	9-0	2	06-Feb-90	Groundwaler Sample	E 3	Received	9.0	ר	UB/L	ĸ
MW1-GW2	0-0	2	06-Feb-90	Groundwater Sample	Σ:	ociyilari.	0.4		na/L	•,
MWI-GWD	9-0	ON.	06-Fob-90	Groundwater Sample	Ξ	Endenine.	2.4.6	•	1/0	9
240	1	C	06-Fab-90	Groundwater Sample	×	Chromium	2.5		1 /6 :	7,0
MAY I - CAVE	0 (2	06-Feb-90	Groundwater Sample	×	Copper			1,700	3 "
MWI-GWZ	ָרָ פְּ		200	Groundwater Sample	2	Load	195	-	חש/ר	
MW1-GWZ	9-0	2 :	08-100-100	Glownian Camping	~	Morcury	1.0	<u> </u>	ng/L	0.2
MW1-GW2	9-C		2001-00	old in the land in	-	Nickel	121		ng/L	4
MW1-GW2	9-0	2	06-Feb-90	Groundwater Sample	ξ :	Colonium	10		UQ/L	
MW1-GW2	0-0	2	06-Fob-90	Groundwater Sample	Σ:	100000		7	na/L	_
CWS-1WM	0-0		06-Fob-90	Groundwater Sample	Σ	1000		=	1/00	-
WW1-GW2	9-0		06-Fob-90	Groundwater Sample	Σ:	Thaildin	678)	T/on	20
7WY - GW2	0-0		06-Feb-90	Groundwaler Sample	Σ	71110	2 5	=	1/011	8
ZWO - WW	9-0		06-Fob-90	Groundwater Sample	Σ	Anumony	7.0) C	1/011	
WW. CWD	0 0			Groundwaler Sample	≊	Arsonic	F (7)		1 1 2	, vo
MYY - CITE	0 (Groundwater Sample	Σ	Beryllum	2.3	5 :	i i	
MWI-GWZ			_	Groundwater Sample	Σ	Cadmium	6.	5 :	1/8n	Ī
MW1-GW2				of care Company	*	Chromlum	0.7	>	ng/L	
MW1-GW2				ordina ionamono		Copper	4.1	>	1/0n	. •
MW1-GW2			_	Groundwater Sample		peol	7.7	S	UB/L	
MW1-GW2			06-Fob-90	Groundwaler Sample		N The state of the	-	E	ח/טח	0.2
MW1-GW2	_		06-Fob-90	Groundwater Sample		A I D A I D	3.50		1/0n	40
MW1-GW2	0-0	YES	06-Fob-90	Groundwater Sample	≨ :	MICKE	2.2.2	=		
MW1-GW2	_		06-Fab-90	Groundwater Sample		muliono o	- 4	-	1/00	
MWI-GWD			6-Fob-90	Groundwater Sample		PAIN F	9 6	-		
MW1-GW2		3 YES	6-Feb-90	Groundwater Sample	≨ :	maildin 7:55				
	(200	Orange Comple	2	2117			1	

RICKENBACKERANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - GROUNDWATER

SAMPLE #	GRID # FILTERED	YES/NO FILTERED	DATE	DESCRIPTION C.	CATEGORY	ANALYSIS	RESULTS QUALIFIER	aUALIFIER	UNITS	DETECTION
CWD CWN	B-7	CZ	07 - Fab90	Groundwater Sample	≨	Antimony	40.2	D	ug/L	09
CWD-CWM	B-7	2	07-Fab-90	Groundwater Sample	Σ	Arsenic	6.5	B	UB/L	10
MWD - GWD	8-7	2	07-Fob-90	Groundwater Sample	Σ	Beryllium	2.3	D	ng/L	S
WWD-GWD	8-7	2	07-Fob-90	Groundwater Sample	Σ	Cadmlum	1.9	2	ng/L	ις.
MWD - GWD	8-7	ON	07-Fob-90	Groundwater Sample	×	Chromlum	ន		Ug/L	10
WWZ - GWZ	B-7	OX	07 - Fob-90	Groundwater Sample	×	Coppor	101		ng/L	22
WWD-GWM	8-7	ON	07-Fob-90	Groundwater Sample	Σ	Load	84.5	S	ng/L	n ·
MW2-GW2	8-7	2	07-Fob-90	Groundwater Sample	Σ	Morcury	0.1	>	UB/L	0.2
MWZ-GWZ	8-7	9	07 - Fob -90	Groundwater Sample	Σ	Nickol	25		UB/L	40
MW2 - GW2	8-7	9	07-Feb-90	Groundwater Sample	Σ	Solonium	1.9	BW	UB/L	in ,
MW2 - GW2	8-7	02	07-Fob-90	Groundwater Sample	Σ	SNer	9.9	5	ng/L	9
MWZ-GWZ	8-7	9	07-Feb-90	Groundwater Sample	Σ	Thallium	6.0	Þ	ח/סח	9
WW2-GW2	8-7	Q	07-Fob-90	Groundwater Sample	Σ	Zinc	340		ng/L	50
MW2-GW2	8-7	YES	07-Fob-90	Groundwater Sample	Σ	Anlimony	34.6	5	ng/L	09
WWD-GWM	8-7	YES		Groundwater Sample	Σ	Arsonic	4.2	BW	ng/L	9
WWD-GWD	8-7	YES		Groundwater Sample	×	Beryllum	3.8	>	ng/L	S
WWO-GWA	B-7	YES		Groundwater Sample	Σ	Cadmlum	4.8	>	ng/L	S
MW2 - GW2	8-7	YES		Groundwaler Sample	Σ	Chromlum	9.6	>	ng/L	9
MW2-GW2	8-7	YES		Groundwater Sample	×	Coppor	ຍ	>	1/6n	25
MWZ-GWZ		YES		Groundwater Sample	Σ	Load	4	≥	7/Bn	ro ,
MW2 - GW2		YES		Groundwater Sample	×	Morcury	0.1		1/6n	0.5
MW2 - GW2	B-7	YES		Groundwater Sample	Σ	Nickel	31.1	5	ng/L	40
MW2-GW2		YES	07-Feb-90	Groundwater Sample	Σ	Solonium	4.)	ug/L	2 (
MW2-GW2	B-7	YES	07-Feb-90	Groundwater Sample	₹	Sivor	8.9	o :	ا ۱۹۸۰	2 5
MW2-GW2		YES	07-Feb-90	Groundwaler Sample	Σ	Thallium		> ;	ug/L	2 8
MW2-GW2	B-7	YES	07-Fob-90	Groundwater Sample	Σ	Zinc	18	3	Ng/L	02

RICKENBACKERANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - GROUNDWATER

Z	09	-	· ·		2	52	6	0.2	0	2	9	9	20	8	9	2	S	9	25	n	0.2	ç	S	2	2	8	
LIMIT	φ.				•	••		0																			
DE	J/Br	7/Br	7/01	٦/د	7/6	ug/L	D/L	9/٢	۵/۲	9/	1/8	1/0	10/L	1/6	ng/L	18/F	J/Br	ng/L	UB/L	10/F	ug/L	UB/L	J/Bn	Ug/L	ng/L	1/60	
UNITS	ân	3	3n	ñ	ñ	5	ž	5	Š	כ		2	2	2	2	٠	_		_	_							
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QUALIFIER																									_		
	40.2	5,9	2.3	6.1	8.7	7	9.7	0.1	23.6	1.7	9	0	35	34.6	2	8	4.8	9.8	9	5.3	0.1	31.1	1.4	8.9	6.0	17	
RESULTS																							_	_	_		
	Antimomy	Arsonic	Boryllium	Cadmium	Chromium	Copper	1000	Moreury	Nickol	Colonium	Care	10 11 11 11	7100	Antimony	Arende	Revilling	Cadmin	Promine.	Copper) and	March	Nicke	Selenium	SNor	Thellium	Zinc	
SIS	Anti	Ā	Bar	Č	240	, C)	Σ		0	5	É	=	And	2	ď	ڻ ڏ	ל ל	5	,	2		Ü)	_		
ANALYSIS FOR																											
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CATEGORY																											
U	4	2	eld E	old m	old m	ojdwi	oldwi	oldmi	oldwi	oldwa	oldma	oldma	oldma	amplo	amplo	amplo	amplo	amplo	amplo	ampio	amplo	Groundwater Sample	Groundwater Sample	Groundwaler Sample	Groundwater Sample	Groundwater Sample	Sampie
TION		ator Su	ator 58	alor Sa	alor Sa	alor So	rator St	rator St	rator Si	valor Si	vator S	valor S	valor S	valor S	valor S	valer S	valor S	wator S	wator S	water	Waler	wator	water	walor	water	water	water
DESCRIPTION		Groundwater Suttiple	Groundwater Sample	Groundwater Sample	Groundwaler Sample	Groundwater Sample	Groundwaler Sample	Groundwater Sample	Groundwater Sample	Groundwater Sample	Groundwater Sample	Groundwater Sample	Groundwaler Sample	Groundwaler Sample	Ground	Ground	Ground	Ground	Ground	Groundwater Sampie							
0.6	9	5	ฮิ	ō	Ö	Ö	ပ	O	O	O	O	Ö	O	O	0	0	Ü	_	_		_	_	_				
		06-0	06-0	06-0	06-0	990	P-90	06-q	06-q	06-q	06-Feb-90	06-Fob-90	06-Fob-90	06-Feb-90	06-Feb-90	06-Feb-90	06-Feb-90	06-Feb-90	06-Fob-90	OR-Fab-90							
DATE		06-Feb-90	06 - Feb - 90	06-Fab-90	06-Feb-90	06-Feb-90	06-Fob-90	06-Fab-90	06-Feb-90	06-Feb-90	06-Fob-90	06-Fab-90	06-Fob-90	06-Fob-90	06-Fob-90	06-Feb-90	06-Fob-90	06-F	06 - F	06 - F	96 - F	96 - F	06 - F	06-F	96 - F	96 - F	96-6
0 ED	:::	_ 	<u> </u>		9 2		9	9	9	õ	9	0	02	2	YES	YES	YES	YES	YES	YES	YES	YES	YES	YES	YES	YES	\ \ \ \ \ \ \ \ \ \
YES/NO.																					_	_					_
		0-4	0-1	C-4	C-4	C-4	4-0	4-0	014	014	4	4-0	1	0-1	0-14	0-14	4-0	C-4	0	CIA	0	C	0-14	C	C	0-14	(
GRID	* _	2	2	2	N	2	2	2	2	2	1 0	2 1	2 5	2 !	2	2	2	2 !	2	2	2	\$	\$	\$	\$	32	-
יובי		MW3-GW2	MW3-GWZ	MW3-GWZ	MW3-GWZ	MW3-GW2	MW3 - GW2	ZW3 - GWZ	MW3 - GW2	200-200	CWD - CWM	TWY - GWD		2WD - 5WW	WW3 - GWZ	ZWD-EWM	CWD - CWM	CWD - EWM	ZWD-EWM	MW3-GWZ	MW3-GW2	WWG - GWZ	MW3 - GW2	MW3 - GW2	MW3 - GW2	MW3-GWZ	
SAMPLE		×	W	×	×	3	×	3	3	***	777	AAA	2 2	2	\ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \	×	2	3	3	Š	ž	. .	 	: \$	2	Σ	: :
		_	_	_		_				_		_															

RICKENBACKERANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS — GROUNDWATER

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SAMPLE . GRID	_		֡	The second secon						
	1						;	=	,	90
	6	Ç	06-Fab-90	Groundwater Sample	Σ	Anlimony	34.6	D	מט/ר	2
WW4 - CW	3 6	2	06-Fab-90	Groundwater Sample	Σ	Arsonic	0.00	0	na/r	2
MW4 - GW	2 0	2	06-Fab-90	Groundwater Sample	Σ	Boryllum	2 4	5 =	no/L	S
MW4-CW1	7 6	2	06-Fab-90	Groundwater Sample	Σ	Cadmium	0, 6)	Non.	9
MW4 - GW1	2-0	2 2	06 - 69 - 90	Groundwater Sample	Σ	Chromlum	2 1		9 5	25
MW4 - GWI	B-2	2 5	06-69-90	Groundwater Sample	×	Copper	73	C	7/02	7 (7
MW4 - GW1	B2	2 :	00-1-00	Octobration Sample	Σ	Load	27.1	0 =	1/6:	0
MW4 - GW1	0-2	2 !	06-08-90	Semolo Samolo	Σ	Morcary	0.7		1/00	2 9
MW4 - GW1	8-2	0	00-100-00		2	Nickel	73		1/Bn	2 1
MW4 - GW1	8-2	0 N	06-Fob-90	Groundwater Sample	2	Selenium	1.5	BW	UB/L	c ;
WWA - GW1	B-2	92	06-Fob-90	Groundwater Sample	2 3	SNot	8.9	2	ug/L	10
44W4 - GW1	8-2	2	06-Fob-90	Groundwater Sample	ξ:	Thallitm	0.9	5	ug/L	10
TANA CONT	10	92	06-Fob-90	Groundwater Sample	ε:	Zing			Ug/L	20
MW4 - GW1	8-2	02	06-Fob-90	Groundwaler Sample	ξ :	Antimomy		D	NB/L	9
MWA - GW1	8-2	YES		Groundwater Sample	Ξ :	Arsonie			ng/L	9
WWA - GW1	B-2	YES	06-Feb-90	Groundwater Sample		Millim		כ	1/6n	S
WW4 - GW1	8-2	YES		Groundwater Sample		Cadmin Cadmin	1.9		Ug/L	S
WAY - GWI	8-2	YES	06-Feb-90	Groundwaler Sample		Chromium		>	UD/L	9
100 - CM1	11-2	YES	06-Fob-90	Groundwater Sample		Jedao C		>	ng/L	25
TWO - NAME	8-2	YES	06-Fob-90	Groundwaler Sample		ped 1	_		UB/L	n
100	B-2	YES	06-Fob-90	Groundwater Sample		Moreille	_	ס	Ug/L	0.5
200	, ,	YES	06-Fob-90	Groundwater Sample				>	UG/L	40
MW4-CW1	9 (- >		Groundwater Sample	×	Nickel	,	-	1/0/1	3
MW4-GW1	7-0	500		Grandwaler Sample	Σ	Solonium	_		1 2 2	10
MW4 - GW1	B-2	YES	_	olomes relember		SNer		5 :	1	2
MW4 - GW1	B-2	YES	_	Groundwatel Sample		Thallium	0.0		UB/L	2 8
MW4 -GW1	8-2	YES	_	Groundwater Sample		Zinc	10	-B	√2/L	2
		1	1	CICE OF POLICE CO.	2					

RICKENBACKERANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - GROUNDWATER

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SAMPLE /	GRID (GRID # FILTERED	DATE	DESCRIPTION	CATEGORY	ANALYSIS FOR	RESULTS	QUALIFIER	UNITS	DETECTION
MW6-GW1	E-5	2	07-Fob-90	Groundwater Sample	Σ	Antimony	35.3	Þ	na/L	9
MW6-GW1	E5	0	07-Feb-90	Groundwater Sample	Σ	Arsonic	9.2	BW	ממ/ר	2
MW6-GW1	E-5	ON N	07-Feb-90	Groundwater Sample	Σ	Boryllium	3.9	כ	ח/הח	10
MW6-GW1	5-5	ON N	07-Fob-90	Groundwater Sample	×	Cadmium	4.9)	No/L	, ru
MWG-GW1	E-5	9	07-Fob-90	Groundwater Sample	Σ	Chromium	32,6	1	na/L	0
MW6-GW1	E-5	2	07-Feb-90	Groundwater Sample	Σ	Copper	101		מא/ר	25
MW6-GW1	E-5	ON ON	07-Feb-90	Groundwater Sample	Σ	Load	84		UB/L	ť
MWG-GW1	E-5	02	07-Fob-90	Groundwater Sample	Σ	Morcury	0.1	D	UB/L	0.2
MW6-GW1	E-5	ON.	07-Fob-90	Groundwater Sample	Σ	Nickol	83.6		ug/L	40
MW6-GW1	E-5	O _N	07-Fob-90	Groundwaler Sample	×	Solonium	13	BW	UB/L	S
MW6-GW1	E-5	O _N	07-Feb-90	Groundwater Sample	Σ	SNor	9.1	D	7/Bn	10
MW6-GW1	E-5	9	07-Feb-90	Groundwater Sample	Σ	Thallium	6.0	2	Ug/L	01
MW6-GW1	E-5	<u>Q</u>	07-Feb-90	Groundwater Sample	Σ	Zlnc	455		ug/L	20
MW6-GW1	E-5	YES	07-Fob-90	Groundwater Sample	Σ	Antimony	40.2	5	Ug/L	9
MW6-GW1	E-5	YES	07-Feb-90	Groundwater Sample	Σ	Arsenic	1.5		ng/L	10
MW6-GW1	E-5	YES	07-Fob-90	Groundwater Sample	Σ	Beryllium	2.3	2	UD/L	S
MW6-GW1	E-5	YES	07-Feb-90	Groundwater Sample	Σ	Cadmium	1.9	2	ug/L	S
MW6-GW1	E-5	YES	07-Feb-90	Groundwater Sample	Σ	Chromlum	8.7	5	ug/L	10
MW6-GW1	E-5	YES	07-Feb-90	Groundwater Sample	Σ	Copper	4.1	2	La/L	25
MWG-GW1	E-5	YES		Groundwater Sample	Σ	Load	3.1	7	ug/L	C
MW6-GW1	E-5	YES		Groundwater Sample	Σ	Moteury	0.1	>	Ug/L	0.2
MW6-GW1	E-5	YES	07-Feb-90	Groundwater Sample	Σ	Nickel	23.6	5	UQ/L	40
MW6-GW1	E-5	YES		Groundwalor Samplo	Σ	Solonium	1.7	2	מש/ר	v
MWG-GW1	E-5	YES	07-Fob-90	Groundwater Sample	Σ	SNor	9.9	2	ug/L	10
MW6-GW1	л 2-	YES	07-Feb-90	Groundwater Samplo	Σ	Thallium	6.0	5	UB/L	10
MW6-GW1	E-5	YES	07-Feb-90	Groundwater Sample	Σ	Zing	S	B	1/00	20

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SAMPLE # GRID # FILTERED	GRID # F	YES/NO FILTERED	DATE	DESCRIPTION	CATEGORY	ANALYSIS FOR	RESULTS	aUALIFIER	UNITS	DETECTION
MW7_CW1	C	2	07 Est 00			i- v	976	=		3
100	0 0	2 2	04 1 20	Cloudonaled Sample	Ξ:	Annual An	2.0	>	7/87	8 5
WAY I CAN	20	2	08-08-70	Groundwater Sample	Ξ	Arsonic	6.71		ng/L	2
MW7-GW1	0-8	<u>Q</u>	07-Fob-90	Groundwater Sample	Σ	Boryllium	3.8	5	ng/L	S
MW7 - GW1	D-8	9	07-Feb-90	Groundwater Sample	Σ	Cadmlum	4.8		Ug/L	S
MW7-GW1	D-8	0X	07-Feb-90	Groundwater Sample	2	Chromlum	27		Ug/L	10
MW7-GW1	0-8	2	07-Feb-90	Groundwater Sample	Σ	Copper	31		ug/L	25
MW7-GW1	D-8	2	07-Feb-90	Groundwater Sample	Σ	Load	25.8	+	תמ/ר	n
MW7 - GW1	0-8	02	07-Feb-90	Groundwater Sample	₹	Morcury	0.1	כ	Ug/L	0.2
· MW7-GW1	D-8	02	07-Feb-90	Groundwater Sample	Σ	Nickel	31.1	>	Ng/L	40
MW7-GW1	D-8	9	07 - Feb-90	Groundwater Sample	Σ	Selenium	1.4	}	US/L	2
MW7-GW1	0-8	9	07-Fab-90	Groundwater Sample	Σ	Sivor	6.9	2	ng/L	2
MW7-GW1	D-8	9	07-Fob-90	Groundwator Samplo	Σ	Thallium	6.0	כ	ng/L	5
MW7-GW1	0 - 0	02	07-Feb-90	Groundwater Sample	Σ	Zinc	168		ng/L	20
MW7-GW1	D-8	YES	07-Feb-90	Groundwater Sample	Σ	Antimony	40.2	D	ng/L	09
MW7-GW1	0-8	YES	07-Fob-90	Groundwater Sample	Σ	Arsonic	6.9	B	ng/L	01
MW7-GW1	D-8	YES	07-Fob-90	Groundwaler Sample	×	Beryllium	2.3	כ	ng/L	\$
MW7 - GW1	D-8	YES	07-Feb-90	Groundwater Sample	×	Cadmium	1.9	כ	Ug/L	S
MW7-GW1	0-8	YES	07-Feb-90	Groundwater Sample	Σ	Chromlum	7.8	D	ng/L	10
MW7-GW1	0-8	YES	07-Fob-90	Groundwater Sample	Σ	Copper	4.1		UB/L	25
MW1-GW1	0-0	YES	07-Feb-90	Groundwater Sample	Σ	Load	4.9	2	UB/L	n
MW7-GW1	D-8	YES	07 - Feb-90	Groundwater Sample	Σ	Morcury	0.1	ס	ng/L	0.2
MW7-GW1	D-8	YES	07-Fab-90	Groundwater Sample	Σ	Nickel	23.6	>	Ug/L	40
MW7-GW1	0-8	YES	07-Feb-90	Groundwater Sample	Σ	Selenium	1.7	D	ng/L	2
MW7-GW1	0-8	YES	07-Fob-90	Groundwater Sample	Σ	SNor	9.9	D	ng/L	0
MW7-GW1	D-8	YES	07-Feb-90	Groundwater Sample	Σ	Thallium	6.0	כ	ng/L	0
MW7 - GW1	0-8	YES	07-Fob-90	Groundwater Sample	W	Zinc	8	BJ.	UB/L	20

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8	0	S	r.	10	25	6	0.2	40	5	10	10	20	9	10	2	13	10	25	6	0.2	40	S	0	10	ć
UG/L	ח/הח	UB/L	ug/L	ug/L	ug/L	Ug/L	UB/L	ug/L	UB/L	Ug/L	ug/L	Ug/L	UB/L	UB/L	ng/L	ng/L	UB/L	UG/L	UB/L	UB/L	UB/L	ng/L	Ug/L	UB/L	1/0::
5	BW	8	5				D		3	D	כ		ס	BW	כ	ס	>	>		ס	5	כ	כ	כ	-
40.2	1.7	C	1.9	65	135	58.1	0.1	134	1.7	6.6	6.0	642	34.6	3,1	3.6	4.8	9.0	9	ဗ	0.1	31.1	4.1	6.9	6.0	2.0
Antimony	Arsonic	Boryllum	Cadmlum	Chromlum	Copper	Load	Marcury	Nickol	Solonium	Sivor	Thallium	Zinc	Antimony	Arsonic	Boryllium	Cadmlum	Chromium	Coppor	Load	Morcury	Nickel	Solonium	SNor	Thallium	Zine
Σ	×	Σ	×	×	Σ	×	¥	×	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	×	Σ	Σ	æ	Σ	Σ	Σ	Σ
Groundwater Sample																									
07-Fob-90	07-Fob-90	07-Feb-90	07-Fob-90	07-Fob-90	07-Feb-90	07-Feb-90	07-Feb-90	07-Feb-90	07-Feb-90	07-Feb-90	07-Fob-90	07-Feb-90	07-Feb-90	07-Fob-90	07-Fob-90	07-Feb-90	07 - Fob - 90	07-Feb-90	07-Fob-90	07-Fob-90	07 - Fob-90	07 - Fob - 90	07 - Fob-90	07-Fob-90	07 - Fab - 90
2	ON N	O _N	2	ON	9	0 2	9	9	0	<u>0</u>	0 2	0	YES												
C-10	C-10	C-10	C-10	C-10	C-10	01-10	C-10	0-10	C-10	01-10	0-10	C-10	0-10	C-10	0-10	C-10	C-10	C-10	01-10	C-10	C-10	C-10	0-10	C-10	C-10
MWB-GW1	MW8-GW1	MWB-GW1	MW8-GW1	MWB-GW1	MWB-GWI	MWB-GWI	MWB-GW1	MWB-GW1	MWB-GW1	MWB-GW1	MWB-GWI	MWB-GWI	MWB-GWI	MWB-GW1	MWB-GWI	MWB-GW1	MWB-GWI	MWB-GW1							

MW9-GW1	A-4	Q	16 - Fob - 90	Groundwater Sample	≥.	Antimony	40.2	כ	UB/L	9
MW9-GW1	¥-4	2	16-Fob-90	Groundwater Sample	Σ	Arsonic	8.4	BS	UB/L	0
MW9-GW1	A-4	2	16-Fob-90	Groundwater Sample	Σ	Boryllium	2.3	8	ng/L	ιO.
MW9-GW1	A-4	9	16-Fob-90	Groundwater Sample	Σ	Cadmlum	1.9	כ	ng/L	ທຸ
MW9-GW1	A-4	02	16-Fob-90	Groundwater Sample	Σ	Chromlum	24	!	ng/L	01
MW9-GW1	A-4	ON.	16-Fob-90	Groundwater Sample	Σ	Copper	15	a	ng/L	25
MW9-GW1	A-4	2	16-Fob-90	Groundwater Sample	Σ	Load	33.3		ا ع/۵۰	ח
MW9-GW1	A-4	02	16-Fob-90	Groundwater Sample	Σ	Morcury	0.1	5	ng/L	0.2
MW9-GW1	A-4	02	16-Fob-90	Groundwater Sample	Σ	Nickol	23.6		ug/L	04
MW9-GW1	A-4	ON N	16-Fob-90	Groundwater Sample	Σ	Solonium	1.7	§ :	מש/ך }	ຸດ ເ
MW9-GW1	A-4	0	16-Feb-90	Groundwater Sample	Σ	SNor	9.9	o :	ng/L	9 9
MW9-GW1	A-4	0	16-Fob-90	Groundwater Sample	Σ	Thallium	6.0	5	7/8n	2 5
MW9-GW1	A-4	0	16-Feb-90	Groundwater Sample	Σ:	Zinc	707	=	7/60	9
MW9-GW1	A-4	YES	16-Fob-90	Groundwaler Sample	٤ :	Anumony	7.0) (f	1/60	5 5
MW9-GW1	A-4	YES	16 - Feb - 90	Groundwater Sample	ξ 3	arillicon	3 6) =	1/8.7	i ru
MW9-GW1	A-4	YES	16-100-90	Groundwater Sample	ξ 3	minus of the second	0.4) =	1/01	, ru
MW9-GW1	A4	YES	16-Feb-90	Groundwaler Sample	ξ 3	This can		> =	1/83	5
MW9-GW1	A-4	Y 2	06-00-01	Groundwater sample	ξ 3	red of C	4	5 3	1/00	25
MW9-GW1	A .	7 17	06-001-01	Groundwarer Sample	2	Tree -		7	1/60	n
MW3-CWI	A-4	מוא	06-001-01	Grownian Sample	E 2	March	100	2	1/67	0.2
MW9-GW1	A-4	YES	16-100-90	Groundwaler sample	ξ 2	19319	23.6) =	1/07	40
MW9-GW1	A-4	YES	06-001-01	Groundwater Sample	ξ :	in land	7 7	2	1/01	i ru
MW9-GW1	A-4	YES	16-100-90	Groundwater Sample	ξ:	IIIDIII IIDI		=	7/01	5
MW9-GW1	A-4	YES	16-rob-90	Groundwaler Sample	٤ 3	I STATE	0.00) =	1/6:1	5 5
MW9-GW1	A-4	YES	16-Feb-90	Groundwater Sample	Σ :	LUGURUL	D. 0.	2	1/60	2 6
MW9-GW1	A-4	YES	16-Fob-90	Groundwater Sample	Σ	Zinc	13	8	ng/t	2
280	Ċ		OB - Fab - 90	Groundwater Sample	>	Acotono	20	3	ug/L	5
MW1-GW2	0 0		06-Fob-90	Groundwater Sample	>	2-Butanone	100	ס	ug/L	10
CW1 - CWN	ט נ		06-Feb-90	Groundwater Sample	>	Virryl Acetate	S	O	ug/L	10
WW.	ט ני		06-Fab-90	Groundwater Sample	>	Bonzene	260	٥	ug/L	10
WW1-GW2	ט ני		06-Fab-90	Groundwater Sample	>	2-Hexanone	S	>	ng/L	10
MW1-GW2	9-0		06-Fob-90	Groundwater Sample	>	4-Mollyl-2-pantanone	S	>	ng/L	10
MW1-GW2	0-0		06-Fab-90	Groundwater Sample	>	Ethylbonzono	110		ng/L	9
MW1-GW2	0-0		06-Fob-90	Groundwater Sample	>	m/p-Xylono	35		ng/L	0
MW1-GW2	0-C		06-Fob-90	Groundwater Sample	>	o-Xylana	99		na/r	2
CWD CWN	B-7		07 - Fab-90	Groundwater Sample	>	Acotono	5		vg/L	01
210-2111	1 0		07 - Fah- 90	Groundwaler Sample	>	2—Butanone	\$	Э	ug/L	5
MW2 - GW2	8-7-8		07 - Feb - 90	Groundwater Sample	>	Viryi Acotato	ß	כ		10
MW2 - GW2	B-7		07-Feb-90	Groundwater Sample	>	2-Hoxanone	S	>	. ug/L	0,
	. :									

SAMPLE # GRID		YES/NO F FILTERED	DATE	DESCRIPTION CA	сатевоях	ANALYSIS FOR	HESULTS	HESULTS QUALIFIER	UNITS	LIMIT
MW3-GW2	2.0		06-Fob-90	Groundwater Sample	>	Acolono	100	D	ng/L	ç
MW3-GWZ	0-14		06-Feb-90	Groundwater Sample	>	2-Bulanono	8	D	ug/L	10
MW3-GWZ	C-4		06-Fob-90	Groundwater Sample	>	Vinyl Acotato	S	э	ug/L	2
MW3-GWZ	C-4		06-Fob-90	Groundwater Sample	>	Trichloroethene	7		. ug/L	10
MW3-GW2	C-4		06-Feb-90	Groundwater Sample	>	2-Hexanone	22	כ	ng/L	10
MW3-GW2	C-4		06-Fob-90	Groundwater Sample	>	4-Mothyl-2-pentanone	50	D	∪g/L	10
IMO - NMM	2		06 - Feb - 90	Groundwoter Semula	>	Acalona	100	3	1/50	101
MW4 - GW1	2 2		06-Fab-90	Groundwater Sample	>	2-8	8)	UQ/L	10
MW4-GW1	B-2		06-Fob-90	Groundwaler Sample	>		20	כ	UB/L	10
MW4-GW1	8-2		06-Feb-90	Groundwater Sample	>	2-Hexanone	20	כ	ng/L	10
MW4-GW1	B-2		06-Fob-90	Groundwater Sample	^	4-Mathyl-2-pantanono	50	ם	√Bn	10
MWG-GW1	E-6		07-Fob-90	Groundwater Sample	>	Acetone	8	>	7/Bn	0
MW6-GW1	E-6		07-Feb-90	Groundwater Sample	>	Trans-12-Dichloroothono	ю		ng/L	10
MWG-GW1	. E-6	.!	07-Feb-90	Groundwater Samplo	>	2-Butanone	100	כ	ng/L	10
MWG-GW1	E-6		07-Feb-90	Groundwater Sample	>	Vinyl Acotato	20	כ	ng/L	10
MW6-GW1	E-6		07-Fob-90	Groundwater Sample	>	Trichloroethene	78		ng/L	0
MW6-GW1	E-6		07-Feb-90	Groundwater Sample	>	2-Hexanone	20	ב	ng/L	2
MW6-GW1	E-6		07 - Feb-90	Groundwater Sample	>	4-Methyl-2-pentanone	20	ם	ng/L	10

SAMPLE /	GRID /	YES/NO FILTERED	YES/NO FILTERED DATE	DESCRIPTION	САТЕВОВУ	FOR	RESULTS al	aualifier.	UNITS	DETECTION
MW7-GW1	0-7		07-Fob-90	Groundwater Sample	>	Chioromothano	901	ס	UB/L	0
MW7-GW1	0-7		07-Fob-90	Groundwater Sample	>	Bromoothane	001	כ	ng/L	10
MW7-GW1	2-0		07-Fob-90	Groundwaler Sample	>	Vinyl Chlorido	100	ס	ng/L	0
MW7-GW1	0-7		07-Fob-90	Groundwater Sample	>	Chloroethane	100	2	ng/L	9
MW7-GW1	0-7		07-Feb-90	Groundwater Sample	>	Mothylana_Chlorida	20	3	ng/L	0
MW7-GW1	0-7		07-Fob-90	Groundwater Sample	>	Aroloin	8	2	UB/L	9
MW7-GW1	0-7		07-Fob-90	Groundwater Sample	>	Acotono	1000	3	UB/L	0
MW7-GW1	D-7		07-Fob-90	Groundwater Sample	>	Arylonitile	8)	1/6n	9
MW7-GW1	2-0		07-Fob-90	Groundwater Sample	>	Carbon Disulido	100	כ	ng/L	9
MW7-GW1	D-7		07-Feb-90	Groundwater Sample	>	Trichlorofluoromothano	100	כ	ng/L	0
MW7-GW1	0-7	.:	07-Fob-90	Groundwater Sample	>	1,1-Dichloroathana	20	D	ng/L	0
MW7-GW1	0-7		07-Fob-90	Groundwater Sample	>	1,1-Dichloroothano	20	5	U9/L	0
MW7-GW1	0-7		07-Fob-90	Groundwater Sample	>	Trans-12-Dichloroathana	20	D	ng/L	<u>0</u>
MW7-GW1	0-7		07-Fob-90	Groundwater Sample	>	Chloroform	20	>	n8/r	0
MW7-GW1	0-7		07-Feb-90	Groundwater Sample	>	1,2-Dichloroathane	20	5	ug/L	2
MW7-GW1	0-7		07-Fob-90	Groundwater Sample	>	2-Butanone	9	D	ng/L	9
MW7-GW1	D-7		07-Feb-90	Groundwater Sample	>	1,1,1-Trichloroothano	20	כ	UB/L	9
MW7-GW1	0-7		07-Fob-90	Groundwater Sample	>	Carbon Totrachloride	20)	UB/L	2
MW7-GW1	0-7		07-Fob-90	Groundwater Sample	>	Viryl Acotato	200	D	ng/L	0
MW7-GW1	2-0		07-Fob-90	Groundwater Sample	>	Bromodichloromethane	22	>	ng/L	9
MW7-GW1	2-0		07-Fob-90	Groundwater Sample	>	1,2-Dichlorpropano	S	כ	ng/L	0
MW7-GW1	2-0		07-Feb-90	Groundwater Sample	>	cls-1,3-Dichlorpropene	20	כ	ng/L	01
MW7-GW1	2-0		07-Feb-90	Groundwater Sample	>	Trichloroothono	20	D	ng/L	0
MW7-GW1	0-7		07-Fob-90	Groundwater Sample	>	Bonzene	200		ng/L	9
MW7-GW1	1-0		07-Fob-90	Groundwater Sample	>	Dibromochlowmathana	20	ס	ng/L	9
MW7-GW1	D-7		07-Feb-90	Groundwater Sample	>	1,1,2 -Trichloroothano	20	2	Ug/L	10
MW7-GW1	0-7		07-Feb-90	Groundwater Sample	>	trans-1,3-Dichlorpropene	S	D	ng/L	10
MW7-GW1	D-7		07-Fob-90	Groundwalor Sample	>	2-Chloroethylvinylether	8	ɔ :	J/Bn	10
MW7-GW1	0-7		07-Fob-90	Groundwater Sample	>	Вготобогт	20	>	7/Bn	01
MW7-GW1	2-0		07-Fob-90	Groundwater Sample	>	2-Hexanone	200	> :	ng/L	10
MW7-GW1	D-7		07-Fob-90	Groundwater Sample	>	4-Mothyl-2-pontanone	200	5	7/Bn	9
MW7-GW1	0-7		07-Fob-90	Groundwator Sample	>	Totrachloroothone	ୟ	>	ng/L	9
MW7-GW1	0-7		07-Fob-90	Groundwater Sample	>	1,122-Tatrachloroethene	20	D	ng/L	9
MW7-GW1	0-7		07-Fob-90	Groundwater Sample	>	Toluene	20)	ng/L	9
MW7-GW1	2-0		07-Feb-90	Groundwater Sample	>	Chlorobenzene	22	>	7/8n	0
MW7-GW1	0-7		07-Feb-30	Groundwater Sample	>	Elhylbanzona	8		ug/L	5
MW7-GW1	D-7		07-Feb-90	Groundwater Sample	>	Styrono	20	בי כ	ng/L	10
MW7-GW1	0-7		07-Feb-90	Groundwater Sample	>	m/p-Xylono	21	7	ng/L	10
MW7-GW1	0-7		07-Feb-90	Groundwater Sample	>	o-Xyleno	20		ug/L	9
MW7-GW1	0-7		07-Fob-90	Groundwater Sample	> :	1,3-Dichlorobonz and	05	> :	ug/L	0 5
MW7-GW1	0-7		07-Fob-90	Groundwater Sample	>	1,2/1,4-Dichlorobenzene	50]	5	NB/L	01

WM9-CWII C = 10 OT = 66 – 90 Groundwater Sample V Motifylana Chioda 150 UJ ug/L WM9-CWII C = 10 OT = 66 – 90 Groundwater Sample V A-Mashy-2-Institution 100 UJL U									
C-10	0,10	07 - Fab-90	Groundwater Sample	>	Methylene Chloride	1 0	3	ug/L	10
C-10 07-Feb-90 Groundwater Sample V A-Buthrone 100 U u C-10 07-Feb-90 Groundwater Sample V 4-Malty-2-profutnore S0 U u u C-10 07-Feb-90 Groundwater Sample V 4-Malty-2-profutnore S0 U u	0 - 10		Groundwater Sample	>	Acetone	8	כ	ug/L	2
C-10	C-10		Groundwater Sample	>	2-Butanone	8	5 :	ng/L	₽ !
C	C-10		Groundwater Sample	>	Vinyl Acetate	20	o :	ng/L	2 9
C	C-10		Groundwater Sample	>	2-Hoxanone	20	5:	מאר	2 9
A-4 16-Feb-90 Groundwater Sample V Cabbanone 100 U			Groundwater Sample	>	4-Mathyl-2-pantanone	20		1767	2
A-4 16 – Feb – 90 Cloundwater Sample V E-Buhanne 10 U				>	Acoton	100	כ	no/L	10
A-4 19-Feb-90 Groundwater Sample V C-Hozanore SO U A-4 16-Feb-90 Groundwater Sample V 4-Menthyl-2-pointenore 50 U <td< td=""><td></td><td>16-160-190</td><td>Groundward Sample</td><td>- ></td><td>2-Burbane</td><td>90</td><td></td><td>no/L</td><td>5</td></td<>		16-160-190	Groundward Sample	- >	2-Burbane	90		no/L	5
A-4 16-Feb-90 Groundwater Sample V A-Methyl-2-pontentone 50 U U A-4 16-Feb-90 Groundwater Sample V 4-Methyl-2-pontentone 50 U			Groundwaler Sample	> >	Mend Acatala	2	5 =	1/00	0,
A-4 16-Feb-90 Groundwater Sample V 4-Methyl-2-pontations CO C			Groundwater Sample	> 3	VIIIVI ACOIDIO	2 2) =	1/0	
A-4 16-Feb-90 Groundwater Sample V Methylene, Chloidde S UJ U U U U U U U U			Groundwater Sample	> >	2 - Mathal - 2 - Maranone	8 6) D	1/60	2 2
C			Groundwalor Sample	>	4 - Monyi-z - pontanono	3	,	7,00	
C-6	\$	07 - Fab - 60	Groundwater Sample	>	Methylene Chloride	.c	3	ug/L	5
C6	2 8	07-154-90	Grandwater Sample	>	Acotono	8	כ	Ug/L	5
O7 - Fab - 90 Groundwater Sample V Viryl Acatalo 50 U U	2	00 11 10	olding loss to the control of	> >	2-Rutanona	50		no/L	0
O7 - Feb-90 Groundwater Sample V 4 - Methyl-2 - pentanen 50 U U	<u>.</u>	07-F9b-90	Groundwaler Sample	> >	Very Acotalo	20.00	5	1/00	10
CG-6 O7-Feb-90 Groundwater Sample V 4-Methyl-2-pontanion SO U C-6 06-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidire 20 U C-4 06-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidire 20 U E-5 06-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidire 20 U E-5 07-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidire 20 U C-10 07-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidire 20 U C-10 07-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidire 20 U C-10 07-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidire 20 U A-4 16-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidire 20 U B-7 17-Ocl-91 Groundwater Sample SV 3,3'-Dichloroborzidire 20 U B-	Q	07 - Feb - 90	Groundwater sample	> :	Alliyi Acelaio	3 5	=		5
C-6 06-Feb-90 Groundwater Sample SV 3,3'-Dichlorobarzidine 20 U L B-7 07-Feb-90 Groundwater Sample SV 3,3'-Dichlorobarzidine 20 U U U B-2 06-Feb-90 Groundwater Sample SV 3,3'-Dichlorobarzidine 20 U U U E-5 07-Feb-90 Groundwater Sample SV 3,3'-Dichlorobarzidine 20 U U U C-10 07-Feb-90 Groundwater Sample SV 3,3'-Dichlorobarzidine 20 U U C-10 07-Feb-90 Groundwater Sample SV 3,3'-Dichlorobarzidine 20 U C-10 07-Feb-90 Groundwater Sample SV 3,3'-Dichlorobarzidine 20 U A-4 16-Feb-90 Groundwater Sample SV 3,3'-Dichlorobarzidine 20 U B-7 17-Oct-91 Groundwater Sample SV 3,3'-Dichlorobarzidine 20 U B-7 17-Oct-91 Gr	Ø.	07-Fob-90	Groundwater Sample	> >	ononalmon - S-livilor	3 6) >	7/00	5 2
C-6 O6 – Feb-90 Groundwater Sample SV 3,3"-Dichloroborzidine 20 U u B-7 07 – Feb-90 Groundwater Sample SV 3,3"-Dichloroborzidine 20 U u C-4 06 – Feb-90 Groundwater Sample SV ALL SEMI-VOLATILES ND J u E-5 07 – Feb-90 Groundwater Sample SV 3,3"-Dichloroborzidine 20 U U C-10 07 – Feb-90 Groundwater Sample SV 3,3"-Dichloroborzidine 20 U U C-10 07 – Feb-90 Groundwater Sample SV 3,3"-Dichloroborzidine 20 U C-10 07 – Feb-90 Groundwater Sample SV 3,3"-Dichloroborzidine 20 U A-4 16 – Feb-90 Groundwater Sample SV 3,3"-Dichloroborzidine 20 U B-7 17 – Oci-91 Groundwater Sample SV 2-Methythaphthalene 20 U B-7 17 – Oci-91 Groundwater Sample SV	81	0/101-30	Groundwaler Sample	>	Manual - 2 - Manual - 1				
B-7 07 – Fob–90 Groundwater Sample SV 3,3" – Dichloroborzidine 20 U C-4 06 – Fob–90 Groundwater Sample SV ALL SEMI–VOLATILES ND J · · · · E-5 06 – Fob–90 Groundwater Sample SV 3,3" – Dichloroborzidine 20 U C-10 07 – Fob–90 Groundwater Sample SV 3,3" – Dichloroborzidine 20 U C-10 07 – Fob–90 Groundwater Sample SV 3,3" – Dichloroborzidine 20 U C-10 07 – Fob–90 Groundwater Sample SV 3,3" – Dichloroborzidine 20 U A-4 16 – Fob–90 Groundwater Sample SV 3,3" – Dichloroborzidine 20 U SV 3,1" – Dichloroborzidine 20 U U TR-Fob–90 Groundwater Sample SV 3,3" – Dichloroborzidine 20 U RB-7 17 – Oct–91 Groundwater Sample SV 2,3" – Dichloroborzidine 20 U RB-7 17 – Oct–91	MW1-GW2 C-6				3,3'-Dichlorolxonzidine	20	ס	ng/L	10
C-4 06 – Feb–90 Groundwater Sample SV ALL SEMI-VOLATILES ND J B-2 06 – Feb–90 Groundwater Sample SV 3,3'-Dichloroborzidine 20 U E-5 07 – Feb–90 Groundwater Sample SV 3,3'-Dichloroborzidine 20 U C-10 07 – Feb–90 Groundwater Sample SV 2,-Methylnaphthalone 5 J C-10 07 – Feb–90 Groundwater Sample SV 3,3'-Dichloroborzidine 20 U C-10 07 – Feb–90 Groundwater Sample SV 3,3'-Dichloroborzidine 20 U A-4 16 – Feb–90 Groundwater Sample SV 3,3'-Dichloroborzidine 20 U B-7 17 – Oct-91 Groundwater Sample SV 2-Methylnaphthalene 50 U B-7 17 – Oct-91 Groundwater Sample SV 3,3'-Dichloroborzidine 50 U B-7 17 – Oct-91 Groundwater Sample SV 3,3'-Dichloroborzidine D U					3,3'-Dichlorobonzidine	20	n	Ug/L	10
B-2 06-Feb-90 Groundwater Sample SV ALL SEMI-VOCATILES TO Teb-90 Groundwater Sample SV 3,3'-Dichloroborzidine ZO U E-5 07-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidine ZO U C-10 07-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidine ZO U C-10 07-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidine ZO U C-10 07-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidine ZO U A-4 16-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidine ZO U A-4 16-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidine ZO U A-4 17-Oct-91 Groundwater Sample SV 3,3'-Dichloroborzidine ZO U B-7 17-Oct-91 Groundwater Sample SV 3,3'-Dichloroborzidine ZO U B-7 17-Oct-91 Groundwater Sample <td< td=""><td></td><td></td><td></td><td></td><td>ALL CEMI VOLVETILES</td><td>2</td><td>•</td><td>1/60</td><td>Y.</td></td<>					ALL CEMI VOLVETILES	2	•	1/60	Y.
B-2 06 - Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidine 20 U E-5 07 - Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidine 20 U C-10 07 - Feb-90 Groundwater Sample SV 2-Methyknaphthalone 5 U C-10 07 - Feb-90 Groundwater Sample SV 2-Methyknaphthalone 50 U A-4 16 - Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidine 20 U A-4 16 - Feb-90 Groundwater Sample SV 2-Methyknaphthalone 20 U A-4 17-Oct-91 Groundwater Sample SV 2-Methyknaphthalone 50 U B-7 17-Oct-91 Groundwater Sample SV 2-Methyknaphthalone 50 U B-7 17-Oct-91 Groundwater Sample SV ALL SEMI-VOLATILES ND NA					ALL SCHILL VOCATILLES			7	
E-5 07-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidine 20 U D-7 07-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidine 20 U C-10 07-Feb-90 Groundwater Sample SV 2-Methylnaphthalene 20 U C-10 07-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidine 20 U A-4 16-Feb-90 Groundwater Sample SV 2-Methylnaphthalene 20 U O7-Feb-90 Groundwater Sample SV 2-Methylnaphthalene 5 J B-7 17-Oct-91 Groundwater Sample SV 2-Methylnaphthalene 5 J B-7 17-Oct-91 Groundwater Sample SV ALL SEMI-VOLATILES ND NA					3,3'-Dichlorobarzidino	20	ס	ug/L	10
D-7 O7-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidine 20 U C-10 07-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidine 20 U C-10 07-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidine 20 U A-4 16-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidine 20 U 07-Feb-90 Groundwater Sample SV 2-Methylnaphthalene 5 J 07-Feb-90 Groundwater Sample SV 2,3'-Dichloroborzidine 50 U 17-Oct-91 Groundwater Sample SV 3,3'-Dichloroborzidine 50 U 17-Oct-91 Groundwater Sample SV 3,3'-Dichloroborzidine NA NA					3.3'-Dichlorobonzidina	50	>	UQ/L	01
D-7 07-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzldine 20 U C-10 07-Feb-90 Groundwater Sample SV 2-Molthyknaphthalene 5 J C-10 07-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzldine 20 U A-4 16-Feb-90 Groundwater Sample SV 2-Methyknaphthalene 20 U B-7 17-Oct-91 Groundwater Sample SV 2-Methyknaphthalene 5 J B-7 17-Oct-91 Groundwater Sample SV ALL SEMI-VOLATILES ND NA									
C10 07-Feb-90 Groundwater Sample SV 2-Methylnaphthalene 5 J C10 07-Feb-90 Groundwater Sample SV 3,3'-Dichloroberzidine 20 U A-4 16-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidine 20 U 07-Feb-90 Groundwater Sample SV 2-Methylnaphthalene 5 J B-7 17-Oct-91 Groundwater Sample SV 3,3'-Dichloroborzidine 50 U R-7 17-Oct-91 Groundwater Sample SV ALL SEMI-VOLATILES ND NA					3,3'-Dichlorobonzidina	20	-	ng/L	10
C=10 O7-Feb-90 Groundwater Sampto SV 3,3'-Dichloroberzidino 20 U A=4 16-Feb-90 Groundwater Sampto SV 3,3'-Dichloroberzidino 20 U B=7 17-Oct-91 Groundwater Sampto SV 2-Methylnaphthaleno 5 J B=7 17-Oct-91 Groundwater Sampto SV ALL SEMI-VOLATILES NA					2-Mothylnaphthalono	ß	ר	ng/L	9
A-4 16-Feb-90 Groundwater Sample SV 3,3'-Dichloroborzidine 20 U 07-Feb-90 Groundwater Sample SV 2-Methylnaphthaleno 5 J 8-7 17-Oct-91 Groundwater Sample SV 3,3'-Dichloroborzidine 50 U 8-7 17-Oct-91 Groundwater Sample SV ALL SEMI-VOLATILES ND NA					3,3'-Dichloroberzidino	8		ng/L	10
O7-Feb-90 Groundwater Sample SV 2-Methylnaphthalene 5 J U C7-Feb-90 Groundwater Sample SV 3,3'-Dichlorobarzidne 20 U NA I7-Oct-91 Groundwater Sample SV ALL SEMI-VOLATILES ND NA					3,3'-Dichlorobonzlding	20	D	ug/L	10
17-Oct-91 Groundwater Sample SV 3,3'-Dichloroborglane 20 U	6-	07 - Fob-90			2-Methylnaphthaleno	NO			10
B-7 (7-Oct-91 Groundwater Sample SV ALL SEMI-VOLATILES ND NA NA NA NA SEMI-VOLATILES ND NA	6-	07-Fob-90			3,3'-Dichloroborzidino	20			10
n=2 17 = Oci=91 Groundwater Sample SV ALL SEMI=VOLATILES ND NA					ALL SEMI-VOLATILES	ON		ng/L	NA
	1 N-WAA				ALL SEMI-VOLATILES	ON		Ug/L	NA

							=	-
	17-0ct-91	Groundwater Sample	SS	ALL SEMI-VOLATILES	Q.	AN	ng/L	¥Z
t	17-0ct-91 17-0ct-91	Groundwater Sample Groundwater Sample	SV	Naphihalono 2-Mothyinophalono	9 8		1/6n	
- {	17-0ct-91	Groundwater Sample	SV	ALL SEMI-VOLATILES	ON	¥Z	J/g/L	Y Y
	17-Oct-91	Groundwater Sample	SV	ALL SEMI-VOLATILES	Q	X X	UB/L	Y Z
	17-0cl-91	Groundwater Sample	SV	ALL SEMI-VOLATILES	N	NA NA	UB/L	Z
	17-Oct-91	Groundwater Sample	SV	ALL SEMI-VOLATILES	Q	NA	vg/L	Y Y
	17-Oct-91	Groundwater Sample	>	1,1,1 -Trichloroethane	2	C	ng/L	
	17-Oct-91	Groundwater Sample	>	Trichloroethene	10		Ug/L	
	17-0c1-91	Groundwater Sample	^	Trichloroethene	B	ſ	Ug/L	
	17-0ct-91	Groundwater Sample	>	ALL VOLATILES	ND	VV	∆g/L	Y Z
	17-0ct-91 17-0ct-91	Groundwater Sample Groundwater Sample	>>	1,2-Dichloroethene(total) Trichloroethene	190		ug/L ug/L	
	17-0cl-91	Groundwater Samplo	>	Vinyi chlorida	17		ng/L	
	17-0ct-91	Groundwater Sample	>	Bonzono	. 19	•	ng/L	
	17-0cl-91	Groundwater Sample	> :	Toluene	~ 5	っ 	ug/L	-
	17-0ct-91 17-0ct-91	Groundwater Sample Groundwater Sample	> >	Elliylbanzana Xylana (loʻzi)	36		ng/L	
-	17-0ct-91	Groundwater Sample	>	1,1,1 - Trichloroethana	6	J.	ng/L	
	17-Oct-91	Groundwater Sample	>	ALL VOLATILES	QN	AN	∆/G∩	Y Y
	17-0c1-91	Groundwater Sample	>	ALL VOLATILES	QN	NA	Ug/L	NA
	17-0ct-91	Groundwater Sample	>	ALL VOLATILES	QN	NA	UB/L	Y Y
	17-Oct-91	Groundwater Sample		ALL VOLATILES	QN	NA	Ug/L	Y Z
	17-Oct-91	Groundwater Sample	>	ALL VOLATILES	QN	NA	vg/L	NA NA

Z			_					_		_		_				_								_		
DETECTION							•																			
UNITS	ug/L	ug/L	UB/L	√gu	ng/L	ng/L	ug/L	ug/L	ug/L	ug/L	ng/L	ug/L	ng/L	ug/L	ng/L	Ug/L	ug/L	ng/L	ng/L	ng/L	nβ/L	ug/L	UB/L	ug/L	ng/L	ug/L
QUALIFIER UNITS	7	7			7					Œ	5	Œ		5		2	>	ח	7	7	ס	כ	3	ס	ס	٦
RESULTS	20.2	22.5	5.1	80	101	347	528	0.17	196	4	n	4	1090	6	5.1	-	2	8	3.1	7.7	0.1	6	2.2	n	2.2	4.7
ANALYSIS	Antimony	Arsonic	Boryllium	Cadmlum	Chromium	Copper	Load	Morcury	Nickol	Salonlum	Sivor	Thallium	Zinc	Antimony	Arsonia	Boryllium	Cadmlum	Chromlum	Copper	Load	Morcury	Nickol	Solonium	SNot	Thallium	Zinc
слтевову	×	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	×	Σ	×	Σ	Σ	Σ	₹
DESCRIPTION	Groundwater Sample	Groundwater Sample	Groundwater Sample	Groundwater Sample	Groundwaler Sample	Groundwater Sample	Groundwaler Sample	Groundwater Sample	Groundwalor Samplo	Groundwater Sample																
DATE	17-0c1-91	17-0cl-91	17-0cl-91	17-Oc1-91	17-0cl-91	17-0cl-91	17-0cl-91	17-0cl-91	17-Oct-91	17-0c1-91	17-0cl-91	17-0cl-91	17-0cl-91	17-0c1-91	17-Oct-91	17-0cl-91	17-0cl-91	17-0ct-91	17-0cl-91	17-0ct-91	17-0ct-91	17-0cl-91	17-0cl-91	17-0ct-91	17-0cl-91	17-Oct-91
YES/NO	02	<u>Q</u>	2	<u>Q</u>	0 2	02	ON N	0	0	02	0 2	02	<u>Q</u>	YES												
GRID / FILTERED	B-7	8-7	8-7	8-7	B-7	8-7	19-7	B-7	B-7	B-7	8-7	8-7	1-8	8-7	8-7	1-9	6-7	B-7	8-7	8-7	8-7	8-7	0-7	8-7	8-7	8-7
SAMPLE	MW-2	WW-2	MW-2	MW-2	WW-2	MW-2	MW-2	WW-2	MW-2	WW-2	WW-2															

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LIMIT																										
UNITS L	ng/L	ng/L	ng/L	UB/L	ng/L	ng/L	Ug/L	UB/L	UB/L	ug/L	ug/L	ug/L	Ug/L	אשע "	ug/L	ug/L	ug/L	ug/L	שא/ר	ug/L	սց/Լ	ا/ ₀ 0	ug/L	עט/ר יי	νgν	7/6n
aUALIFIER	7	<u> </u>			7				1	ac ;	D 1	Ï	•	>	-	> :	5	> :	> .	-	>	>	3:	o :	5.	5
Ts au	29.5	6.9	3.7	4.6	2.96	208	109	0.29	202	8	n ·	4	645	6	2	-	7	n	n	6:	0.1	o.	2.2	r)	2.2	3
RESULTS					_				_	_		_	0	_	0	_	=		_	- -	<u>~</u>	_	F	_	F	0
SIIS	Antimony	Arsonic	Boryllium	Cadmium	Chromlum	Coppor	Load	Morcury	Nickel	Solonium	SNor	Thallium	Zinc	Antimony	Arsonic	Boryllium	Cadmium	Chromium	Copper	Load	Morcury	Nickol	Selenium	SNo	Thallium	Zinc
ANALYSIS FOR																										
Æ	≥	Z	Σ	Σ	Σ	×	Σ	Σ	Σ	Σ	Σ	Σ.	≥	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Z	Σ	Σ	≥	Σ	Σ
CATEGORY																										
	r Sample	r Sample	r Samplo	r Samplo	r Samplo	r Samplo	r Sample	r Samplo	r Samplo	r Sample	r Samplo	r Sample	r Sample	r Sample	or Sample	or Samplo	or Sample	or Samplo	or Sampla	or Samplo	er Sample					
DESCRIPTION	Groundwater Sample	Groundwater Sample	Groundwater Sample	Groundwater Sample	Groundwater Sample	Groundwater Sample	Groundwator Samplo	Groundwater Sample																		
DATE	7-0cl-91	17-0ct-91	17-Oct-91	17-Oct-91	17-Oct-91	17-0cl-91	17-Oct-91	17-0cl-91	17-0cl-91	17-0c1-91	17-0ct-91	17-0cl-91	17-0ct-91	17-0cl-91	17-0cl-91	17-Oct-91	17-Oct-91	17-0ct-91	17-0ct-91	17-Oct-91	17-Oct-91	17-Oct-91	17-Oct-91	17-0ct-91	17-Oct-91	17-Oct-91
	0 14	NO 17		_		_	NO 17	_	_	_	_	_	_		_		_		_	YES 17	_		YES 17	_	YES 17	YES 17
YES/NO FILTERED	Z	Z	Z	z	Z	Z	Z	Z	Z	2	Z	_		χ.	×	Ϋ́	χ.	Σ.	<u></u>	>	>	>	>			
	B-2	8-2	B-2	B-2	B-2	8-2	8-2	B-2	B-2	B-2	8-2	B-2	B-2	B-2	8-2	8-2	B-2	B-2	8-2	8-2	B-2	B-2	B-2	B-2	8-2	B-2
SAMPLE CRID	WW-4	WW-4	MW-4	WW-4	WW-4	WW-	WW-4	MW-4	WW-4	MW-4	WW-4	MW-4	MW-4	WW-4	MW-4	MW-4	WW-4	MW-4	MW-4	WW-4	MW-4	WW-4	MW-4	MW-4	MW-4	MW-4

YES/NO FILTERED	DATE	DESCRIPTION CA	CATEGORY	ANALYSIS FOR	RESULTS	RESULTS QUALIFIER UNITS		DETECTION
-	17-0cl-91	Groundwater Sampla	Σ	Anlimony	24.8	7	ug/L	
F	17-0cl-91	Groundwater Sample	Σ	Arsonic	2	7	J/Bn	
Ç	17-0cl-91	Groundwater Sample	Σ	Boryllium	12.1		ng/L	
#	17-0cl-91	Groundwater Sample	Σ	Cadmium	10.9		J/Bn	
=	17-Oct-91	Groundwater Sample	Σ	Chromlum	247		J/Bn	
=	17-0ct-91	Groundwater Sample	Σ	Copper	868		UB/L	-
=	17-Oct-91	Groundwater Sample	Σ	Load	647		ng/L	
17	17-0cl-91	Groundwater Sample	Σ	Morcury	0.55		UB/L	
11	17-0ct-91	Groundwater Sample	Σ	Nickol	730		nB/L	
17	17-0c1-91	Groundwater Sample	Σ	Solonium	89	Œ	ng/L	
17	17-0cl-91	Groundwater Sample	Σ	SNor	e	>	UB/L	
=	17-0cl-91	Groundwater Sample	Σ	Thallium	4	cc	ug/L	
=	17-0c1-91	Groundwater Sample	Σ	Zinc	3260		ng/L	
7	17-0c1-91	Groundwater Sample	Σ	Antimony	O)	э	ng/L	
7	17-0cl-91	Groundwater Sample	Σ	Arsonic	2	5	ng/L	
=	17-0cl-91	Groundwater Sample	Σ	Boryllium	-	>	ng/L	
-	17-0cl-91	Groundwater Sample	Σ	Cadmlum	2	Þ	ng/L	
-	17-0cl-91	Groundwater Sample	Σ	Chromlum	n	D	ug/L	
_	17-0cl-91	Groundwater Sample	Σ	Coppor	6.2	7	ng/L	
_	17-0c1-91	Groundwaler Sample	Σ	Load	2	7	ug/L	
_	17-0c1-91	Groundwater Sample	Σ	Morcury	0.1	ס	ug/L	
_	17-Oct-91	Groundwater Sample	Σ	· Nickol	6	5	ng/L	
_	17-Oct-91	Groundwater Sample	Σ	Solonlum	2.2	3	ng/L	
	17-0cl-91	Groundwater Sample	Σ	SNor	C	>	ug/L	
_	17-0cl-91	Groundwater Sample	Σ	Thallium	2.2	כ	Ug/L	
_	17-0cl-91	Groundwater Sample	M	Zinc	20.1		J/gv	

]

DETECTION																										
UNITS	na/F	ug/L	ממ/ר	na/L	ug/L	ug/L	ug/L	ng/L	ממ/ך	ug/L	ng/L	ug/L	ng/L	ng/L	ug/L	ng/L	ug/L	ug/L	ng/L	UB/L	UB/L	Ug/L	ng/L	ug/L	1/6n	, no./
JALIFIER	7	7						ס		Œ	כ	Œ		2		כ	ס	ר	ס	7	כ	2	3	כ	כ	7
RESULTS QUALIFIER	18.9	4.5	n	2.8	66.4	155	147	0.1	152	8	c	4	513	6	7.6	-	2	0	n	3.0	0.1	6	2.2	n	2.2	4.7
ANALYSIS FOR RI	Antimomy	Arsonic	Boryllium	Cadmlum	Chromium	Copper	Lead	Marcury	Nickel	Selenium	Siver	Thallium	Zluc	Antimony	Arsonic	Beryllium	Cadmlum	Chromlum	Copper	Load	Marcury	Nickel	Solenium	SNor	Thallium	Zine
CATEGORY	×	×	Σ	×	×	Σ	×	Σ	×	. S	×	Σ	Z	Σ	×	×	Σ	×	. ≅	<u>×</u>	Σ	Z	Σ	Σ	×	×
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	D-7	7-0 ·	D-7	D-7	D-7	D-7	0-7	D-7	2-0	0-7	0-7	D-7	2-0	D-7	D-7	D-7	D-7	D-7	0-7	2-0	D-7	D-7	D-7	0-7	0-7	0-7
SAMPLE . GRID	MW-7	MW-7	MW-7	WW-7	MW-7	MW-7	MW-7	MW-7	WW-7	MW-7	WW-7	WW-7	WW-7	WW-7	WW-7	MW-7	MW-7	MW-7	MW-7	WW-7	WW-7	MW-7	WW-7	WW-7	WW-7	MW-7

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DETECTION																											
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	Antimony		Arsonic	Boryllum	Cadmium	Chromium	Coppor	Load	Morcury	Nickol	Solonlum	SNor	Thallium	Zino	Antimony	Arsonic	Boryllium	Cadmium	Chromium	Copper	Load	Morcury	Nickol	Solonium	SNor	Thallium	Zinc
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CATEGORY	2		Σ	Z	X	Σ	Σ	Σ	Σ	≊	Σ	Σ	Σ	Σ	×	Z	×	Σ	Σ	≆	×	Σ	×	Σ	2	Z	2
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YES/NO FILTERED	2	2	2	2	2	S	2	Z	ž	ž	2	ž	02	02	YES	YE	YES	Ϋ́Ε	ΥE	Ϋ́E	Ϋ́	YES	YES	YES	YES	YES	YES
GRID F		9-V	9-V	A-8	9-V	A-8	A-8	A-B	9-V	A-6	A-8	A-8	A-8	9-Y	A-8												
SAMPLE # C	-	OI - WW	MW-10	MW-10	MW-10	MW - 10	MW-10	MW - 10	MW-10	MW-10	MW-10	WW-10	MW-10	MW - 10	MW-10	MW - 10	MW-10	MW-10	MW-10	MW-10	MW - 10	MW-10	MW-10	MW-10	MW-10	MW-10	MW-10
SAM																								_			

RICKENBACKERANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - GROUNDWATER

SAMPLE #	GRID (YES/NO FILTERED	OATE	DESCRIPTION	сатевоях	ANALYSIS FOR	RESULTS	QUALIFIER UNITS		DETECTION
MW-11	F-7	ON.	17-Ocl-91	Groundwater Sample	Σ	Antimony	65.4	7	ug/L	
MW-11	F-7	92	17-Oct-91	Groundwater Sample	Σ	Arsonic	45.9	7	ng/L	
MW-11	F-7	9	17-Oct-91	Groundwater Sample	Σ	Boryllium	6.2		nayr	
MW-11	F-7	ON N	17-Oct-91	Groundwater Sample	Σ	Cadmlum	8.7		ug/L	-
MW-11	F-7	ON	17-0ct-91	Groundwater Sample	Σ	Chromium	149	7	ng/L	
MW-11	F-7	2	17-0cl-91	Groundwater Sample	Σ	Copper	322		Ug/L	•
MW-11	F-7	ON ON	17-Oct-91	Groundwater Sample	\$	Load	225		ng/L	٠.
MW-11	F-7	9	17-0cl-91	Groundwater Sample	Σ	Morcury	0.14		ng/L	
MW-11	F-7	ON.	17-Oct-91	Groundwater Sample	Σ	Nickol	270		ug/L	
MW-11	F-7	9	17-Oct-91	Groundwater Sample	Σ	Muluolos .	8	.	ug/L	
MW-11	F-7	8	17-Oct-91	Groundwaler Sample	Σ	Sivor	6		ng/L	
MW-11	F-7	02	17-Oct-91	Groundwater Sample	Σ	Thallium		α.	ug/L	
MW-11	F-7	ON.	17-0c1-91	Groundwater Sample	Z	Zinc	1090		ug/L	
MW-11	F-7	YES	17-0ct-91	Groundwater Sample	Σ	Antimony	6	>	ug/L	
MW-11	F-7	YES	17-0cl-91	Groundwater Sample	Σ	Arsonic	4.1	. د	UB/L	
MW-11		YES	17-Oct-91	Groundwater Sample	Σ	Boryllium	-	>	ng/L	
MW-11	F-7	YES	17-0c1-91	Groundwater Sample	Σ	Cadmlum	7	>	J/Bn	
MW-11		YES	_	Groundwater Sample	Σ	Chromlum	n	>	ng/L	
MW-11	F-7	YES		Groundwater Sample	Σ	Copper	n	>	ng/L	
MW-1		YES	17-Oct-91	Groundwater Sample	Σ	Load	2.2	٦	ng/L	
MW-11		YES	17-0cl-91	Groundwater Sample	Σ	Morcury	o	>	ug/L	
MW-11		YES	17-Oct-91	Groundwater Sample	Σ	Nickol		> ;	מאלים	
MW-11			17-0c1-91	Groundwater Samplo	Σ	Solonium	2.2	3	ng/L	
MW-11	F-7		17-Oct-91	Groundwater Sample	×	SNor		> :	ng/L	
MW-11			17-Oc1-91	Groundwater Sample	Σ	Thallium	2.2	-	/bn	
MW-11	1 F-7		17-0cl-91	Groundwater Sample	M	Zinc	3) 	Vgu	

RICKENBACKERANGB HAZARDOUS WASTE STORAGE AREA ANALYTICAL RESULTS - GROUNDWATER

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DETECTION																										
UNITS	ug/L	ug/L	ng/L	ng/L	ng/L	תם/ך	ng/L	UB/L	ug/L	ng/L	ug/L	ng/L	ug/L	ng/L	1/0n	\08/L	ug/L	ug/L	ח/מח	UB/L	no/L	UB/L	ug/L	ng/L	ng/L	77
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RESULTS AU	33.9	12.9	4.9	6.5	97.1	328	185	0.18	178	ю	က	4	898	o	3.6	_	2	n	3.9	2.4	0.1	o,	2.2	n	2.2	6.2
RES	Antimony	Arsonic	Boryllium	Cadmium	Chromlum	Coppor	Lead	Mercury	Nickal	Solonlum	SNor	Thallium	Zinc	Antimony	Arsonic	Boryllium	Cadmium	Chromlum	Copper	Load	Morcury	Nickol	Selonium	Sivor	Thallium	Zinc
ANALYSIS FOR	¥.		6 0	O	さ					u,				∢		-	O	Ö					•			
סוזיל	Σ	Σ	Σ	Σ	Z	Σ	Z	×	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Σ	Z	Σ
CATEGORY	old	old	old	old	plo	plo	plo	old	ple	plo	ple	plo	old	oldi	. lold	oldi	ple	oldi	plo	oldı	oldu	oldr	oldu	oldu	oldu	nplo
DESCRIPTION	Groundwater Sample	Groundwater Sample	Broundwater Sample	Groundwater Sample	Groundwater Sample	Groundwater Sample	Groundwater Sample	Broundwater Sample	Groundwater Sample	Groundwater Sample	Groundwater Sample	Groundwater Sample	Groundwater Sample	Groundwater Sample												
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DATE	17-0ct-91	17-Oct-91	17-Oct-91	17-Oct-91	17-Oct-91	17-0cl-91	17-Oct-91	17-0cl-9	17-Oct-91	17-Oct-9	17-Oct-9	17-Oct-9	17-Oct-9	17-Oct-9	17-Oct-9	17-Oct-9	17-Oct-9	17-0c1-9	17-Oct-9	17-Oct-9	17-Oct-9	17-Oct-9	17-Oct-9	17-0cl-9	17-0cl-9	17-Oct-9
YES/NO FILTERED	9	9	ON.	9	ON.	Q.	2	ON	ON	OZ	02	O _Z	Q	YES	YES	YES	YES	YES	YES	YES	YES	YES	YES	YES	YES	YES
A GRID	E-9	6-W	E-9	E-9	E-9	E-9	E-9	E-9	E-9	6-3	E-9	6-W	B-19	6-3	В - В	6 1 1	E-9	6. L	E-9	E-9	6-11	п 6-п	E-9	E-9	E-9	E-9
SAMPLE	MW-12	MW-12	MW-12	MW-12	MW-12	MW-12	MW-12	MW-12	MW-12	MW-12	WW-12	MW-12	MW-12	MW-12	MW-12	MW-12	MW-12	MW-12	MW-12							
	1									_		_				_	_					_		_	_	

TABLE A.1

SOLUBILITY CONSTANTS HAZARDOUS WASTE STORAGE AREA RICKENBACKER AIR NATIONAL GUARD BASE

	Water
Compound	Solubility (1)
Acetone	1000000 mg/L (2)
Benzene	1750 mg/L
Carbon Tetrachloride	757 mg/L
Chloroethane	N/A (3)
Chloroform	82000 mg/L
Creosol	31000 mg/L
1,1-Dichloroethane	280 mg/L
1,2-Dichloroethane	5500 mg/L
cis-1,2-Dichloroethene	3500 mg/L
trans-1,2-Dichloroethene	6300 mg/L
Ethylbenzene	152 mg/L
Dichloromethane	4600 mg/L
2-Butanone	268000 mg/L
Phenol	93000 mg/L
1,1,1,2—Tetrachloroethane	2900 mg/L
1,1,2,2-Tetrachloroethane	2900 mg/L
7 etrachloroethene	150 mg/L
Methylchloroform	N/A
1,1,2-Trichloroethane	4500 mg/L
Trichloroethene	1100 mg/L
Vinyl Chloride	2670 mg/L
Xylenes	198 mg/L

- (1) Source Superfund Public Health Evaluation Manual, EPA/540/1-86/060, October 1986.
- (2) mg/L- milligrams per liter
- (3) Not available.

REVISED MARCH 1993

APPENDIX B-2

SUMMARY OF PREVIOUS INVESTIGATIONS, INCLUDING 1991 ANALYTICAL DATA

APPENDIX B

PREVIOUS INVESTIGATIONS

1.0 1988 FIELD INVESTIGATION

Engineering-Science (ES) completed the first phase of the field investigation at the Hazardous Waste Storage Area (HWSA) in October 1988. The purpose of the investigation was to determine if the soil or groundwater beneath the site had been contaminated due to spills or leaks form on-site storage containers.

The results of this investigation are presented in detail in the <u>Field Investigation</u>
Report - Hazardous Waste Storage Area: Rickenbacker Air National Guard Base,
Columbus, Ohio (1989), and the results are summarized in the following paragraphs.

1.1 SUMMARY OF ANALYTICAL RESULTS

The soil-gas survey identified with elevated concentrations of benzene, toluene and ortho-xylene (BTX). Concentration of total BTX in the soil gas ranged from undetectable to 29.8 ppm.

Analyses of the soil samples indicated elevated semi-volatile organic and metals concentrations. The characteristics of the semi-volatile organics found were typical of coal-tar derivatives and phthalates. Metals identified included cadmium, chromium, copper, lead and zinc.

Three of the auger borings made during soil sampling were completed as monitoring wells in the shallow aquifer. Water samples from two of these wells exhibited volatile organic concentrations in excess of Federal Maximum Contaminant Levels (MCLs). Water from MW1 contained 94 μ g/l benzene, 20 μ g/l xylenes and 13 μ g/l methylnapthalene. Water from MW3 contained 44 μ g/l trichloroethene. Samples from all wells had total unfiltered metals concentrations in excess of Federal Drinking Water Standards for arsenic, cadmium, chromium and lead.

2.0 1990 FIELD INVESTIGATION

The phase of the field investigation at the HWSA was completed in March 1990. The purpose of this additional field investigation was to determine the extent of contamination and to allow revision of the Closure Plan to affect a "clean" closure of the site.

The pre-closure sampling activities included soil sampling at the surface and at depth and the installation of six new monitoring wells in and around the HWSA (MW4 through MW9). The results of this investigation are presented in detail in the <u>Pre-Closure Sampling Report - Hazardous Waste Storage Area: Rickenbacker Air National Guard Base. Columbus, Ohio</u> (1992), and a summary of the results follows.

2.1 SUMMARY OF ANALYTICAL RESULTS

2.6.1 Metals

Total metals were found over the site with higher levels within the fenced area.

Detected above background criteria were beryllium, cadmium, copper, lead, mercury, silver and zinc.

2.2 VOLATILE ORGANIC COMPOUNDS

Volatile organic compounds in samples from the 0-2 foot interval, were only analyzed for at six hand boring and two monitoring well locations. The only VOCs detected were $440,000 \mu g/kg$ o-xylene at HB1.

Volatile organic compounds were detected at concentrations up to 1,900,000 μ g/kg in soils from the 3-5 foot interval. Elevated ethylbenzene and o-xylene concentrations were found in HB1, near Building 560, while benzene was detected in AB2.

Volatile organics were found in samples from the 8-10 foot interval at levels up to $27,000 \mu g/kg$ of o-xylene. The highest concentrations were found at AB1, AB14 and MW7. Specific compounds include: benzene, ethylbenzene, xylenes and 1,1,1-trichloroethane.

Samples from the 13-15 foot interval containing volatile organic compounds were found in the southern corner and along the northeast side of the area. These include:

benzene, ethylbenzene, toluene, xylenes, acetone, trichloroethene, trans-1,2-dichloroethene, 1,1-dichloroethene, and vinyl chloride. The highest concentration was $1,000 \mu g/kg$ trans-1,2-dichloroethene at MW6.

At the greater than fifteen foot interval, sand and gravel is present to a depth of approximately 25' with a thin layer of clay from 18'-19'. Detected volatile and semi-volatile organics were confined to the southeast side of the area. Semi-volatile organics were found only at MW1 at a total concentration of 1,830 μ g/kg. The highest volatile organic concentrations were also found at this location. They were benzene, ethylbenzene, and o-xylene at concentrations of 1,900, 11,000, and 20,000 μ g/kg, respectively.

2.3 SEMI-VOLATILE ORGANIC COMPOUNDS

Semi-volatile organics were detected in the soils at various depths and ranged from non-detect to $4,630 \mu g/kg$.

2.6.2 Groundwater

Volatile and Semi-Volatile Organics

On the analytical results map (Sheet 6), both the 1990 and 1988 sampling data are shown. The only semi-volatile organic compound found in the groundwater was 2-methylnaphthalene at 5J μ g/L in MW8.

Volatile organics compounds were detected in MW1, MW3, MW6 and MW7, and include benzene, ethylbenzene, o-xylene, p-xylene, trichloroethene, and trans-1,2-dichloroethene. In addition, four feet of phase-separated hydrocarbons were floating in MW5. Fingerprint analysis of the liquid hydrocarbons identified it as a 30 to 40 percent weathered gasoline mixed with jet fuel.

Filtered Metals

Four metals were detected at all concentrations below the Federal Drinking Water Standards. These four metals were arsenic (found at 2.0 to 9.4 μ g/L), lead (found at 3.1 to 14.0 μ g/L), zinc (found at 5.0 to 35 μ g/L) and mercury (found at 0.11 μ g/L).

3.0 1991 FIELD INVESTIGATION

The third phase of the field investigation at the HWSA was completed in October 1991. Field activities conducted during this investigation include groundwater screening, monitoring well installation and soil sampling from the well borings, surface soil sampling, and groundwater sampling.

The additional sampling was conducted to fill data gaps existing after the original pre-closure sampling report. Specifically, these data gaps are:

- The anomalously high concentrations of semi-volatile organic compounds (SVOCs) found at the surface soils of the westernmost corner of the HWSA.
- The extent of VOCs previously detected in the groundwater.

The results of this investigation were reported in the Addendum to the Pre-Closure Sampling Report - Hazardous Waste Storage Area: Rickenbacker Air National Guard Base, Columbus, Ohio (1992) and a summary of the results follows.

The data obtained through the groundwater sampling indicate that petroleum hydrocarbon and chlorinated organic contamination is restricted to the area upgradient and downgradient of the four underground storage tanks (USTs) numbered 47, 48, 49, 50. Wells MW1 and MW5, where phase-separated hydrocarbons were observed, lie in the northern and furthest upgradient portion of this contaminant plume. In the downgradient direction, the dissolved organic plume does not extend to MW11 and MW12.

Volatile organic results of the groundwater sampling events indicate that chlorinated organics are present in MW3 and MW6. The compound 1,1,1-trichloroethane was found at an estimated concentration of 3 μ g/L in MW8 during the 1991 sampling event. This compound was also found in MW2 at an estimated quantity of 2 μ g/L. Although this compound was found in the associated trip blank, it is still possible that it is present at this site since it has been detected in the past.

No groundwater samples were collected from MW5 in either sampling event due to the presence of PSH. In 1990, MW1 had concentrations of dissolved benzene, ethylbenzene and xylenes; however, due to the presence of PSH in 1991, this well was

not resampled. Dissolved benzene, ethylbenzene and xylenes were found in MW7. No volatile organics were detected in MW4 and MW9.

The semi-volatile organic compounds, 2-methylnaphthalene and naphthalene were found in MW7 at estimated concentrations 2 and 6 μ g/L respectively.

Groundwater analyses indicate the presence of total metals in the water samples; however, the filtered aliquot analysis showed a decrease in metals concentrations. Therefore, the presence of metals is associated with the silt suspended in the water sample.

Four metals were detected in filtered groundwater samples, all at concentrations below the Federal Drinking Water Standards. These four metals were arsenic found at 2.9 to 12.1 μ g/L, copper at 3.1 to 6.2 μ g/L, lead at 1.9 to 7.7 μ g/L, and zinc at 3.7 to 20.1 μ g/L.

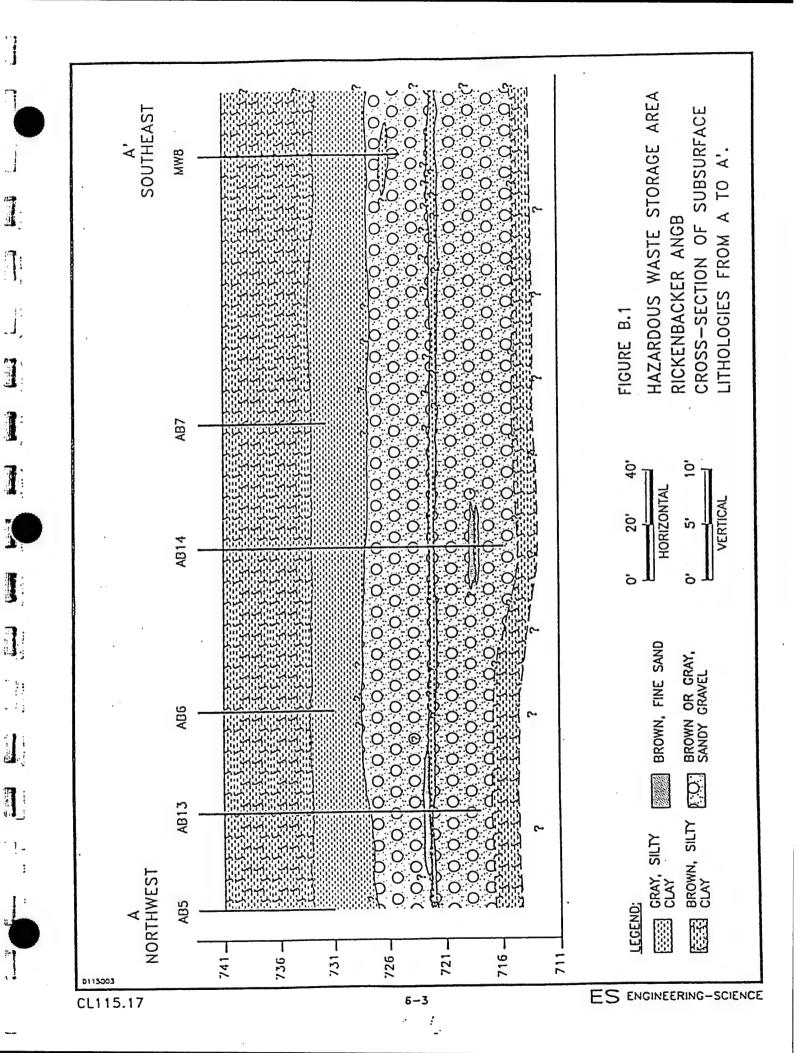
Soil Results

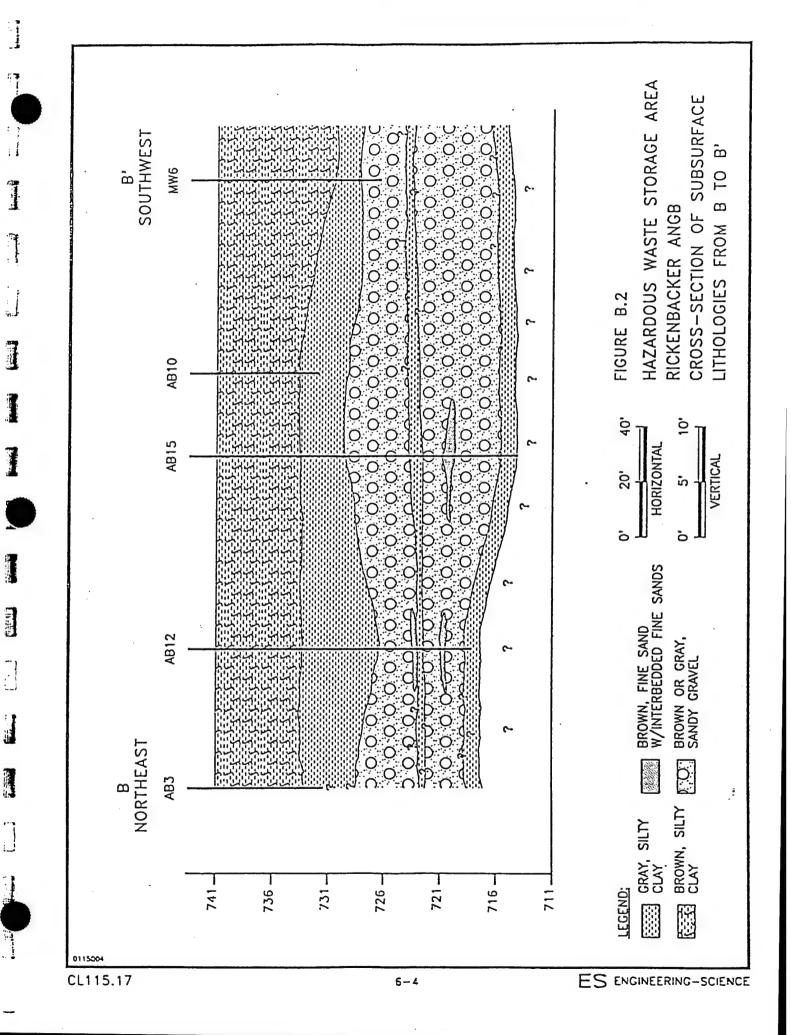
Five of the ten surface soil samples were collected at the fenceline surrounding the site, the remaining five from locations off site. One sample (SS3) was collected in duplicate. All surface soil samples were analyzed for SVOCs and the priority pollutant metals.

Samples with no detected SVOCs are SS1, SS2 and SS7. Samples SS3, SS9 and SS10 have the highest total SVOC concentrations, 1108, 2250 and 977 μ g/kg respectively. The SVOCs detected can be classified as coal tar derivatives.

Metals analysis from these surface soil samples were compared to background levels that were established for the Base during the Site Investigation for the Installation Restoration Program. Detected above background criteria were arsenic, cadmium, copper, lead, nickel, silver and zinc.

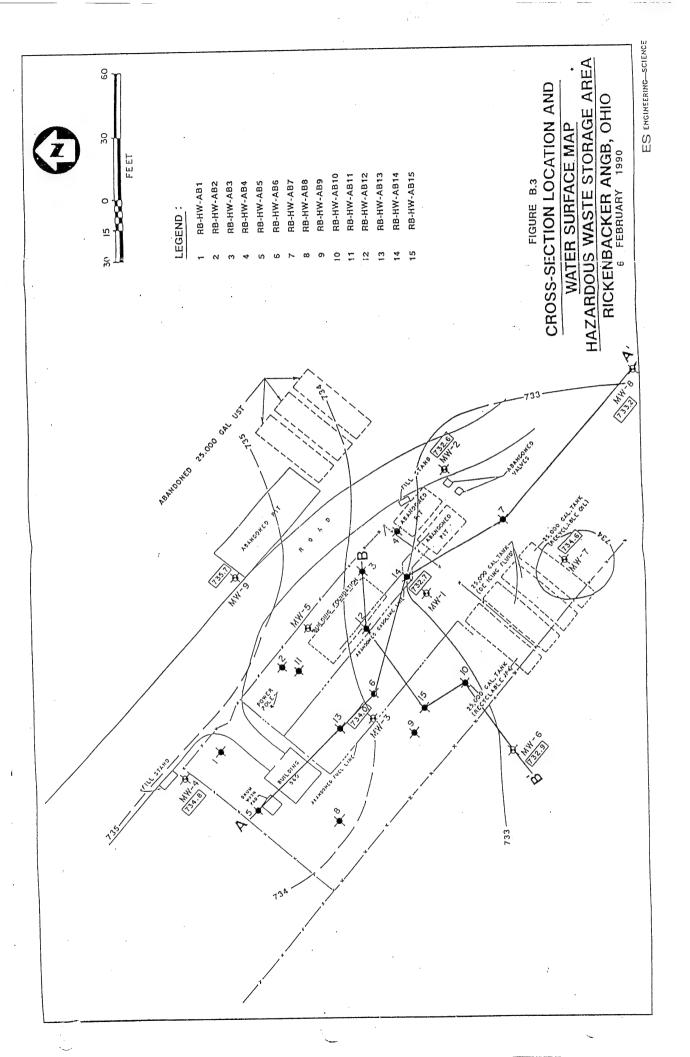
Soil boring samples were collected from two depths (3 to 5 feet and 13 to 15 feet) in each of the three soil borings. Each sample was analyzed for SVOCs, VOCs, and priority pollutant metals. SVOCs were not detected in either of the two samples from MW10 and MW11. The soil sample from the 13 to 15 foot horizon of MW12 had a total semi-volatile concentration of 1569 μ g/kg although no SVOCs were detected in the shallow soil sample (3 to 5 feet), or in the groundwater sample from this well. These compounds are coal tar derivatives.





No VOCs were detected in the soils from borings from MW10, MW11 and MW12.

Metals analysis from the soil samples obtained from the soil borings were also compared to the background levels for the Base. Detected above background criteria were arsenic, beryllium, cadmium, chromium, copper.



GROUNDWATER SURVEY RICKENBACKER ANGB, OHIO

October 1991

Prepared for:

Engineering Science Cleveland, Ohio

Project 533935

BURLINGTON ENVIRONMENTAL, MATHES DIVISION 4091 Venture Place Groveport, Ohio 43125

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GROUNDWATER SURVEY RICKENBACKER ANGB, OHIO

1. INTRODUCTION

Engineering Science (ES) requested that John Mathes & Associates, Inc., (Mathes) perform a groundwater survey at the site located at Rickenbacker ANGB, Ohio.

The purpose of this survey was to evaluate the potential presence, and estimate the extent of impact, of volatile organic compounds (VOCs) at the above mentioned facility. This survey was performed by obtaining and analyzing groundwater samples on site.

Mathes arrived at the facility on October 7, 1991. Groundwater sampling was performed from October 7 to October 9, 1991.

2 SUMMARY

Twenty one groundwater samples from 17 probe hole locations were collected and analyzed. Three duplicate groundwater samples, and ten sample blanks were also analyzed. Samples were analyzed for the following chemicals:

- o benzene;
- o ethylbenzene;
- o m&p-xylene;
- o o-xylene;
- o toluene;
- o trichloroethylene;

The analytical results are summarized in Table 1. Site maps with sample locations are being prepared by ES.

3 SAMPLING LOCATIONS

Groundwater samples were collected on site at locations suspected by ES to be impacted with VOCs. Sampling locations were selected by ES based on on-site soil boring data previously obtained by ES, the location of underground utilities, groundwater flow, and vehicle accessibility.

Sampling locations were mapped by ES. At the request of ES, a sampling location map is not included in this report.

TABLE 1

GROUNDWATER ANALYTICAL RESULTS

RICKENBACKER ANGB, OHIO

					Concentration	(ng/L)		
Sample I.D.	Probe Hole Number	Depth (Feet)	Benzene	Trichloroethylene	Toluene	Ethylbenzene	Total Xylenes	Comments
BLANK-01			ND(1)	ND(1)	ND(1)	ND(1)	ND(1)	QC-System Blank
BLANK-02	•		ND(1)	ND(1)	(1) ON	ND(1)	ND(1)	QC-Probe Rod Blank
GW-01	PH-01	16.0	æ	ND(1)	ND(1)	ND(1)	ND(1)	Groundwater
GW-02	PH-02	21.0	ND(1)	ND(1)	ND(1)	ND(1)	ND(1)	Groundwater
GW-02D	PH-02	21.0	ND(1)	ND(1)	ND(1)	ND(1)	ND(1)	QC-Duplicate
GW-03	PH-03	20.0	ND(1)	ND(1)	ND(1)	ND(1)	ND(1)	Groundwater
GW-04	PH-04	16.0	ND(1)	. 55	1385	393	. 259	Groundwater
BLANK-03	•	•	ND(1)	ND(1)	ND(1)	ND(1)	ND(1)	QC-System Blank
BLANK-04	•	•	ND(1)	ND(1)	ND(1)	ND(1)	ND(1)	QC-System Blank
BLANK-05	•		ND(1)	ND(1)	ND(1)	ND(1)	ND(1)	QC-Probe Rod Blank
GW-05	PH-05	20.0	ND(1)	ND(1)	ND(1)	ND(1)	ND(1)	Groundwater
90-MD	PH-06	20.0	ND(1)	ND(1)	ND(1)	ND(1)	ND(1)	Groundwater
GW-07	PH-07	20.0	ND(1)	ND(1)	ND(1)	ND(1)	ND(1)	Groundwater
GW-08	PH-08	24.0	ND(1)	ND(1)	ND(1)	ND(1)	ND(1)	Groundwater
GW-09	PH-09	24.0	ND(1)	. ND(1)	ND(1)	ND(1)	ND(1)	Groundwater

Not detected at the detection limit in parentheses 8 8 gi

Quality Control Micrograms per liter Not Applicable

TABLE 1

GROUNDWATER ANALYTICAL RESULTS

RICKENBACKER ANGB, OHIO

					Concentration	(ug/L)		
Sample	Probe Hole	Depth				:) }	2000
	Number	(Feet)	Benzene	Trichloroethylene	Toluene	Ethylbenzene	iotal Aylenes	COULINATION
						177.00	147014	Groundwater
GW-10	PH-10	24.0	ND(1)	(1) ND(1)	ND(1)	(L)QN	(i)ON	GIODIOWALEI
O.W. 11	DH-11	24.0	(1)ON .	ND(1)	ND(1)	ND(1)	ND(1)	Groundwater
GW-11	DH.13	24.0	NOCI	ND(1)	ND(1)	ND(1)	ND(1)	Groundwater
5W-12	21.00	0.1.0	168	ND(1)	ND(1)	ND(1)	ND(1)	Groundwater
GW-13	2.	7.7	NOCE	ND(1)	ND(1)	ND(1)	ND(1)	QC-System Blank
BLANK-06	•		(1)GN	(E) CN	NDCO	ND(1)	ND(1)	Groundwater
GW-14	PH-14	18.0	(1)ON	(1)011	ND(4)	C CN	(DCN	QC-Duplicate
GW-14D	PH-14	18.0	ND(1)	(1) (1)		(1)011	(1)QN	OC System Blank
RI ANK-07	•	•	ND(1)	ND(1)	ND(1)	ND(1)		CC-System Diam
OLANIK OB	•	•	ND(1)	ND(1)	ND(1)	ND(1)	ND(1)	QC-System Blank
BLANK-00			CON	ND(1)	ND(1)	ND(1)	ND(1)	QC-Probe Rod Blank
BLANK-09	•		ND(4)	C)CN	ND(1)	ND(1)	ND(1)	Groundwaler
GW-15	PH-06	72.0	(1)00		(F)CIN	ND(1)	ND(1)	Groundwater
GW-16	PH-06	20.0	ND(1)	(L)QN	(1)(1)	(1)(1)	(s) IN	Groundwater
GW-17	PH-06	18.0	ND(1)	ND(1)	ND(1)	(L)QN	(1)(1)	Glodionator
GW-18	90.Hd	16.0	ND(1)	ND(1)	ND(1)	ND(1)	ND(1)	Groundwater
GW-10	04.46	0 00	ND(1)	ND(1)	ND(1)	ND(1)	ND(1)	Groundwater
61-M5		2 6		ND(1)	ND(1)	ND(1)	ND(1)	Groundwater
GW-20	PH-16	20.0	r		ND(4)	ND(1)	ND(1)	QC-Duplicate
GW-20D	PH-16	20.0	က	(L)ON	(1)011	(1)	NO(1)	Groundwater
GW.21	PH-17	20.0	172	ND(1)	ND(1)	٥	(1)01	Groundhaidh
01 4 10		•	ND(1)	ND(1)	ND(1)	ND(1)	ND(1)	QC-System Blank
BLAINN-IN								

4 GROUNDWATER SAMPLING AND ANALYSIS

Twenty one groundwater samples and three duplicate groundwater samples were collected from 17 locations using the RECONSM System equipment. Samples were collected from a depth of 16 to 24 feet. Analytical results for the groundwater samples are presented in Table 1.

A hydraulic probe unit was used to drive and withdraw the groundwater sampling probes. A hydraulic hammer was used where necessary to assist in driving probes through unusually hard soil. The probes consisted of three-foot lengths of 0.75-inch-diameter threaded steel pipes with detachable drive points.

After the probe was inserted into the groundwater, the probe was withdrawn approximately one foot to create an annular space from which to collect a representative sample. A check valve attached to a length of polyethylene tubing was inserted through the probe into the water table. The sample was collected in the tubing, the tubing was pulled up, and the water in the tubing drained into a 40-milliliter (mL) glass vial sealed with a Teflon-lined septum screw cap. The sample was given to the GC technician for on-site analysis.

A Hewlett-Packard Model 5890A Series II gas chromatograph (GC) was used for the analysis of groundwater samples. Compound separation and detection were performed using a 30-meter wide-bore DB-624 volatile organic column and a flame-ionization detector (FID). Appendix A contains the GC field work sheets.

The groundwater samples were analyzed by modified U.S. Environmental Protection Agency (USEPA) Method 601/602. Modifications include use of a FID, a static headspace analysis, single-point calibration, and limited quality control measures.

Each sample vial was shaken for two minutes and heated at 70° C for 10 minutes to equilibrate the volatile components between the liquid and the air in the vial. An aliquot of up to 400 microliters of the headspace was collected by inserting a syringe through the septum of the vial and pulling the headspace sample into the syringe. The sample was then injected directly into the GC.

Concentrations were measured based on an external standard calibration. Known concentrations of benzene, ethylbenzene, meta & para-xylene, ortho-xylene, toluene, and trichloroethylene were injected as a calibration gas mixture into the GC. Compound peak area versus standard concentration was used to calculate sample concentrations.

Compound identification was based on comparison of target compound retention times with sample unknowns.

Analytical results for the groundwater sample analyzed by this technique will not necessarily be the same as those obtained by submitting the same groundwater sample for laboratory analysis. Different extraction techniques are used in each case and, although method sensitivities and accuracies are comparable, different results are possible.

The detection limit is the lowest concentration of a compound that can be practicably measured relative to the calibration standard. Detection limits are a function of the injection volume, as well as detector sensitivity. The detection limit is calculated from the current response factor, the sample size, and the estimated peak area that would have been detected under the given conditions. For this survey, the detection limit for each of the target compounds, was 1 microgram per liter (ug/L).

5 QUALITY CONTROL

Quality control is an essential part of an analytical test methodology. Quality control procedures increase the confidence in the analytical results and are used to evaluate the reproducibility of the data.

The GC was calibrated using a known concentration of each of the target compounds of interest at the beginning of the day before analysis of any samples. The USEPA recommends instrument calibration be performed at least once every 12 hours. The calibration helps to evaluate the operating conditions of the GC.

A chromatographic system blank is analyzed every 10 samples as a means of indication that sample carryover has not occurred. If sample carryover has occurred, the concentration detected in the system blank can be subtracted from any of the subsequent samples containing that compound. A probe rod blank is analyzed prior to sample collection to ensure that rods are free of contamination.

A duplicate sample, which is a second volume of soil-gas, groundwater, or soil collected from the same sample location, is analyzed once every 20 samples, or at least once daily for each survey. Three duplicate groundwater samples were collected at Probe Holes PH-2, PH-14, and PH-20. Duplicates are used to evaluate the reproducibility of the analytical data. The analytical results for each of the respective duplicate samples collected at Probe Holes PH-2, PH-14, and PH-20 were within the specified limit of plus or minus 20 percent.

Appendix A

Gas Chromatograph Field Work Sheets

RECON"SAMPLE ANALYSIS WORK SHEET

Task Phase 1003 Equipment: GC Mewlett-Packard 5890 A Project Number 533435 Project Name Ricker backer ANGB رهمه Nick 10-1-91 C. Operator ate

Carrier Gas: H2 V He Pressure (kPa) 100

Temp. 1 70 Temp. 2 Time 1 10... Time 2 Final Temp.

njector Temp. (C) 200

)ven Temp. Profile
(C)

Datector FID Temp. (C) 300 Column DB-624 / 30 mater 0.53 SE-54 / 15 meter 0.53

Sample I.D.	Probe Hole Number	Depth (ft)	Analysis Time	Inj. Vol.; (ul)	Multiplier	VAC in. Hg	Comments
6:A14K-01			9:19	994	54.0		QC- SYSTEM BLANK
C 001 - 015	1	ı	9:34	100	1.0	(CALIBRATION SID
BLANK-07	.,	١	11:09	00 h	٥.٤٥		GC- ROD BLANK
10,70	10-Hd	16	11:34	. 00 h	5, 25	•	CROUND WATER
(~ · C)	40-H9	٠ 1٤	17:35	400	٥٠٤٥	\	GRONDUATER
0 to-~j	40Hd	الو	12:48	400	٥. عح	ţ	al- Duplicate
57.03	PH-03	30	רס: או	. 004	54.0	1	CROWD VATER
h0-14)	40-Hd	,91	58:31	20	2,0		CRONDUMFER
PLANK. 03	•	١	ht.71	00 h	St'0	3	ac-Sistem BLANK
KT-01	,	,	16.36	001	1.0	١	GC. RT CHECK
					•		•

"duplicate sample analysis C"quality control

page 2 of 5

RECON SAMPLE ANALYSIS WORK SHEET

77	
Task	
1003 Task	
Phase	5890 A
ber 531935 ·	Equipment: GC Hewlett-Packard 5890 A
Number	GC Hew
Project Number	quipment:
ANCB	ы
R. KKAN bucker	•
Name	
Project Name R	Mick Crass
16-8-6	
te 10	·· Onerator

Detector FID Pressure (kPa) 100 rrier Gas: 112.

Temp. 2 Time 2 Final Temp.

و ٥

en Temp, Profile Temp, (C) Time

jector Temp. (C) 200

Column DB-624 / 30 mater 0.53 / SE-54 / 15 mater 0.53

Temp. (C) 300

:nplo I.D.	Proba Hole Number	Depth (ft)	Analysis Time	Inj. Vol.	Multiplior	VAC in. Hg	Comments
81 ANK-04	j	1	7:16	. 00 <i>F</i>	56.0	١	al-system BLANK
510-1008	,	3	LH:(901	1.0	,	CALTERATION STO.
CLANF-05	\	١	50.8	hco	0.25	,	Q1-R00 BLANK
30-00	PH-05	70,	F:57	. 001-	54.0	,	CROUND WATER
70-30	јо-н J	70,	60:01	400	54.0	١	GROUND WATER
(0-m)	F0- H9	90,	10:33	aoh	0.75	١	GROUND WATER
30-70	80-tld	,74	13:04	, 00h	24.0	3	GROUND WATER
80.7.7	PH-05	7.7.	13:40	004	54.0	1	GROUND JATER
01-79	P H -10	74	£ : T -	400	24.0	١	GROUND WATER
	11-11	、みな	14:41	ooh	0.25	3	GROUND WATER
-/-/	P.H - 12	7.74	15:13	400	0.25	1	GROUND WATER.
5 -> U	PH . 12	34,	ار: ۱۲	, 90 h	0.25)	CROUND WATER
<u>.</u>							

-duplicate sample analysis C.quality control

RECON"SAMPLE ANALYSIS WORK SHEET

te 12 4-611 Project Name Rickshucker ANGB P1	Project Number 5'33935 Phase 1003 Task 77
Oparator Nick Cranu Equi	Equipment: GC Hewlett-Packard 5890 A
rrier Gas: H2 / He Pressure (kPa) 100	Detector FID Temp. (C) 300
ren Temp. Profile Temp. 1 70° Temp. 2 (C) Time 1 610 id Time 2	Column DB-624 / 30 mater 0.53 / SE-54 / 15 meter 0.53
jector Temp. (C) 200 . Final Temp.	

						•					
Comments	QCSYSTEM BLANK	CROUND WATER	QL-PUPLECATE	QC-SYSTEM BLANK	QC-RT CHECK			•		•	
VAC in. Hg	1	,))	J						
Multipliar	0,25	٥.25	54.0	\$7'0	0.1					•	
Inj. Vol; (ul)	400	400	400	. doc	100				-		•
Analysis Time	16:00	P4:39	14:91	H5:91	₹9:11						
Depth (ft)	J	18,	,81	,	١						
Probe Hole Number		H-Hd	P1-114	,	}			•			
sample I.D.	8 LANK-06	H-~3	9H- ~ 9	רטאאאט	Å1-01						

*duplicate sample analysis

page 4 of 5

RECON SAMPLE ANALYSIS WORK SHEET

co 10.9-91 Project Name Rickshacks, ANGB	Project Number 533935 Phase 1003 Task 7
Oparator Nik Cravo	Equipment: GC Hewlett-Packard 5890 A
rier Gas: H2 He Pressure (kPa) 100	Datector FID Temp. (C) 300_
on Temp. Profile Temp. 1 70° Temp. 2 (C) Time 1 10 10 10	Column DB-624 / 30 meter 0.53 SE-54 / 15 meter 0.53
Jector Temp. (C) 200 Final Temp.	

ample I.D.	Probe Hole Number	Depth (ft)	Analysis Time	Inj. Vol.; (ul)	Multipliar	VAC in. Hg	Comments
SLANK . CS	,	1	נו:ר	ooh	0.35	,	al-system BLANK
5 TC- 1009.	١	1	ZH:L	100	1.0	3	CALIBRA TION STD
BLANK-09))	8:11	ooh	0.45	١	ac- Rod BLANK
21-73	90-H3	7.5.6	20:6	را 100	25,0		CROUND WATER
91-~)	90-Hd	,01	9:30	o0 h	٥٠,٤٢	ı	CROUND WATER
(, >)	90-Hd	18/	9:50	400	52.0	3	GROUND WATER
ر ۱۰۱۰	Р.Н. об	,91	11:01	. 00h	0.25	١	GROUND WATER
F1 - M.)	P.H - 15	,0%	11:44	904	0.35	1	CROUND WATER
07 7.	81-Hd	, ot	トにやこ	400	0.75	١	CROUND WATER
0 04 7	91-HJ	,00	۲4:41	004	٥. پح	ı	ac-puplescaté
C~.7-	C1-419	3.0	17:40	09h	. 54.0	J	CROUND JATER.
BLANK-10)	a a	٦٠٠٤٦	. ooh	0.25	`	ac-sistem BLANK

duplicate sample analysis Jequality control

recon^msamplė analysis Mork sheet

late 10-9-91	Project Name	Project Name Rickshirker ANGB	- 1	Project Number 533935 Phase 1003 Task	۲
:C'Operator Nick	Nick Cruno	-	Equipm	Equipment: GC Hewlett-Packard 5890 A	
arrier Gas; H2 / He	. 1	Pressure (kPa) 100	•	Detector FID Temp. (C) 300	
ven Temp, Profile Temp. (C) Time	Temp. 1 70 Time 1 10 -, 5		2	Column DB-624 / 30 matar 0.53 SE-54 / 15 meter 0.53	
njector Temp. (C)	200	Final Temp.	Temp.		

	-	 		•				
Comments	ac-at check					-	•	
VAC in, Hg	١							
Multiplior	1.0.				•		•	
Inj. Vol; Multiplior (ul)	100		•					
Analysis	13:08							
Depth (ft)	١							
Probe Hole Number	1							
Sample I.D.	RT-03							

"duplicate sample analysis" Cuquality control

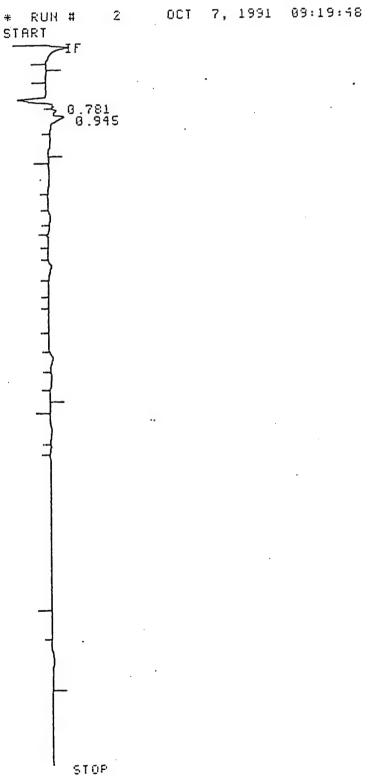
Appendix B

Gas Chromatograph Plots

- + TIME BREAK
- OP # 7

DEFAULT SAMPLE INFORMATION USE SAMPLE TABLE IN MANUAL RUN CY/H+J:

ISTO AMT 00.0000E+00]: SAMPLE AMT [0.0000E+00]: MUL FACTOR [2.5000E-01]: RECALIBRATION EY/N*]: NAME: BLANK-01 REPORT MEMO:



RUN# 2 OCT 7, 1991 09:19:48

SAMPLE NAME: BLANK-01

SIGNAL FILE: B:Q362ACB5.BNC

MATHES RECON MULTIMEDIA ANALYSIS

NO CALIB PERKS FOUND

AREA%

RRER% AREA TYPE WIDTH RT 8677 PU .079 6.07259 .781 .215 18.92742 UU 27045 .945

TOTAL AREA= 35722 MUL FACTOR=2.5000E-01 · DEFAULT SAMPLE INFORMATION USE SAMPLE TABLE IN MANUAL RUN CY/N*J:

ISTO AMT [0.0000E+00]:
SAMPLE AMT [0.0000E+00]:
MUL FACTOR [2.5000E-01]:
RECALIBRATION [Y/N*]:
NAME: BLANK-02
REPORT MEMO:

0.732

* RUN # 4 OCT 7, 1991 11:09:20 START コンショリリ シェビリロギー・エンド マナル・マーンびじゅういけん

RUN# 4 OCT 7, 1991 11:09:20

SAMPLE NAME: BLANK-02

SIGNAL FILE: 8:03620662.BNC

MATHES RECON MULTIMEDIA ANALYSIS

NO CALIB PEAKS FOUND

AREA%

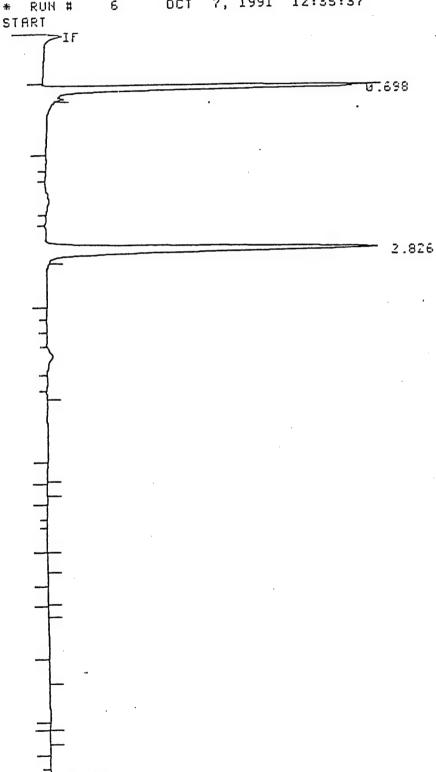
RT AREA TYPE WIDTH AREA% .732 9054 BU .144 25.00000

TOTAL AREA: 9054 MUL FACTOR:2.5000E-01

```
DEFAULT SAMPLE INFORMATION
USE SAMPLE TABLE IN MANUAL RUN [Y/N*]:
```

ISTO AMT 60.0000E+00]: SAMPLE AMT [0.0000E+00]: MUL FACTOR [2.5000E-01]: RECALIBRATION CY/N*J: NAME: 6W-82 REPORT MEMO: PH-02

* RUN # 6 OCT 7, 1991 12:35:37



Closing ...gral file B:03620AFA.SHC

RUN# \$ 007 7, 1991 12:35:37

SAMPLE MAME: SU-02

PH-02

SIGNAL FILE: B:Q362DA9A.BNC

MATHES RECON MULTIMEDIA ANALYSIS

ESTO-AREA

ug/L NAME AREA WIDTH CAL# RT TYPE .000

70913 .061 .698 PB .000 INT. STD.

99953 .089 3R PB 2.826

TOTAL AREA = 178866 MUL FACTOR=2.5000E-01

> ug/L NAME GRP# 0.0000E+00 TOTAL XYLENES 1

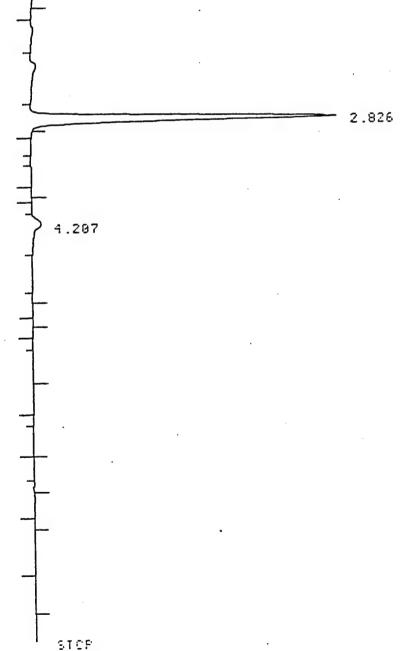
```
DEFAULT SAMPLE INFORMATION
USE SAMPLE TABLE IN MANUAL RUN CYZN+J:
```

ISTO AMT C0.0000E+00]:
SAMPLE AMT C0.0000E+00]:
MUL FACTOR C2.5000E-01]:
RECALIBRATION CY/N*]:
NAME: GU-02D
REPORT MEMO: PH-02

* RUN # 7 OCT 7, 1991 12:48:59
START

IF

0.700



Places stempt feld Det

PUN# 7 0CT 7, 1991 12:48:59

SAMPLE MAME: GW-020

PH-02

SIGNAL FILE: B:Q362008C.BNC

MATHES RECON MULTIMEDIA ANALYSIS

ESTO-AREA

ug/L NAME AREA WIDTH CAL# RT TYPE .000 .700 PU 68409 .059 .000 INT. STD. 3R 85853 .082 2.826 PB .000 5293 .155 UP 4.207

TOTAL AREA= 159555 MUL FACTOR=2.5000E-01

> ug/L NAME GRP# 0.0000E+00 TOTAL XYLENES 1

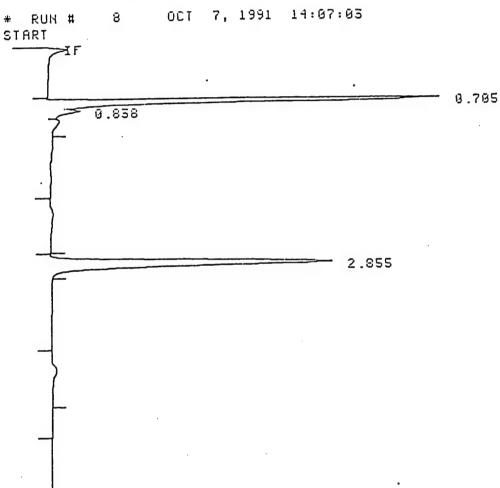
```
+ OF # 7
```

DEFAULT SAMPLE INFORMATION USE SAMPLE TABLE IN MANUAL RUN CY/N+):

ISTO AMT [0.0000E+00]: SAMPLE AMT [0.0000E+00]: MUL FACTOR [2.5000E-01]: RECALIBRATION CY/N*3: NAME: GW-03 REPORT MEMO: PH-03

STOP

* RUN #



RUN# 8 OCT 7, 1991 14:07:03

SAMPLE NAME: GW-03

PH-03

SIGNAL FILE: B:Q362F008.BHC

MATHES RECOM MULTIMEDIA ANALYSIS

ESTO-AREF RT .705 .838	TYPE	. 83711 6755			.000 000.	NAME
2.855	PB	87170	.690	3R	.000	INI. SID.

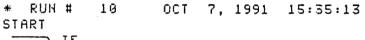
TOTAL AREA= 177636 MUL FACTOR=2.5000E-01

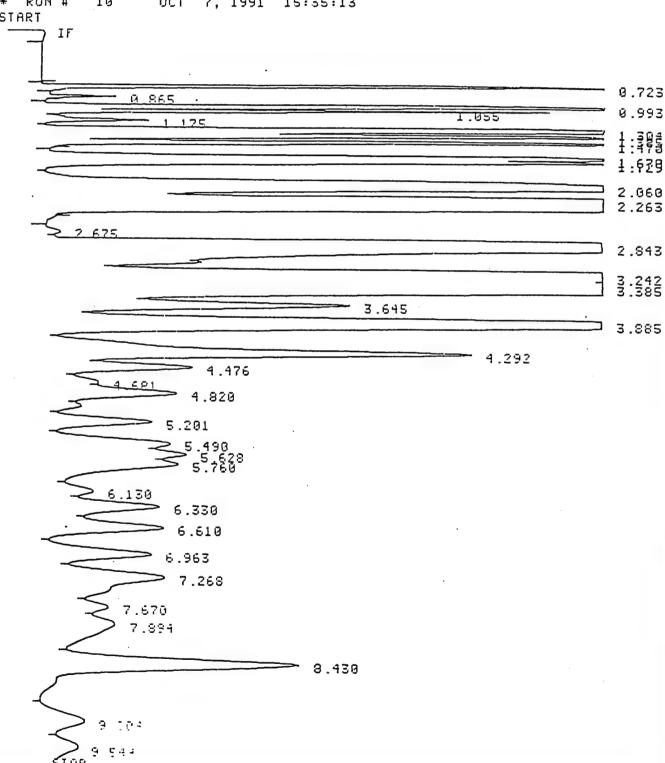
> GRP# Ug/L NAME 1 0.0000E+00 TOTAL XYLENES

```
DEFAULT SAMPLE INFORMATION
USE SAMPLE TABLE IN MANUAL RUN [Y/N*]:
```

ISTD AMT [0.0000E+00]: SAMPLE AMT [0.0000E+00]: MUL FACTOR [2.5000E-01]: 2 RECALIBRATION CY/N*]: NAME: GW-04

REPORT MEMO: PH-04





Closing signal File 8:03630482.8NC

groups and the second

RUN# 18

SAMPLE NAME: GW-04

PH-84

SIGNAL FILE: B:Q36304B2.BNC

MATHES RECON MULTIMEDIA ANALYSIS

ESTO-AREA						NAME
RT T	YPE	AREA	WIDTH C	AL#	- J	RAPIL
.723	PB	109018	.021		.000	
.865	88	6165	.025		.003	
.993	εU	196771	.029		.000	
1.055	UU	50451	.029		.000	
1.175	UP	16644	.045		.009	
1.504	PU	342242	.041		.000	
1.385	บบ	112840	.042		000	
1.470	VВ	97149	.039		.000	
1.670	BU	165499	.053		.000	
1.729	UP	198061	.050		.000	
2.060	PU	425747	.105		.000	
2.263	VB	1761999	.085		.000	7.05
2.263	PU	5987	.070	2	54.948	TCE
2.843	υÜ	817443	.115	3R	.000	INT. STD.
3.242	ŪÜ	745023	.094		.000	
	υU	1091792	.116		.000	
3.385	ΰŪ	116194	.111		.000	
3.645	ÜÜ	449209	.128		1385.000	Toluene
3.885	ÜÜ	190415	.129		.000	
4.292	UU	62069	.128		.000	
4.476	υU	20031	.182		.000	*
4.681	ÜÜ	67472	.145		.000	
4.820	ÜÜ	50624	.155		.000	
5.201	VV	61399	.138		.088	
5.490	ŲÜ	65131	.131		.000	
5.628	บบ	84455	.179		.000	
5.760 6.130	UU	28627	.155		.000	
	UU	69667	.171		.000	
6.330 6.610	ÜÜ	69522	.154		.000	
	ÜÜ	64344	.168		.000	
6.953	ŲŲ	120983	.282	5	392.725	ETHYLBENZENE m + P - Xy/e~e
7.268	บับ	42048			135.000	
7.678	VU	88864		حيد		-HEP-HILLEHE
7.894	UP	169007			.000	
8.430 9.204	PU	38319		7	123.556	0-XAFENE
=	I UH	35276			.000	
9.544	T OU	55210	,			

TOTAL AREA=7944448 MUL FACTOR=2.0000E+00

> ug/L HAME GRP# 4.0956E+02 TOTAL XYLENES

DEFAULT SAMPLE INFORMATION
USE SAMPLE TABLE IN MANUAL RUN CYZN+J:

ISTO AMT [0.0000E+00]:
SAMPLE AMT [0.0000E+00]:
MUL FACTOR [1.0000E+01]: .25
RECALIBRATION [Y/N*]:
NAME: BLANK-03
REPORT MEMO: PH-MEM

*
* RUN # 12 OCT 7, 1991 16:24:48
START

IF

STOP

PUN# 12 OCT 7, 1991 16:24:48

SAMPLE NAME: BLANK-03

SIGNAL FILE: B:03631051.BHC

MATHES RECON MULTIMEDIA ANALYSIS

NO RUN PERKS STORED

```
+ OP # 7
```

DEFAULT SAMPLE INFORMATION USE SAMPLE TABLE IN MANUAL RUN CYZH+J: ISTO AMT [0.0000E+00]: SAMPLE AMT [0.0000E+00]: MUL FACTOR [1.0000E+00]: RECALIBRATION EY/N*J: HAME: RT-01 REPORT MEMO: OCT 7, 1991 16:49:03 * RUN # 14 START IF 0.873 1.965 2.169 2.595 2.805 4.857 7.274 7.684 9.173 STOP

(lasing signa: -::e 8:03631600.880

SAMPLE NAME: RT-01

SIGNAL FILE: B:03631600.BNC

MATHES RECON MULTIMEDIA ANALYSIS

EST	D-AREA			UTETU	COLH	ug/L	NAME
	RT	TYPE	AREA	WIDTH	UNLH		
	.873	SHB	27081440	.031		.000	
		PB	71868	.056		.000	
	1.965		330730	.852	1	466.685	BENZENE
	2.169	88	_	•	2	458.019	TCE
	2.595	PB	66477	.058		.000	INT. STD.
	2.805	FB	39011	.072	ΞR	•	
	3.948	PB	298462	.082	4	459.967	TOLUENE
	-	FB	57835	.101		.000	
	4.857	• –		.137	5	508.130	ETHYLBENZENE
	7.274	PU	313070		_	529.990	M&P-XYLENE
	7.684	ŲB	329346	.143		•	
	9 173	PB	335121	.167	7	540.284	O-XYLENE

TOTAL AREA=2.8923E+07 MUL FACTOR=1.0000E+00

> HRME ug/L GRP# 1.0703E+03 TOTAL XYLENES

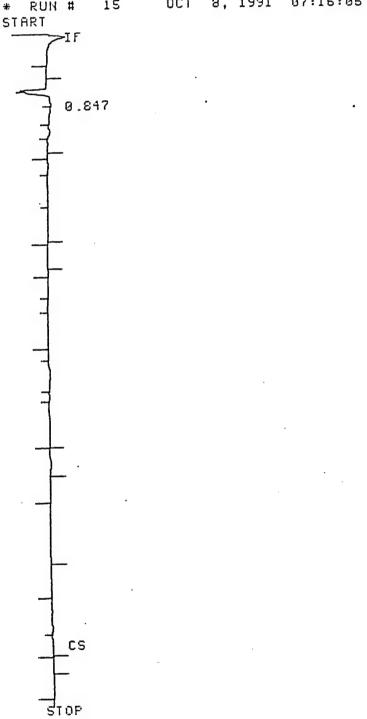
14-1-1

ខណ្ឌ ដ

DEFAULT SAMPLE INFORMATION USE SAMPLE TABLE IN MANUAL RUN EY/N+3:

ISTD AMT [0.0000E+00]: SAMPLE AMT [0.0000E+00]: MUL FACTOR C1.0000E+00 J: .25 RECALIBRATION EY/N*J: NAME: BLANK-04 REPORT MEMO:

15 OCT 8, 1991 07:16:05 # RUN #



Closing signal file B:Q363E137.680

2001 4 10.000 PUN# 15 0CT 8, 1991 07:16:06

SAMPLE NAME: BLANK-04

SIGNAL FILE: B:Q363E137.BNC

MATHES RECON MULTIMEDIA ANALYSIS

NO CALIB PEAKS FOUND

AREA%

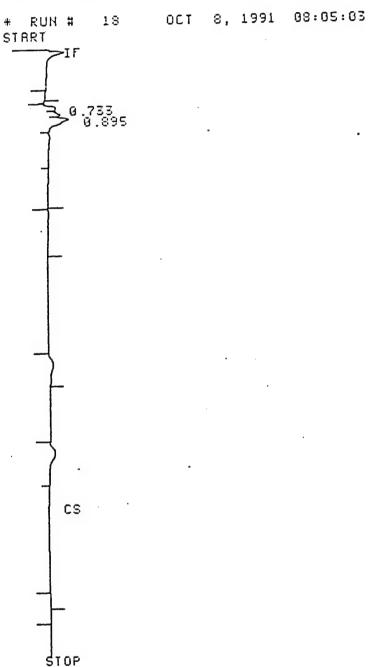
** RT AREA TYPE WIDTH AREA% .847 10106 BU .141 25.00000

TOTAL AREA: 10106 MUL FACTOR:2.5000E-01

```
DEFAULT SAMPLE INFORMATION
USE SAMPLE TABLE IN MANUAL RUN CY/N+J:

ISTO AMT C0.0000E+00 J:
SAMPLE AMT C0.0000E+00 J:
MUL FACTOR C2.5000E-01 J:
RECALIBRATION CY/N+J:
NAME: BLANK-05
REPORT MEMO:

* RUN # 18 OCT 8, 1991 08:05:03
```



Closing signal file B:Q363ECB0.BNC

SAMPLE HAME: SLAHK-05

SIGNAL FILE: B:Q363ECB0.BNC

MATHES RECON MULTIMEDIA ANALYSIS

NO CALIB PERKS FOUND

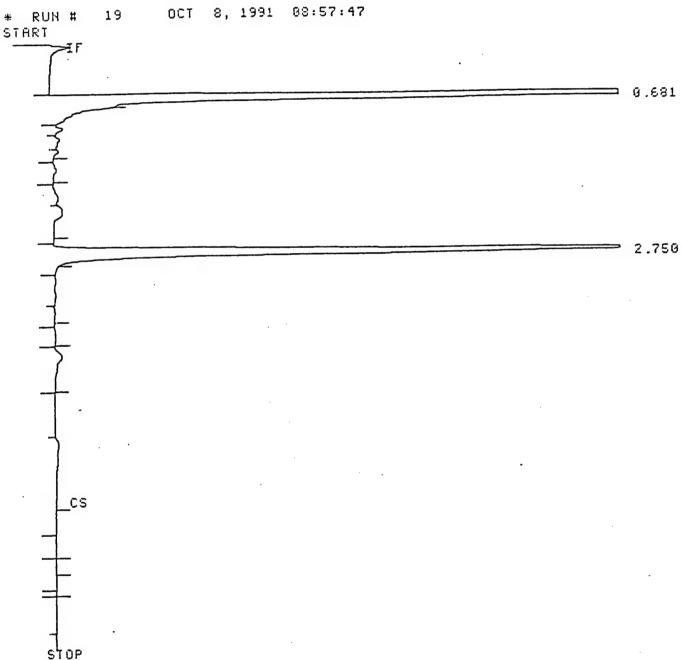
AREA%

RT AREA TYPE WIDTH AREA% .733 2919 BU .065 5.36169 .895 8552 UU .098 18.63830

TOTAL AREA: 11471 MUL FACTOR:2.5008E-01

```
DEFRULT SAMPLE INFORMATION
USE SAMPLE TABLE IN MANUAL RUN [Y/N+]:
```

```
ISTO AMT [0.0000E+00 ]:
SAMPLE AMT [8.9888E+00 ]:
MUL FACTOR [2.5000E-01 ]:
RECALIBRATION [Y/N*]:
NAME: GW-05
REPORT MEMO: PH-05
```



Closing signal file B:Q363F900.BMC

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SAMPLE MAME: GW-05

PH-05

SIGNAL FILE: B:Q363F90C.BNC

MATHES RECON MULTIMEDIA RNALYSIS

ESTD-AREA

RT TYPE - AREA WIDTH CAL# ug/L NAME

.681 PB 467780 .058 .000

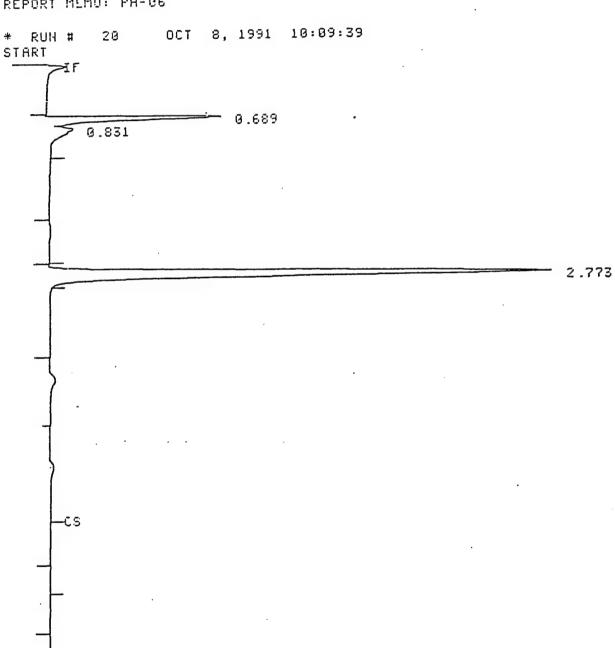
2.750 PB 193914 .086 3R .000 INT.STD.

TOTAL AREA= 661614 MUL FACTOR=2.5000E-01

GRP# Ug/L NAME .
1 0.0000E+00 TOTAL XYLENES

```
DEFAULT SAMPLE INFORMATION
USE SAMPLE TABLE IN MANUAL RUN CY/N*J:
```

```
ISTO AMT C0.0000E+00 ]:
SAMPLE AMT C0.0000E+00 ]:
MUL FACTOR C2.5000E-01 ]:
RECALIBRATION CY/N*]:
NAME: GW-06
REPORT MEMO: PH-06
```



Closing signal file B:036409E4.BNC

RUN# 27 OCT 8, 1991 15:43:63

SAMPLE NAME: GW-13

PH-13

SIGNAL FILE: 8:Q3645808.BNC

MATHES RECON MULTIMEDIA AHALYSIS

ESTD-AREA	1					NOME
	TYPE	AREA	HICIU	CAL#	ug/L	HAME
.690	PU	166080	.062		.000	
.815	υŬ	26657	.064		.000	
.944	ŲŲ	160558	.072		.090	
	ŬŬ	9268	.069		.000	
1.117		174431	.084		.000	
1.247	UU	22550	.864		.000	
1.403	υυ	87859	.105		.000	
1.605	UU	88901	.105			
1.988	00		.095	1	167.764	BENZENE
2.179	ŲВ	423715	.109	3R	.000	INT.STD.
2.765	P8	230514		O1X	.000	
3.131	BV	132817	.100		.000	
3.265	UU	191772	.117			
3.510	UP	6518	.097		.000	
3.758	PP	43936	.131		.666	
4.111	PU	9306	.160		.000	
4.328	VV	7593	.119		.000	
5.568	PU	17465	.363		.000	
7.549	PB	41234	.205	ميد	-11.350	· TOTAL THE ETTE
8.022	88	1329	.849	•	.000	
0.024	0.0					

TOTAL AREA=1842503 MUL FACTOR=2.5000E-01

> ug/L NAME GRP# 1.4368E+01 TOTAL XYLENES

```
DEFAULT SAMPLE INFORMATION
USE SAMPLE TABLE IN MANUAL RUN CV/H+J:

ISTD AMT C0.0000E+00 J:
SAMPLE AMT C0.0000E+00 J:
MUL FACTOR C2.5000E-01 J:
RECALIBRATION CY/N*J:
NAME: BLANK-06
REPORT MEMO:

# RUN # 28 OCT 8, 1991 16:00:04
START

IF

0.749
```

Closing signal file 8:03645005.BNC

0S

STOP

0, 1001 NUMBER 69

SAMPLE NAME: BLANK-06

SIGNAL FILE: 8:Q3645C05.BNC

MATHES RECOM MULTIMEDIA ANALYSIS

NO CALIB PEAKS FOUND

AREA%

AREA% MIDIH AREA TYPE P.T 25.00000 5347 BB .095 .749

TOTAL AREA= 5347 MUL FACTOR=2.5000E-01

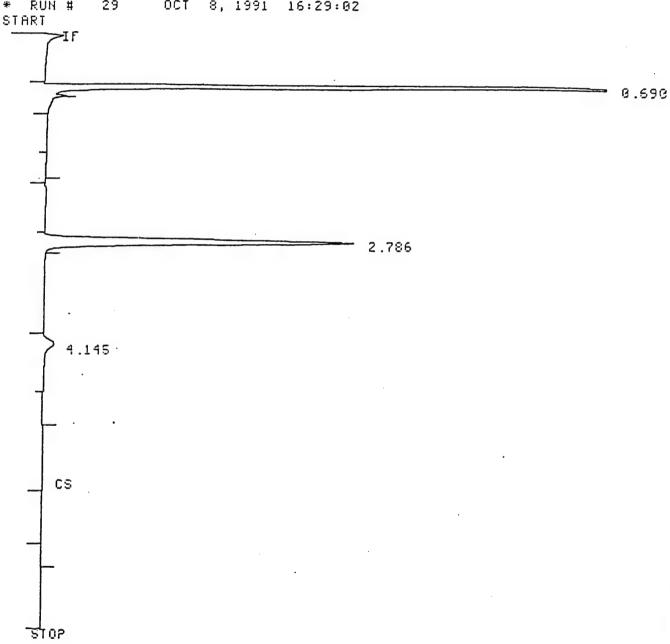
c....: =

```
+ 02 = 7
```

```
DEFRULT SAMPLE INFORMATION
USE SAMPLE TABLE IN MANUAL RUN [Y/N*]:
```

```
ISTD AMT [0.9000E+00 ]:
SAMPLE AMT CO.0000E+00 ):
MUL FACTOR [2.5000E-01 ]:
RECALIBRATION EY/N*1:
NAME: GU-14
REPORT MEMO: PH-14
```

* RUN # 29 OCT 8, 1991 16:29:02



Closing signal file B:Q3646267.886

SAMPLE NAME: GU-14

PH-14

SIGNAL FILE: B:Q36462CF.BNC

MATHES RECOM MULTIMEDIA ANALYSIS

ESTD-AREA

ug/L NAME RT TYPE AREA WIOTH CAL# .690 PB 118099 .845 .000 .000 INT.STD. 87224 .082 3R 2.785 PB 5427 .143 .000 4.145 BP

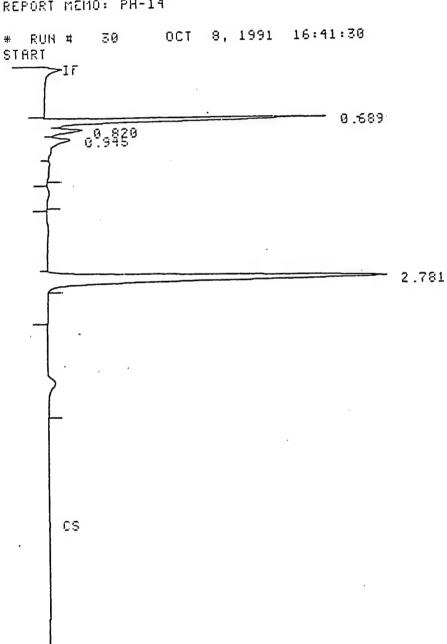
TOTAL AREA= 210750 MUL FACTOR=2.5000E-01

> GRP# ug/L NAME 1 0.0000E+00 TOTAL XYLENES

```
+ '68 # 7
```

```
DEFAULT SAMPLE INFORMATION
USE SAMPLE TABLE IN MANUAL RUN CY/H+J:
```

```
ISTO AMT C0.0000E+00 ]:
SAMPLE AMT C0.0000E+00 ]:
MUL FACTOR C2.5000E-01 ]:
RECALIBRATION CY/N*]:
NAME: GU-14D
REPORT MEMO: PH-14
```



Closing signal file B:Q36465BC.BNC

STOP

KONA 50 - 501 8, 1991 15:31:50

SAMPLE MAME: GW-140

PH-14

SIGNAL FILE: B:Q364658C.BNC

MATHES RECON MULTIMEDIA ANALYSIS

ESTO-AREA

DI	TYPE	AREA	HIDIH	CRL#	ug/L	NHITE
		• • • •	.048		.000	
.689	۴IJ				.000	
.820	99		.057		.000	
.945	UU	9866	.107			TUT 676
2 781		98259	.085	3R	.688	INT.STO.

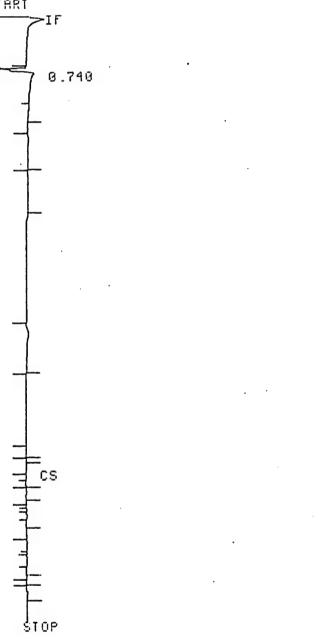
TOTAL AREA= 162816 MUL FACTOR=2.5000E-81

> GRF# Ug/L NAME 1 0.0000E+00 TOTAL XYLENES

DEFAULT SAMPLE INFORMATION
USE SAMPLE TABLE IN MANUAL RUN CY/N+J:

ISTO AMT C0.0000E+00]:
SAMPLE AMT C0.0000E+00]:
MUL FACTOR C2.5000E-01]:
RECALIBRATION CY/N+]:
HAME: BLANK-07
REPORT MEMO:

* RUN # 31 OCT 8, 1991 15:54:08 START



Closing signal file B:Q3648881.BMC

SAMPLE HAME: BLANK-07

SIGNAL FILE: B:Q36468B1.BNC

MATHES RECON MULTIMEDIA ANALYSIS

NO CALIB PERKS FOUND

AREA%

AREA TYPE WIOTH

AREA%

RT .740

1951 PU .06

.061 25.00000

ȚOTAL AREA= 1951 MUL FACTOR=2.5000E-01

.

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DEFAULT SAMPLE INFORMATION USE SAMPLE TABLE IN MANUAL RUN EY/N+J:

ISTO AMT [0.0000E+00]:
SAMPLE AMT [0.0000E+00]:
MUL FACTOR [2.5000E-01]: 1
RECALIBRATION [Y/N*]:
HAME: RI-02
REPORT MEMO:

RUN # 32 OCT 8, 1991 17:05:24 START) IF 0.870 1.960 2.164 2.594 2.805 ⊒ 3.956 4.871 CS 7.303 7.715 9.215 STOP

Closing signal file 8:03646B55.BNC

SAMPLE MAME: RT-02

SIGNAL FILE: B:Q3646B55.BNC

MATHES RECON MULTIMEDIA ANALYSIS

ESTO-AREA	1					NOME
RT	TYPE	AREA	MIDIH	CAL#	ug/L	HAME
.870	SBB	32647792	.031		.000	
1.988	FP	74315	.057		.000	
2.164	 ₽В	316196	.052	1	500.775	BENZENE
		74970	.059	2	586.968	TCE
2.594	PU	35025	.075	3R	.000	INT.STD.
2.805	ŲΒ			4	504.934	TOLUENE
3.956	89	325717	.082	. 7		, , , , , , , , , , , , , , , , , , , ,
4.871	68	79946	.100	_		ETHYLBENZENE
7.303	PU	375398	.137	5	540.314	
7.715	UB	397295	.143	6	. 553.761	M&P-XYLENE
9.215	PB	411078	.168	7	554.828	O-XYLENE

TOTAL AREA=3.4729E+07 MUL FACTOR=1.0000E+00

> ug/L NAME GRP# 1.1086E+03 TOTAL XYLENES

DEFAULT SAMPLE INFORMATION USE SAMPLE TABLE IN MANUAL RUN EYZH+J:

ISTO AMT C0.0000E+00 J:
SAMPLE AMT C0.0000E+00 J:
MUL FACTOR C1.0000E+00 J: .25
PECALIBRATION CY/N*J:
NAME: BLANK-08
PEPORT MEMO: '

REPORT MEMO: ' + RUN # 33 OCT 9, 1991 07:13:21 START \sim IF CS STOP

Closing signal file B:Q3653212.BNC

RUN# 33 001 9, 1991 07:13:21

SAMPLE HAME: BLANK-08

SIGNAL FILE: B:03653212.BNC

MATHES RECON MULTIMEDIA ANALYSIS

NO RUN PEAKS STORED

DEFAULT SAMPLE INFORMATION USE SAMPLE TABLE IN MANUAL RUN EY/M+J:

ISTO AMT C0.0000E+00]:
SAMPLE AMT C0.0000E+00]:
MUL FACTOR C2.5000E-01]:
RECALIBRATION CY/H+]:
NAME: BLANK-009
REPORT MEMO:

* RUN # 35 OCT 9, 1991 08:11:40 START ≻I F 0.820 CS STOP

Closing signal file B:Q3653FBD.BNC

RUN# 35 OCT 9, 1991 08:11:40

SAMPLE NAME: BLANK-009

SIGNAL FILE: B:Q3653FBQ.BNC

MATHES RECON MULTIMEDIA ANALYSIS

NO CALIB PEAKS FOUND

AREA%

"RT AREA TYPE WIDTH AREA% .820 10480 88 .257 25.00000

TOTAL AREA= 10480 MUL FACTOR=2.5000E-01 RUN# 20 OCT 8, 1991 10:09:39

SAMPLE NAME: GU-06

PH-06

SIGNAL FILE: B:Q36409E4.BHC

MATHES RECON MULTIMEDIA ANALYSIS

ESTO-AREA

.689	TYPE PU	AREA 36673 12709	WIDTH .052 .136		.000 .000	NAME
.831 2.773	UB PB	149907	.088	3R	.000	INT.STD.

TOTAL ARER= 199289 MUL FACTOR=2.5000E-01

GRP# Ug/L NAME
1 0.0000E+00 TOTAL XYLENES

CS

Closing signal file B:Q3640F77 SHC

STOP

RUN# 21 OCT 8, 1991 10:33:26

SAMPLE NAME: 69-87

PH-07

SIGNAL FILE: B:Q3640F77.BHC

MATHES RECON MULTIMEDIA ANALYSIS

ESTO-AREA

-AREH RT TYPE AREA WIOTH CAL# UG/L NAME .689 PB 43620 .063 .000 .689 PB 43620 .063 .000

2.774 PB 169196 .085 3R .000 INT.STD.

TOTAL AREA= 232816 MUL FACTOR=2.5000E-01

GRP# Ug/L NAME .
1 0.0000E+00 TOTAL XYLENES

Closing signal tile B:03642483.BNC

Row = Il vii

STOP

22 OCT 8, 1991 12:04:02

HAME: GW-08

L FILE: B:Q36424B3.BHC

HES RECOM MULTIMEDIA ANALYSIS

TO-AREA	AREA	HTOIN	CAL#		HAME
RT TYPE .717 BV 2.807 PB	99359 114725	.072 .087		•	INT.STD.
4 172 PB	6823	.174		.000	

TOTAL AREA= 220908 MUL FACTOR=2.5000E-01

> GRP# Ug/L NAME 1 0.0000E+00 TOTAL XYLENES

```
* OP # 7
DEFAULT SAMPLE INFORMATION
USE SAMPLE TABLE IN MANUAL RUN CYZH+J:
ISTO AMT [0.0000E+00]:
SAMPLE AMT E0.0000E+00 ]:
MUL FACTOR C2.5000E-01 ]:
RECALIBRATION EY/N*]:
NAME: GW-09
REPORT MEMO: PH-09
          23 OCT 8, 1991 13:40:10
# RUN #
START
     >IF
                                                                     0.700
                                                                 2.842
       CS
   , STOP
```

Closing signal file B:Q3643830.800

oga 25 OCT 8, 1991 13:40:10

AMPLE HAME: GW-89

PH-09

SIGNAL FILE: B:Q3643B3C.BNC

MATHES RECON MULTIMEDIA ANALYSIS

ESTO-AREA

RT TYPE AREA WIDTH CAL# ug/L NAME .700 PU 175726 .063 .000 2.842 PB . 164258 .090 3R .000 INT.STD.

TOTAL AREA= 339984 MUL FACTOR=2.5000E-01

GRP# UQ7L NAME .
1 0.0000E+00 TOTAL XYLENES

```
+ OP # 7
DEFAULT SAMPLE INFORMATION
USE SAMPLE TABLE IN MANUAL RUN CY/N+]:
ISTO AMT 00.0000E+00 ]:
SAMPLE AMT [0.0000E+00 ]:
MUL FACTOR [2.5000E-01]:
RECALIBRATION EY/N*]:
MAME: GU-10
REPORT MEMO: PH-18
                  OCT 8, 1991 14:12:57
           24
* RUN #
START
      >IF
         0.969
         1.280
        2.308
                                                            2.845
       CS
```

Closing signal file B:Q36442EA.BNC

STOP

RUN# 24 OCT 8, 1991 14:12:57

SAMPLE NAME: 6W-10

PH-10

SIGNAL FILE: B:Q36442EA.BNC

MATHES RECON MULTIMEDIA ANALYSIS

ESTD-AREA

RT	TYPE	AREA	WIDTH	CAL#	ug/L	HAME
.700	PB	489862	.063		.000	
.969	PP	4949	.108		.000	•
1.280	PU	4523	.104		.000	
2.308	VB	7467	.161		.000	
2.845	88	148413	.092	38	.000	INT.STD.

TOTAL AREA= 655214 MUL FACTOR=2.5000E-01

> GRP# Ug/L NAME 1 0.0000E+00 TOTAL XYLENES

25

LE HAME: GW-11

1

SHAL FILE: B:Q364499D.BHC

ATHES RECON MULTIMEDIA ANALYSIS

ESTO-AREA

NAME AREA WIDTH CAL# ug/L RT TYPE .000

.058 321323 P.B

.698 2.815 INT.STD. .000 .089. 3R 139508 PB

TOTAL AREA = 460831 MUL FACTOR=2.5000E-01

> ug/L NAME GRP# TOTAL XYLENES 0.8000E+00 . 1

```
+ OP # 7
DEFAULT SAMPLE INFORMATION
USE SAMPLE TABLE IN MANUAL RUN CY/H+]:
ISTO AMT [0.0000E+00 ]:
SAMPLE AMT C0.0000E+00 ]:
MUL FACTOR E2.5000E-01 3:
RECALIBRATION EY/N*J:
HAME: GW-12
REPORT MEMO: PH-12
                 OCT 8, 1991 15:12:02
           26
# RUN #
START
      -IF
                                                                      0.682
        0.948
                                                 2.773
       -os
     $102
```

Closing signel file 8:03645003.BNC

50001 =

RUN# 26 OCT 8, 1991 15:12:02

SAMPLE NAME: 6U-12

PH-12

SIGNAL FILE: 8:036450C3.BNC

MATHES RECOM MULTIMEDIA ANALYSIS

ESTD-AREA

212 1115-	•				_	
RI	TYPE	AREA	WIDIH	CAL#	ug∕L	NEME
.682	ВV	303820	.861		.000	
	UÜ	7469	.132	•	.000	
	PB	113302	.088	3R	.000	INT.STD.

TOTAL AREA= 424591 MUL FACTOR=2.5000E-01

> GRP# Ug/L NAME 1 0.0000E+00 TOTAL XYLENES

```
DEFAULT SAMPLE INFORMATION
USE SAMPLE TABLE IN MANUAL RUN EY/N*]:
ISTO AMT [0.0000E+00]:
SAMPLE AMT [0.0000E+00 ]:
MUL FACTOR [2.5000E-01 ]:
RECALIBRATION CY/N*]:
NAME: GU-13
REPORT MEMO: FH-13
                  OCT 8, 1991 15:43:03
           27
# RUN #
START
      >IF
                                                                         0.690
                 1 815
                                                                         0.944
         S 1 117
                                                                         1.247
                  1.403
                                   1.605
                                   1.980
                                                                        2.179
                                                                         2.765
          3.510
                  3.758
          4.111
          4.328
         5.568
        CS
        3.5<u>62</u>
        STOP
```

RUM # 27-882

Closing signal file B:Q3645808.880

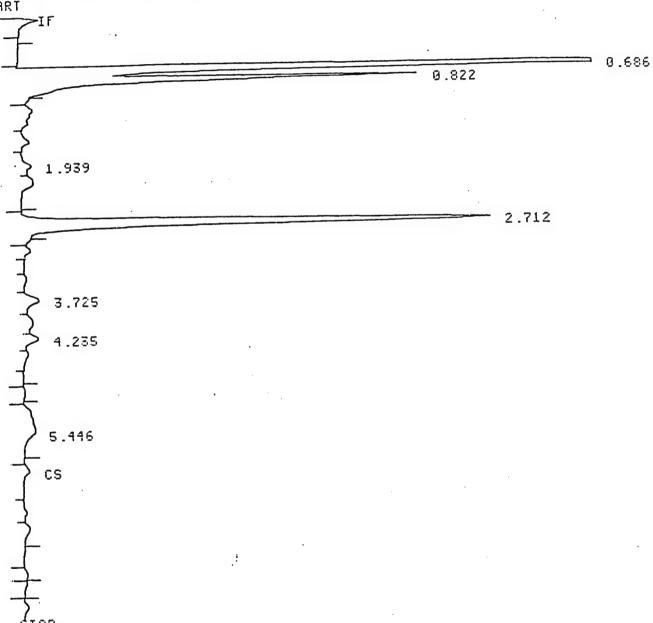
+ 0P # 7

```
* OP # 7
```

```
DEFRULT SAMPLE INFORMATION
USE SAMPLE TABLE IN MARUAL RUN CYZN*J:
```

```
ISTO AMT C0.0000E+00 ]:
SAMPLE AMT C0.0000E+00 ]:
MUL FACTOR C2.5000E-01 ]:
RECALIBRATION EY/N*]:
HAME: GGU-15
REPORT MEMO: PH-06
```

* RUN # 36 OCT 9, 1991 09:05:26 START



Closing signal file B:Q3654058.8KC

SAMPLE NAME: 66W-15

PH-06

SIGHAL FILE: B:Q3654C58.BNC

MATHES RECON MULTIMEDIA ANALYSIS

NO CALIB PEAKS FOUND

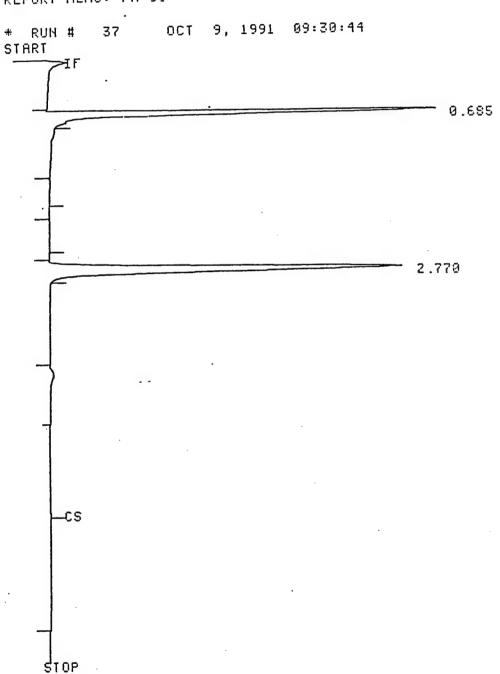
ARE	8%
-----	----

. 1117	0000	TYPE	WIDTH	AREA%
RT	AREA	ITPL		
.686	172472	PU	.075	9.73272
.822	96464	VВ	.071	5.44354
1.939	4312	PV	.118	.24333
2.712	1.41915	PB	.090	8.00836
3.725	7100	PU	.129	.40065
-	6699	บบ	.127	.37803
4.235	14059	68	338	.79336
5,445	14057	00		

TOTAL AREA= 443021 MUL FACTOR=2.5000E-01

```
DEFAULT SAMPLE INFORMATION
USE SAMPLE TABLE IN MANUAL RUN CY/H+J:
ISTO AMT C0.0000E+00 J:
```

ISTO AMT [0.0000E+00]:
SAMPLE AMT [0.0000E+00]:
MUL FACTOR [2.5000E-01]:
RECALIBRATION [Y/N+]:
HAME: GW-16
REPORT MEMO: PH-06



Closing signal file 5:Q3655246.BNC

OCT 9, 1991 09:30:44 RUN# 37

SAMPLE HAME: GW-16

PH-06

SIGNAL FILE: B:Q3655246.BNC

MATHES RECOM MULTIMEDIA ANALYSIS

ESTO-AREA

HAME ug/L AREA WIDTH CAL# RI TYPE .000 .059 78113 .685 FB .000 INT. STD. 3R 104623 .087 PB 2.779

TOTAL AREA= 182736 MUL FACTOR=2.5000E-01

> ug/L NAME GRP# 0.0000E+00 TOTAL XYLENES 1

```
* OP # 7
```

```
DEFAULT SAMPLE INFORMATION
USE SAMPLE TABLE IN MANUAL RUN EY/N+J:
```

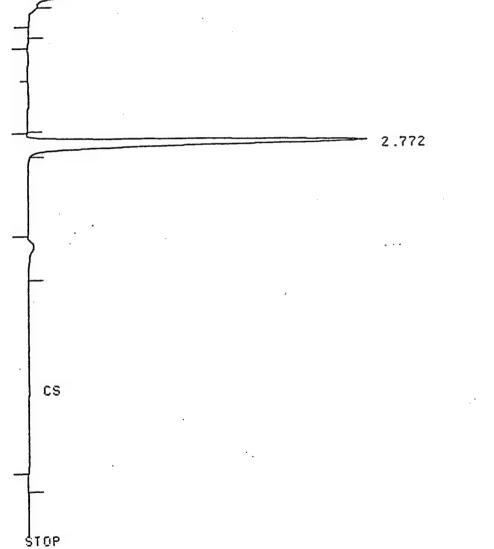
```
ISTO AMT C0.0000E+00 ]:
SAMPLE AMT C0.0000E+00 ]:
MUL FACTOR C2.5000E-01 ]:
RECALIBRATION CY/N*]:
NAME: GW-17
REPORT MEMO: PH-06
```

RUN # 38 OCT 9, 1991 09:50:45

START

IF

0.702



(losing signal file B:Q36556F6.800

OCT 9, 1991 09:50:45 RUH# 38

SAMPLE MAME: 6W-17

PH-05

SIGNAL FILE: B:Q36556F6.BNC

MATHES RECOM MULTIMEDIA ANALYSIS

ESTO-AREA

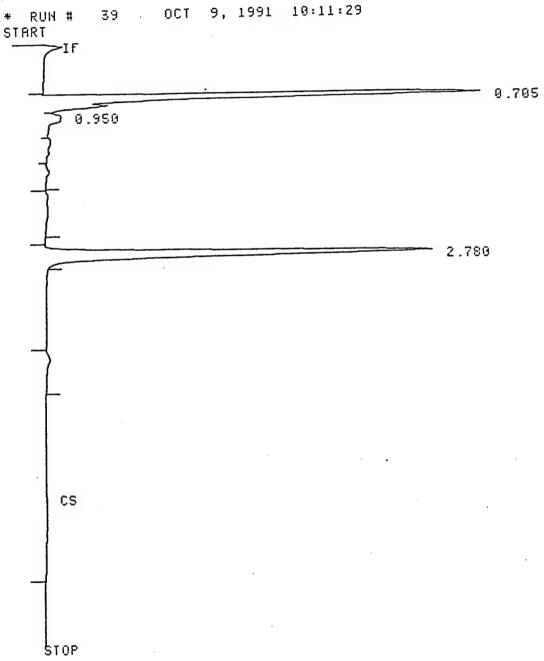
ug/L NAME WIDTH CAL# AREA RT TYPE .000 .075 106548 28 .702 INT. STD. .000 ВR .089 101910 PB 2.772

TOTAL AREA= 208458 MUL FACTOR=2.5000E-01

> NAME ug/L GRP# 0.0000E+00 TOTAL XYLENES 1

```
DEFAULT SAMPLE INFORMATION
USE SAMPLE TABLE IN MANUAL RUN CY/H+J:
ISTO AMT C0.0000E+00 J:
SAMPLE AMT C0.0000E+00 ]:
```

MUL FACTOR C2.5000E-01 3: RECALIBRATION CY/N*J: NAME: GU-18 REPORT MEMO: PH-06



Closing signal file 8:03655803.880

North 27 22 40:11:29

SAMPLE MAME: 6W-18

PH-06

SIGNAL FILE: B:Q3655803.BNC

MATHES RECON MULTIMEDIA ANALYSIS

ESTO-AREA

AREA WIDTH CAL# MAME uazL RT TYPE .000 .705 .079 PU 118826 .950 VV 9454 .145 .000 .088 .000 INT. STO. 115672 2.788 PB

TOTAL AREA= 243952 MUL FACTOR=2.5000E-01

> GRP# Ug/L NAME 1 0.0000E+00 TOTAL XYLENES

≟ HAME: GU-19

AL FILE: 8:0365718A.BHC

.HES RECON MULTIMEDIA ANALYSIS

TOTAL AREA= 284417 MUL FACTOR=2.5000E-01

GRP# Ug/L NAME .
1 0.0000E+00 TOTAL XYLENES

```
# 0P # T
DEFAULT SAMPLE INFORMATION
USE SAMPLE TABLE IN MANUAL RUN [Y/H+]:
ISTO AMT [0.0000E+00 ]:
SAMPLE AMT C0.0000E+00 ]:
MUL FACTOR C2.0000E+00 ]: .25
RECALIBRATION CY/N*]:
NAME: GW-28
REPORT MEMO: PH-16
           43 OCT 9, 1991 12:14:56
# RUN #
START
      >IF
          0.950
         1.262
          2.194
                                             2.796
       CS
     STOP
 (losing signal file 8:03657801.5∷0
```

From # 43-000

AMPLE MAME: 6U-20

H-16

\$IGNAL FILE: B:Q36578C1.BNC

MATHES RECON MULTIMEDIA ANALYSIS

ESTO-AREA	ES	T	n-	А	R	Ε	Ĥ	
-----------	----	---	----	---	---	---	---	--

RI	TYPE	AREA	WIDTH	CAL#	ug/L	HAME
.693	PB	136165	.051		.880	
.950		7752	.074		.000	
1.262		6856	.090		.000	
	UB	12185	.110	1	4.373	BENZENE
2.796	PB	106721	.089	. 3R	.000	INT. STD.

TOTAL AREA= 269679 MUL FACTOR=2.5000E-01 .

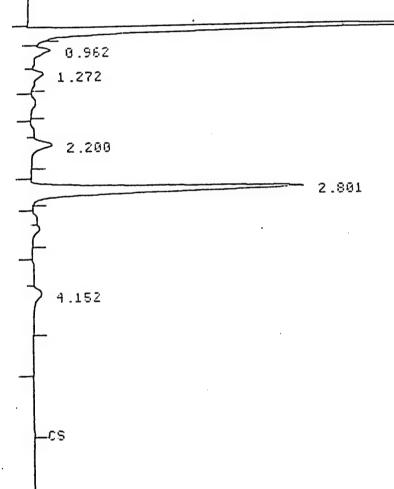
> ug/L NRME GRP# 0.0000E+00 TOTAL XYLENES

```
DEFAULT SAMPLE INFORMATION
USE SAMPLE TABLE IN MARUAL RUN CY/R*J:
```

```
ISTO AMT [0.0000E+00]:
SAMPLE AMT [0.0000E+00]:
MUL FACTOR [2.5000E-01]:
RECALIBRATION [Y/N*]:
NAME: GW-200
REPORT MEMO: PH-16
```

```
# RUN # 44 OCT :9, 1991 12:27:41
START
IF
```

0.712



Closing signal file 8:0365788E.BNC

STOP

RUN# 44 OCT 9, 1991 12:27:41

SAMPLE HAME: GW-200

PH-16

SIGNAL FILE: B:Q365788E.BNC

MATHES RECOM MULTIMEDIA ANALYSIS

ESTO-ARE	7					
	TYPE	AREA	WIDTH	CAL#	ug/L	NAME
.712	28	100928	.068		.000	
.962	BP	3998	.077		.000	
1.272		4428	.105		.000	
2.200	UB	8848	.115	1	3.176	BEHZEHE
2.801	PB	87658	.035	3R	.000	INT. STD.
4.152	UB	5715	.169		.899	

TOTAL AREA = 211567 MUL FACTOR = 2.5000E - 01

> GRP# ug/L NAME 1 0.0000E+00 TOTAL XYLENES

```
DEFAULT SAMPLE INFORMATION
USE SAMPLE TABLE IN MANUAL RUN CY/N+J:
ISTO ANT C0.0000E+00 J:
SAMPLE AMT [0.0000E+00 ]:
MUL FACTOR C2.5000E-01 3:
RECALIBRATION EY/N*]:
HAME: GU-21
REPORT MEMO: PH-17
                 OCT 9, 1991 12:40:11
           45
* RUN #
START
      -IF
                                                                       0.688
                B 815
                                          0.945
                                  1.251
            1.418
                        1.610
                                                                       2.186
                                                                        2.775
                                                                        3:144
         > 3.526
                                        3.765
         4.131
         4.346.
           4.679
         5.331
        cs
          7.068
        STOP
```

Closing signal tile 8:Q3657EAC.BNC

0P # 7

34 45

AMPLE HAME: GU-21

4H-17

SIGNAL FILE: B:03657EAC.BNC

MATHES RECON MULTIMEDIA ANALYSIS

ESTO-AREA RT TYPE 688 PU	AREA 528259	WIDTH .047	CAL#	ug/L .000	HAME
.815 UV .945 UV 1.251 UU	21105 88022 71005 11038	.060 .077 .082 .063		899. 899. 889. 889.	
1.410 UU 1.510 UU 1.937 UU 2.186 UB	58796 68374 479313	.112 .100 .092	1 3R	.000 .000 172.029 .000	BENZEKE INT. STO.
2.775 PU 3.144 VV 3.278 VV 3.526 VV	275309 235747 322150 16869	.103 .118 .118	310	.000 000 000 000	
3.765 UU 4.131 UU 4.346 UU	127435 13022 8604 21776	.129 .181 .132 .175		.090 .999 .909	
4.679 UU 5.331 PB 7.068 UU	3104 22007	.118		.000 6.299	ETHYLBENZENE

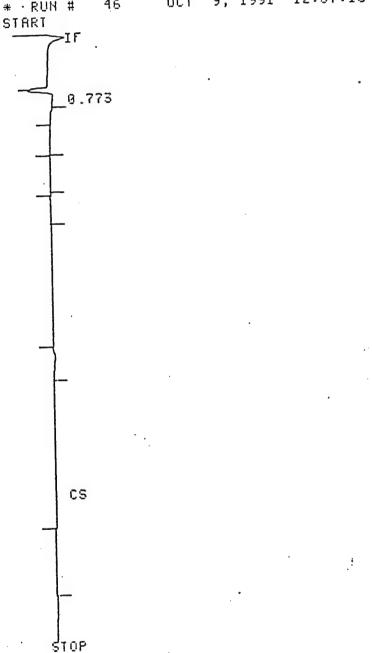
TOTAL AREA=2371934 MUL FACTOR=2.5000E-01

> GRP# Ug/L NAME 1 0.0000E+00 TOTAL XYLENES

DEFAULT SAMPLE INFORMATION USE SAMPLE TABLE IN MANUAL RUN CY/N+]:

ISTO AMT C0.0000E+00 J: SAMPLE AMT [8.0000E+80]: MUL FACTOR [2.5000E-01]: RECALIBRATION [Y/N*]: NAME: BLANK-10 REPORT MEMO:

OCT 9, 1991 12:57:13 * · RUN #



Closing signal file 8:03658288.880

RUH# 46 0CT 9, 1991 12:57:13

SAMPLE HAME: BLANK-10

SIGNAL FILE: B:Q36582AA.BNC

MATHES RECON MULTIMEDIA ANALYSIS

NO CALIB PEAKS FOUND

AREA%

RT RREA TYPE WIDTH RREA% .773 7486 PB .112 25.00000

TOTAL AREA= 7486 MUL FACTOR=2.5000E-01

```
DEFAULT SAMPLE INFORMATION
USE SAMPLE TABLE IN MANUAL RUN EY/H+J:
ISTO AMT 60.0000E+00 ]:
SAMPLE AMT [0.0000E+00 ]:
MUL FACTOR [2.5000E-01 ]: 1
RECALIBRATION CY/N*J:
HAME: RT-03
REPORT MEMO:
                  OCT 9, 1991 13:08:19
           47
* RUN #
START
      IF
                                                                     0.875
                                                    1.974
                                                                    2.180
                                                   2.613
                    2.825
                                                                     3.985
                               4.986
      CS
                                                                      7.352
                                                                      7.767
      STOP
```

Closing signal file 8:03658544.880

SAMPLE NAME: RT-03

SIGNAL FILE: B:Q3658544.BNC

MATHES RECON MULTIMEDIA ANALYSIS

r	S	Т	n	_	A	R	F	Ĥ
r_			$\mathbf{\nu}$			1.	_	, .

COLD INSEL	•					
RT	TYPE	AREA	HIDIH	CRL#	ug/L	NAME
.875	888	30938704	.032		.999	
1.974	PB	81848	.059		.000	
2.180	88	335924	.054	1	482.262	BENZEHE
2.613	BU	81725	.061	2	474.573	TCE
2.825	UB	32022	.976	3R	.000	INT. STD.
3.985	PB	334276	.085	4	453.932	TOLUENE
4.906	PB	78471	.103		.000	
7.352	PU	386095	.141	5	442.059	ETHYLBENZENE
7.767	VB	415050	.147	6	448.594	M&P-XYLENE
9.276	PB	430869	.171	7	427.554	O-XYLENE

TOTAL AREA=3.3115E+07 MUL FACTOR=1.0000E+00

> NAME ug/L GRP# 8.6815E+02 TOTAL XYLENES

APPENDIX B-3

LABORATORY REPORTS FROM 1995 NATURAL ATTENUATION INVESTIGATION



Ref: 95-DK9/vg May 9, 1995

Dr. Don Kampbell

R.S. Kerr Environmental Research Lab

U.S. Environmental Protection Agency

P.O. Box 1198

Ada, OK 74820

THRU: S.A. Vandegrift SV

Dear Don:

This report contains the results of my GC/MSD analysis of methylene chloride extracts of core samples and one free floating product sample (MW-5) from Rickenbacker ANGB for quantitation of benzene, trichloroethylene (TCE), tetrachloroethylene (PCE), toluene, ethylbenzene (EB), p-Xylene (p-X), m-Xylene (m-X), o-Xylene (o-X), 1,3,5-trimethylbenzene (1,3,5-TMB), 1,2,4-trimethylbenzene (1,2,4-TMB), 1,2,3-trimethylbenzene (1,2,3-TMB), 1,2,4,5-tetramethylbenzene, 1,2,3,5-tetramethylbenzene, 1,2,3,4-tetramethylbenzene, naphthalene; 2-methylnaphthalene and 1-methylnaphthalene performed under Service Request #SF-1-118.

The analytical method was a modification of RSKSOP-124. Cool (38°C) on-column injection (0.5 $\mu l)$ was used with electronic pressure control set for a constant flow of 0.9 ml/min. A 30M X 0.25 mm Restek Stabilwax (Crossbonded Carbowax-PEG, 0.5 μm film) capillary GC column with 9 inch long X 0.53 mm ID uncoated capillary precolumn was used. Quantitation was based on calibration curves of selected target ions (2 or 3 ions , total area) for each compound. A high level (5-250 $\mu g/ml$, naphthalenes 0.1-50 ug/ml) and low level (0.05-5 $\mu g/ml$) calibration curve was applied to each sample for quantitation. Complete reports detailing the acquisition method and calibration curves have been recorded. The soil samples were extracted by Mark Blankenship on March 6, 1995 and the free-floating product was received March 2, 1995. All samples were analyzed by GC/MSD on March 15-16, 1995.

If I can be of further assistance, please feel free to contact me.

Sincerely,

David A. Kovacs

xc: R.L. Cosby

J.L. Seeley

G.B. Smith

GC/MSD (SIM)

Vylono	ווו-עאופוופ	4.77E+00	5.18E+00	7.63E+01	5.08E+01	N/A	QN	BLQ	7.00E-01	3.08E-02	N/A	2.63E+02	5.30E-01
orolly a	D-VAIGILE	4.87E+00	5.14E+00	2.80E+02	5.20E+01	N/A	QN	Q	3.14E-01	1.86E-02	N/A	2.62E+02	4.97E-01
741	Etuyipenzene	4.79E+00	5.04E+00	2.59E+02	5.30E+01	N/A	Q	QN	4.48E-01	BLQ	N/A	2.62E+02	5.10E-01
	loluene	4.76E+00	5.06E+00	9.51E+00	5.27E+01	N/A	QN	BLQ	BLQ	BLQ	. A/N	2.57E+02	5.17E-01
1	밁	4.71E+00	4.87E+00	QN	4.97E+01	N/A	Q	BLQ	QN	QN	A/N	2.38E+02	5.14E-01
1	TCE TCE	4,60E+00	4.81E+00	N	4.98E+01	A/N	Q	ВГО	QN	2	N/A	2.39E+02	5.17E-01
	Benzene	4.81E+00	5.17E+00	QN	5.36E+01	A/N	QN	2	2.06E-01	QN	A/N	2.46E+02	5.08E-01
	Sample	OC (5 ua/ml)	50 in/mi	MW-5 Free Product (un/ml)	50 chk	o 4 ma/ml Nanhthalenes	Method Blank SE-1-118	SS-4		1 6000	10 uo/ml Nanhthalenes	250 mg/ml	0.5 ug/ml

SF-1-118 Dr. D. Kampbell

5/9/95 Page 2 of 3

				,	A/N							
1,2,4,5-MeBenzene	A/N	4.71E+00	1.98E+02	5.08E+01	N/A	ON	QN	3.90E-02	BLQ	N/A	2.76E+02	4.97E-01
1,2,3-TMB	4.61E+00	4.92E+00	3.02E+02	5.25E+01	N/A	QN	Q	1.35E-01	QN	N/A	2.70E+02	5.20E-01
1,2,4-TMB	4.64E+00	5.00E+00	7.68E+02	5.33E+01	N/A	ND	QN	4.87E-01	2.48E-02	N/A	2.76E+02	5.21E-01
1,3,5-TMB	4.70E+00	4.98E+00	4.20E+02	5.30E+01	A/N	QN	QN	1.21E-01	BLQ	N/A	2.71E+02	5.12E-01
o-Xviene	4.81E+00	5.08E+00	5.32E+01	5.31E+01	N/A	QN	QN	6.56E-02	2.10E-02	N/A	2.62E+02	5.20E-01
Sample	OC (5 119/ml)	(m, m, o,	MW.5 Free Product (IId/ml)	50 chk	0.4 ug/ml Naphthalenes	Method Blank SF-1-118	25.4	6.58	1 C. 00	40 mg/ml Naphthalenes	o ugaill rapilliaiches	gm.cz 0.5 ug/ml

1-MeNaphthalene	N/A	N/A	3.21E+00	N/A	9.93E-02	ND	ND	2.33E-02	ND	9.84E+00	N/A	A/N
2-MeNaphthalene	N/A	N/A	4.93E+00	N/A	1.00E-01	ND	"QN	2.37E-02	ND	9.81E+00	N/A	N/A
Naphthalene	N/A	N/A	3.45E+01	N/A	1.00E-01	QN	QN .	8,35E-02	QN	9.79E+00	N/A	N/A
1,2,3,4-MeBenzene	N/A	5.12E+00	1.03E+03	5.44E+01	NA	QN	BLQ	3.13E-02	BLQ	N/A	2.72E+02	5.05E-01
Sample		5.0 ug/ml	(lm/gr		lalenes	;	SS-1		88-3	10 ug/ml Naphthalenes		•

SF-1-118)	
(SR#	
TOC	
FOR	
AFB	
RICKENBACHER	
IIS	

SOLLS	SOLLS FINE KICKENBACHER AFB FOR IOC (SK SF 1-118)	AED FOR	TOC (SK# SF#	(011_1
SAMPLES	SOIL FILTRATES	Ω ₩	TOTAL SOIL STOC	MEAN & TOC
SS-1-1 SS-1-2	0.010	1.505	1.515	1.515 1.514 1.512
SS-2-1 SS-2-2	0.058	0.798	0.856	0.799
SS-3-1 SS-3-2	0.004	1.458	1.462	1.409
LECO		1.020		

WPO33-I 7.7 MG/L 7.9 MG/L

TRUE VALUES: LECO = 1.00 +/- 0.04% C WPO33-I = 7.70 MG/L OC

TPH, mg oil/Kg

650

C 50

2 50

& on Fampleh

% moisture

per Don Kampbell, phone conversation 33

Core Sa	emples MoisTure, %	TPH mg oil/kg	Toc,?
55-1	13,7	< 50	1.5
55-2	17,2	250	0.8
55-3	13.3	250	1.4

	MW- 5 Product	55	-1	55-2 mg/kg	58-	2
Benzene	115/ml < 5	~	5	< 5	< 5	
Toluene	TO	}	,	7	5	
Ethyl benzene	240	}		282	5	
P-xylene	250)			198	1)	
On-xylene	76 (377			443	19	
o-xylene	5))			.41.	12	
1,3,5 THY MB	388			76	5	
1,2,4 TRIMB	740	-		309	16	
1,2,3 THIMB	540			85	< 5	
1,2,3,5 TeTra MB	440			25	< 5	
1,2,3, 4 TeTra MB	860	}		44	< 5	
TCE	< 5			< 5	< 5	
PCE	25	1		45	< 5	
				^		
		1	BTE X	1 TPH	Fm Kamp	helf
			Extra	ctions	4/14)	95

Field Data Rickenbacker ANG, Ohio								
Sample	Date	Carbon Dioxide mg/l	Total Alkalinity mg/l	Ferrous Iron mg/l	Hydrogen Sulfide mg/l			
ES mp-5S	2-27-95	150	293	<.05	-			
ES mp-5D	2-27-95	228	370	1.9	<.1			
ES mp-10D	2-27-95	296	426	1.8	<.1			
ES mp-10S	2-27-95	192	314	.3				
ES mp-8D	2-27-95	100	380	3,1	<.1			
ES mp-11D	2-27-95	220	376	2.5	<.1			
ES mp-7S	2-28-95	124	296	.1				
ES mp-7D	2-28-95	208	212	1.6	<.1			
ES mp-6D	2-28-95	168	385	.9	<.1			
ES mp-9D	2-28-95	150	393	.8	<.1			
ES mp-9S	2-28-95	188	254	.1	-			
MW-12	2-28-95	300	347 •	.3	-			
ES mp-3D	2-28-95	216	389	1.0	-			
MW-6	2-28-95	226	387	<.05				
MW - 11	2-28-95	94	211	<.05	-			
ES mp-4D	2-28-95	220	394	.1	-			
ES mp-4S	2-28-95	186	378	.8	-			
ES mp-14S	2-28-95	288	440	3.2	<.1			
ES mp-14D	2-28-95	214	393	1.4	<.1			
ES mp-2D	2-28-95	190	344	1.3	-			
ES mp-1D	2-28-95	170	336	.1	•			
MW-3	2-28-95	276	368	<.05	-			
MW-8	2-28-95	208	391	<.05	-			
ES mp-13S	2-28-95	330	386	3.2	.1 .			

Sample	mg/L _Cl_	mg/L SO ₄ =	mg/L NO ₂ + NO ₃ (N)	mg/L NH ₃
ESMP-1D	10.2	41.4	<.05	0.22
ESMP-2D	13.1	56.7	0.09	0.10
ESMP-4D	7.31	87.1	0.08	0.06
ESMP-4S	5.10	144	0.08	0.10
ESMP-13D	17.0	54.9	0.09	0.07
ESMP-13D Field Dup	17.8	57.9	0.09	0.07
ESMP-13D Field Ddp	23.5	38.3	0.09	0.43
ESMP-13S Dup	23.3	38.2		
ESMP-14D	17.0	58.7	0.11	0.09
ESMP-14D Dup			0.11	0.08
ESMP-14D Field Dup	16.3	57.1	0.11	0.08
ESMP-14S	7.29	19.2	0.09	0.13
ESMP-14S Field Dup		SAMPLE -	0.09	0.12
ESMP-15D	20.1	264	0.41	0.10
ESMP-15S	16.1	206	0.32	0.19
ESMP-15S Dup	15.6	206		
ESMP-16D	19.5	938	<.05	0.75
ESMP-16D Field Dup	20.7	895	<.05	0.74
ESMP-16S	53.3	208	<.05	1.22
ESMP-17S	7.26	41.0	0.09	0.29
MW-2	7.79	61.2	0.09	0.05
MW-3	21.1	127	0.10	0.06
MW-4	19.0	103	0.15	0.09
MW-4 Dup			0.15	0.09
MW-5	8.00	6.57	0.08	0.45
MW-8	8.00	20.0	0.06	0.10
MW-9	18.9	496	0.08	0.27
MW-9 Dup	20.0	498		
MW-10	23.4	296	0.07	0.49
Blank	<.5	<.5	<.05	<.05
WP032	106	75.1	2.51	2.08
WP032 T.V.	106	75.0	2.81	2.30
Spike Rec.	101%	95%	98%	100%

•

<u>Sample</u>	рН	Conductivity	Redox
ESMP-5S	7.30	730	200
	7.07	751	-45.8
	7.01	823	-30.0
	7.22	667	152
ESMP-8D	7.10	779	-93.8
	7.08	781	-89.5
ESMP-11D	7.09	786	-70.1
ESMP-7S	7.33	632	199
ESMP-7D	7.29	703	-53.5
	7.18	840	-24.4
	7.11	803	2.70
ESMP-9D Dup	7.10	817	
	7.13	799	115
MW-12	7.04	854	38.6
ESMP-3D	7.06	810	-16.7
	6.96	1017	181
	6.98	1057	178
MW-11	7.38	566	194
ESMP-4D	7.02	874	140
ESMP-4S	6.98	965	23.1
ESMP-14D	7.09	767	-116
ESMP-14S	7.28	760	-115
ESMP-2D	7.16	761	-63.3
ESMP-1D	7.22	703	190
MW-3	7.08	943	212
MW-3 Dup	7.10	961	213
MW-8	7.34	719	209
ESMP-13D	7.14	775	-136
ESMP-13S	7.21	841	-136
ESMP-17S	7.24	773	-125
ESMP-15S	7.55	731	-95.0
ESMP-15D	8.22	764	72.1
MW-2	7.16	832	212
MW-10 .	7.11	1172	-92.1
MW-5	7.07	942	-115
ESMP-16S	7.00	2150	-143
ESMP-16D	6.94	2070	-170
ESMP-16D Field D.		2080	-172
MW-9	6.82	1596	19.1
MW-4	6.95	859	210

GROUNDWATER SAMPLES FROM RICKENBACHER AFB FOR TOC & TIC (SR# SF-1-118)

SAMPLES	MG/L TC	MG/L OC	MG/L TIC		MG/L TC	MG/L OC	MG/L TIC	!
MW-2	112.0	5.3	106.7	ESMP-15D	132.8	114.2	18	18.6
MM-03	105.6	4.6	101.0	ESMP-15D DUP	132.8	115.0	17	8.7
MW-4	112.4	7.2	105.2	ESMP-158	134.0	97.0	37	0.7
MW-5	197.2	139.6	57.6	ESMP-16D	156.4	61.3	95	5.1
MW-6	1	5.5		ESMP-16S	594.0	523.0	71	0.1
MW-6 DUP		5.6	1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1	ESMP-17S	104.0	2.0	102	5.0
MW-9	150.0	7.6	142.4					
MW-9 DUP	149.6	7.6	-	WPO32-II	10.2 MG/L			
MW-8	105.2	13.4	0		10.3 MG/L			
MW-10	100.4	5.7			10.9 MG/L			
MW-11	56.0	1.0			9.5 MG/L	•		
MW-12	8.86	3.5			9.8 MG/L			
MW-12 DUP	8.86	3.6			1			
ESMP-1D	92.0	3.3		WP033-I	7.7 MG/L			
ESMP-1D DUP	92.0	3.3			σ.			
ESMP-2D	94.8	2.6						
ESMP-3D	108.4	1.6						
ESMP-4D	109.6	2.6						
ESMP-4S	109.2	2.7	-					
ESMP-5D	0.86	2.0						
ESMP-5D DUP	98.4	2.0						
ESMP-5S	82.0	1.7						
ESMP-6D	112.0		-					
ESMP-7D	83.6							
ESMP-78	83.2						٠	
ESMP-8DD	99.2	1.3						
ESMP-8DD DUP	9.66		98.3					
ESMP-8S		5.1						
ESMP-8S DUP		5.3	1					
ESMP-9D	108.0		-					
ESMP-98 REP 1	92.6							
ESMP-9S REP 2	92.6							
ESMP-10D	122.4		118.5					
ESMP-10S	83.2							
ESMP-11D	100.4		97.8					
ESMP-11D DUP	100.8							
ESMP-13D	97.2		95.1					
ESMP-13D DUP	97.2							
ESMP-13S	129.6	27	-					
ESMP-14D	0.96							
ESMP-14D REP 3	0.96		92.4					
ESMP-14S REP 1	107.6	e	104.5					
ESMP-14S REP 1 DUP	1	3.1	104.5					
ESMP-14S REP 3	107.6	3.7	103.9					

TRUE VALUE: WPO32-II = 9.9 MG/L WPO33-I = 7.7 MG/L

A 1.2.3.4-TETBA		3.94 34.07 NO
1235-TETBA	5.00 5.00	8.84 32.54 NO
1.2.4.5-TETBA	8.8.8.8.8.8.8.8.8.8.8.8.8.8.8.8.8.8.8.	4.84 22.71 ND
12.3-TMB	25.28 8.28	17.08 57.44 ND
12.4-TMB	51.72 80.17 10.03 81.03 80.03	31.88 62.76 ND
1.3.5-TMB	55.04 81.03 81	8.42 NO NO NO
P-XYLENE	75.48 81.49 81	31.63 91.43 NO
M-XYLENE	65.69 50.06 50.06 81.09 81	7.80 18.02 ND
P-XYLENE	101.75 50.02 50.03 81.03	22.60 44.20 ND
ETHYLBENZENE	101.85 49.45 49.45 80.06 81.00 81.00 80.00 8	26.87 228.50 ND
TOLUENE	101.33 52.37 55.08 81.09 81.09 81.09 81.00	B B C O
BENZENE	101.52 53.37 50.09 81.00 81.00 81.61 11.61	6.46 89.18 BLO
SampleName	00C, OBSERVED, PPB 00C, TRUE VALUE, PPB MW-2 MW-3 MW-4 MW-4 MW-4 MW-6 MW-9 MW-10 MW-10 MW-11 10 PPB MW-10 MW-11 10 PPB MW-10 MW-11 10 PPB MW-10 ESMP-3D ESMP-3D ESMP-3D ESMP-3D ESMP-3D ESMP-4D ESMP-4D ESMP-4D ESMP-4D ESMP-4D ESMP-4D ESMP-8D ESMP-1D ESMP-11D	ESMP-16D ESMP-16S ESMP-17S

Table 1. Quantitation Report for S.R. # SF-1-118 from Rickenbacker.

Concentration = ppb

Compound	MW-2	MW-03	MW-4	MW-6	MW-6	MW-6 Fleid Dup	MW-8	WW-9	MW-10	MW-11	MW-11 Lab Dup
VINYL CHLORIDE 1,1-DICHLOROETHENE T-1,2-DICHLOROETHENE C-1,2-DICHLOROETHENE 1,1,1-TRICHLOROETHANE CARBON TETRACHLORIDE BENZENE 1,2-DICHLOROETHANE TRICHLOROETHENE TOLUENE TETRACHLOROETHENE TOLUENE TETRACHLOROETHENE TOLUENE TETRACHLOROETHENE TOLUENE TETRACHLOROETHENE TOLUENE	2222222222222	555 555555555	222222222222	ND ND ND ND 30.7 30.8 30.6 30.7 30.7 30.8 30.8 30.8 30.8 30.8 30.8 30.8 30.8	23.1 6.5 6.5 873 873 873 873 873 873 873 873 873 873	21.9 21.9 6.55 43.2 2.2 2.2 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0	99999999 999999	99999999999999	22222222222222	22222222222222	9999999999999
	MW-12	MW-12	ESMP 1D	ESMP 1S	ESMP 2D	ESMP 2S	ESMP 03	ESMP 3D	ESMP 4D	ESMP 4S	ESMP 5D
VINYL CHLORIDE 1,1-DICHLOROETHENE T-1,2-DICHLOROETHENE C-1,2-DICHLOROETHENE 1,1,1-TRICHLOROETHANE CARBON TETRACHLORIDE BENZENE 1,2-DICHLOROETHANE TRICHLOROETHENE TOLUENE TETRACHLOROETHENE TOLUENE TETRACHLOROETHENE TOLUENE T	99999999999999999999999999999999999999	ND N	N N N N N N N N N N N N N N N N N N N	AN SAN SAN SAN SAN SAN SAN SAN SAN SAN S		ND N	Z Z Z Z Z Z Z Z Z Z Z Z Z Z Z Z Z Z Z	16.0 ND ND ND ND ND ND ND ND ND N	16.0 ND ND N	0.1.0	2222222222222222
	ì ì								;		

Table 2. Quantitation Report for S.R. # SF-1-118 from Rickenbacker.

Concentration = ppb

ESMP 10S	5 5 5 5 5 5 5 5 5 5 5 5 5 5 5 5 5 5 5	ESMP 14S NO
ESMP 10D	22222222222222	A
ESMP 9S	22222221122222	AND ON
ESMP 9D	999999999999999	FSMP 14D Field Dup 1.0 ND
ESMP 8S	999999999999999	13S ESMP 14D 1.0 ND ND 2.2 ND
ESMP 8DD	222222221122222	7. 2. 2. 2. 2. 2. 3. 3. 4. 4. 4. 4. 4. 4. 4. 4. 4. 4. 4. 4. 4.
ESMP 7S	222222222222222	ESMP 13D ESMP 13D EST ND
ESMP 7D	22222222212222	ESMP 13D ND
ESMP 6D		220 NO
ESMP 6D	<u></u>	OS ESMP 11D ES ND
ESMP 5S	222222222222222	Lab Dup ND ND ND ND ND ND ND ND ND ND ND ND ND
	VINYL CHLORIDE 1,1 DICHLOROETHENE T-1,2 DICHLOROETHENE C-1,2 DICHLOROETHENE CHLOROFORM 1,1,1 TRICHLOROETHANE CARBON TETRACHLORIDE BENZENE TRICHLOROETHANE TOLUENE TOLUENE TETRACHLOROETHENE TOLUENE TOLUEN	VINNL CHLORIDE 1,1-DICHLOROETHENE T-1,2-DICHLOROETHENE C-1,2-DICHLOROETHENE CHLOROFORM 1,1,1-TRICHLOROETHANE CARBON TETRACHLORIDE BENZENE 1,2-DICHLOROETHANE TRICHLOROETHANE TRICHLOROETHENE TOLUENE TETRACHLOROETHENE CHLOROBENZENE ETHYLBENZENE ETHYLBENZENE O-XYLENE

Table 3. Quantitation Report for S.R. # SF-1-118 from Rickenbacker.

Concentration = ppb

QC0307F 200 ppb	207 203 191 195 200 198 219 219 197 197 197 202	
QC0307E 20 ppb	21.3 20.8 20.1 19.9 20.0 22.2 23.8 22.1 20.6 19.8 20.6 20.6 20.6 20.6 20.6 20.8	
QC0307D 200 ppb	Cracked Vial sample lost	(qda
QC0307C 20 ppb	20.3 20.4 19.4 20.5 19.7 19.6 22.0 22.6 21.1 19.6 22.6 18.9 39.6 *	
QC0307B 200 ppb	222 226 202 201 219 228 216 229 210 221 420 ***	 helow Callbrat
QC0307A 20 ppb	22.3 23.0 20.2 19.0 22.0 22.7 21.8 19.6 22.8 20.9 21.2 21.2 21.2 21.2 20.3	200 ppt 200 207 189 195 208 208 209 218 226 201 191 212 203 191 212 201 201 201 201 201 201 201 201 20
ESMP 17S	1570 11.7 152 4913 ND ND N	20 ppb 22.0 21.8 19.9 19.7 21.4 21.1 22.6 21.9 22.5 20.7 20.0 39.8 *
ESMP 16S	ND ND ND ND ND ND ND ND ND ND ND ND ND N	Ppb 200 ppb 196 196 199 203 199 203 199 2003 2002 2003 2005 2024 2025 203 2006 20.1 199 20.1 198 20.1 198 20.1 198 20.1 198 20.1 198 20.1 198 20.1 198 20.1 198 20.1 198 20.1 198 20.1 198 20.1 198 20.1 198 20.1 196 20.1 196 20.1 196 20.1 196 20.1 196 20.1 196 20.1 196 20.1 196 20.1 196 20.1 196 20.1 196 20.1 196 20.1 19.0 20.1 20.1 20.1 20.1 20.1 20.1 20.1 20
ESMP 16D	ND ND ND 6.2 ND ND ND ND ND ND ND ND ND ND ND ND ND	20 ppb 21.9 19.9 19.3 19.3 20.3 20.5 22.5 22.5 22.5 22.5 21.3 20.6 20.1 40.3 * 19.7
ESMP 16S	2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2	200 ppb 201 201 208 204 208 208 209 208 209
ESMP 16D	8 8 8 8 8 8 8 8 8 8 8 8 8 8 8 8 8 8 8	8 88558888885858
Compound	VINAL CHLORIDE 1,1—DICHLOROETHENE 1,2—DICHLOROETHENE C—1,2—DICHLOROETHENE 1,1,1—TRICHLOROETHANE CARBON TETRACHLORIDE BENZENE 1,2—DICHLOROETHANE TRICHLOROETHANE TRICHLOROETHENE TOLUENE	VINAL CHLORIDE 1,1-DICHLOROETHENE T-1,2-DICHLOROETHENE C-1,2-DICHLOROETHENE 1,1,1-TRICHLOROETHANE CARBON TETRACHLORIDE BENZENE 1,2-DICHLOROETHANE TRICHLOROETHANE TRICHLOROETHENE TOLUENE TETRACHLOROETHENE TOLUENE TETRACHLOROETHENE TOLUENE TETRACHLOROETHENE TOLUENE TETRACHLOROETHENE TOLUENE TETRACHLOROETHENE OHLOROBENZENE ETHALBENZENE ETHALBENZENE M+p-XYLENE 0-XYLENE O-XYLENE

units = mg/L

Lower quantitation limits: methane = 0.001 ms/2

ethylene = 0.003 mg/L
per Do-kampbell phone conv.
4/12/95

ANALYZED 3/7/95 SAMPLE	METHANE	ETHYLENE
LAB BLANK ESMP-3D ESMP-5D	ND 0.067 0.106	ND ND ND
ESMP-5S ESMP-6D	0.002 0.079	ND ND
ESMP-7D ESMP-7S	0.478 0.017	ND ND
ESMP-8DD ESMP-8S	0.006 0.003	ND ND
ESMP-9D	0.008	ND
ESMP-9S " FIELD DUP	0.015 0.016	ND ND
ANALYZED 3/8/95		
SAMPLE	METHANE	ETHYLENE
LAB BLANK ESMP-10D	BLQ 0.012	ND ND
ESMP-10S	0.012	ND
ESMP-11D	0.105	ND
MW6	0.013	ND
MW11	BLQ	ND
MW12	0.001 0.058	ND ND
ESMP-1D ESMP-2D	0.058	ND
ESMP-4D	0.007	$0.001 = \angle 0.003$
ESMP-4S	0.109	ND
ESMP-13D	0.110	ND
" FIELD DUP	0.114	ND
ESMP-13S	7.830	0.001 = < 0.003
ESMP-14D	0.106	ND
ESMP-14S	0.462	ND ND
ESMP-15D ESMP-15S	0.007 0.136	ND
"LAB DUP	0.130	ND
ESMP-16D	1.150	ND
* FIELD DUP	1.182	ND
ESMP-16S	3.067	ND
ESMP-17S	2.296	0.057
MW2	0.661	ND
MW3	0.003	ND ND
MW4 MW5	0.002 7.693	ND ND
"LAB DUP	7.693 7.178	ND
MW8	0.015	ND
MW9	0.004	ND
MW10	0.040	ND



Ref: 95-DF20

April 17, 1995

Dr. Don Kampbell

R.S. Kerr Environmental Research Lab

U.S. Environmental Protection Agency

P.O. Box 1198 Ada, OK 74820

THRU: S.A. Vandegrift 9

Dear Don:

As requested in Service Request SF-1-118, GC/MS analysis for phenols and aliphatic/aromatic acids was done on one water sample labelled MW-5 from Rickenbacken ANGB. Liquid-liquid extraction was done by Mark Blankenship on April 6, 1995. The extract was analyzed by GC/MS on April 13, 1995. A SOP describing the extraction, derivatization and GC/MS analysis is in preparation.

Liquid-Liquid Extraction of Phenols and Aliphatic/Aromatic Acids.

For the extraction of the phenols and aliphatic/aromatic acids from the water sample, 100 ml of the water sample is placed in a dried, silanized 125 ml separatory funnel. Spike solutions if applicable were added to the sample at this time. The pH of the water is adjusted to 2.0 using 1:1 H₂SO₄. For a water blank without Na₃PO₄ added, a pH of 2 is reached with ten drops. For 100 ml of water sample preserved with Na₃PO₄, twenty drops of acid is required. Next 25 g of NaCl is added to the separatory funnel after which the liquid is swirled to dissolve the salt.

The water sample is extracted four times with 5 ml aliquots of acid free methylene chloride. To remove acids from methylene chloride and other solvents, 10 g of Celite Micro-Cel T-49 is added to one liter of GC/MS grade solvent. This mixture is stirred for one hour, allowed to settle and is filtered through a Millipore organic filter pad using Millipore vacuum apparatus. The methylene chloride extracts are collected in silanized 40 ml VOA vials. The total extract volume is recorded.

Phenol/Acid Derivatization to Form PFB Ethers and Esters.

A 200 μ l aliquot of the methylene chloride extract is delivered to a 2 ml screw cap vial containing 2.5 mg of dried potassium carbonate. Next 790 μ l of acid free acetonitrile, 10 μ l of 100 ppm benzoic acid-d, and 10 μ l of pentafluorobenzyl bromide is added to the vial. Benzoic acid-d, is the internal standard for the analysis. The vials are momentarily placed in a sonic bath to free the solid salt from the bottom of the vial. The screw caps of the vials are tightened and the vials are heated in a oven at 60°C for 2 hours. When the vials are removed from the oven, 500 μ l of 0.1M Hcl is added. The vials are shaken for 30 seconds and 200 μ l of the top organic layer is delivered to the liner of a 2 ml crimp cap autosampler vial.

Negative Ion Chemical Ionization GC/MS Analysis of PFB-Derivatives.

For negative ion chemical ionization GC/MS, a chemical ionization ion volume is placed in the ion source block of the Finnigan 4615 GC/MS. Methane gas is regulated using a needle valve until the ionizer pressure reaches 0.40 torr. With the ionizer at this pressure, the high vacuum pressure indicates 1.0x10⁻⁵ torr. The mass spectrometer is tuned using the calibration gas, FC-43, to obtain good peak shape for ions 414 and 633 m/z and a relative intensity of 100:14:4 for ions 633, 414 and 127 m/z. The ion source is heated at 150°C. The injector and transfer lines are held at 275°C.

The Hewlett Packard 7673 autoinjector delivered 0.5 μ l of the sample or standard to the GC injection port. A splitless injection for 1 minute was used for the analysis. The analytical column was a 60 meter, 0.25 mm J&W DB5-MS capillary column with 0.25 μ m film thickness. The column was temperature programmed from 50°C to 100°C at 30°C/min and then to 300°C at 6°C/min. The helium linear velocity measured with air was 36 cm/s when the oven temperature was 100°C and the helium head pressure on the column was 29 psi. The Finnigan 4615 GC/MS was scanned from 42 to 550 m/z in 0.5 sec.

Standard curves are prepared using a mixture containing thirteen phenols, twenty-five aliphatic acids and nineteen aromatic acids. Calibration curves for acetic acid was not prepared due to artefact levels of this acid in solvents. Derivatization of the standard solutions and samples was done in the same manner. Standards are prepared at 5, 10, 25, 50, 100, 500 and 1000 ppb. Quality assurance was maintained during the sample analysis by running check standards, derivatization blanks, extraction banks, extraction recovery check standards and spiked field samples.

Quantitative Results of Phenols and Aliphatic/Aromatic Acids.

Table I provides the concentrations of phenols and aliphatic/aromatic acids found in the water sample taken at the Rickenbacken ANGB site and quality assurance samples run at the same time as the samples. The lowest reported value of phenol or acid in this table is at or about 5 ppb.

Spike recoveries for each of the acids and phenols were determined in 50 ppb spikes of 100 ml of water blank. Recovery of the 50 ppb concentration was poor for low molecular weight aliphatic acids due to the poor extraction efficiencies of these acids from water. Higher molecular weight aliphatic acids and all the phenols and aromatic acids exhibit good recoveries.

Sample MW5 contained high levels of lower molecular weight acids and branched heptanoic and octanoic acids. The branched octanoic acids are present in the water sample at or above 1 ppm. These compounds are labelled in the attached chromatograms. One chromatogram shows the extracted ion profile of C_6 , C_7 , C_8 and C_9 branched chain acids. The extracted ion for each corresponds to the carboxylate ion of each acid. Also included please find extracted ion profiles for m/z ions 141 and 155. The 141 ion could correspond to octenoic acids, methylcyclohexanecarboxylic acids or dimethylcyclopentanecarboxylic acids. Compounds such as nonenoic acids or dimethyl-(or ethyl)cyclohexanecarboxylic acids could give carboxylic ions at 155 m/z.

Please note that a problem has occurred in the determination of benzoic acid. A larger amount of benzoic acid was found in the extraction blank than in the sample. We will determine the source of the benzoic acid artifact before the next acid/phenol sample queue is started.

If you should have any questions, please feel free to contact me.

Sincerely

Dennis D. Fine

xc: J.L. Seeley

G.B. Smith

R.L. Cosby

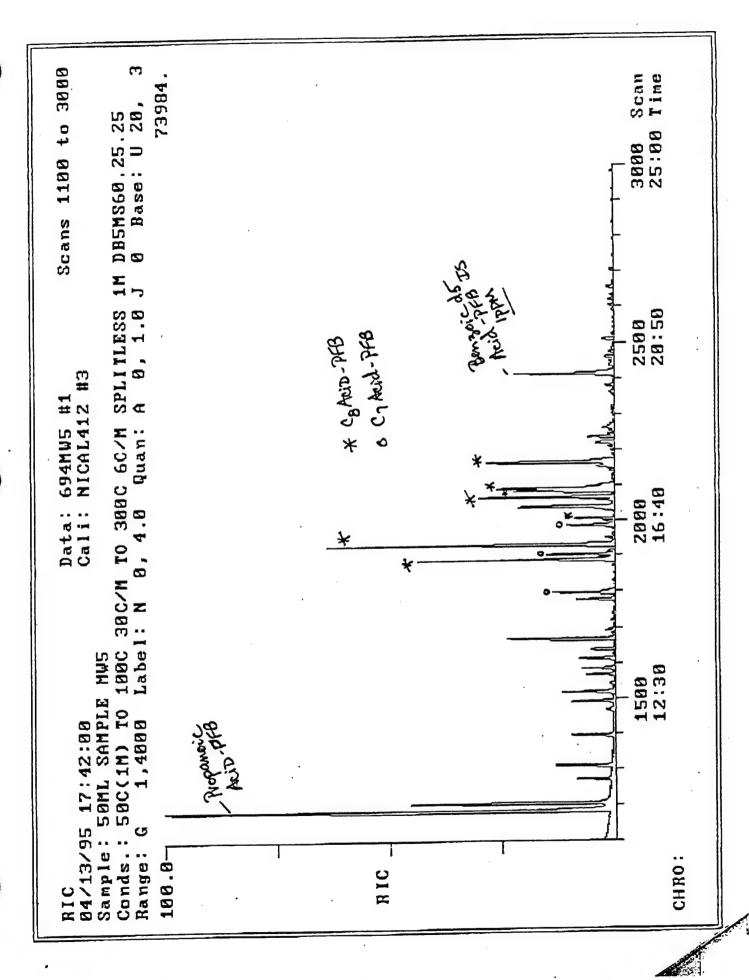
APR-24-19

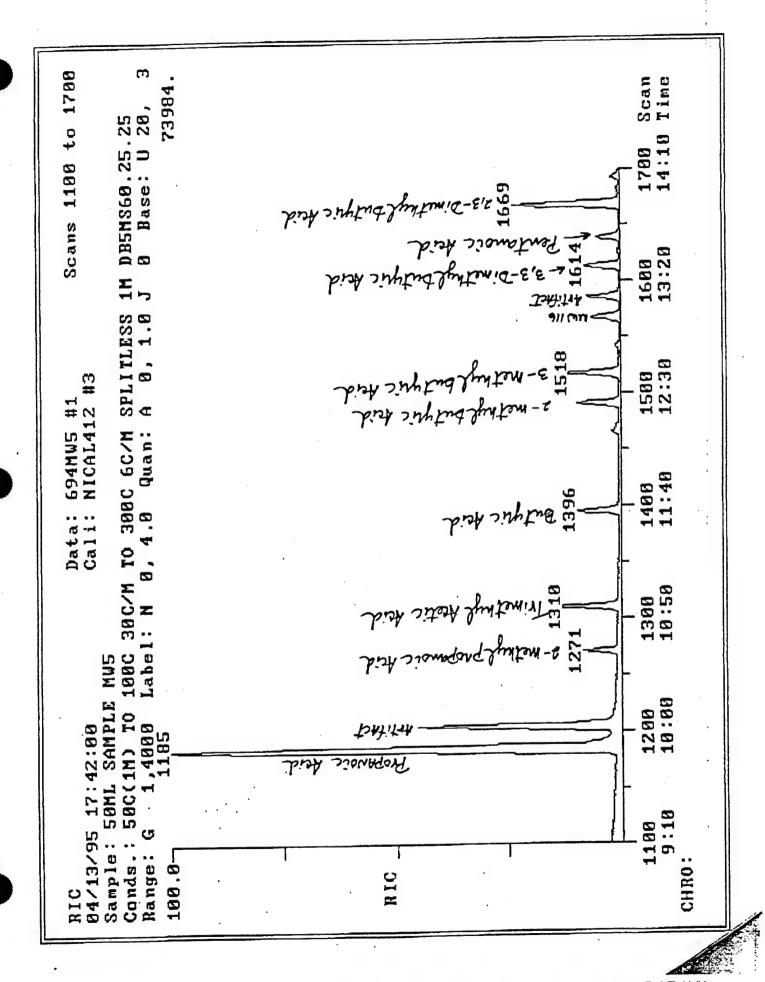
Table I. Quantitative Report and QC Data for Phenois and Aliphatic and Aromatic Adds for Samples from Fickenbacken ANGS (Service Request SF-1-118).

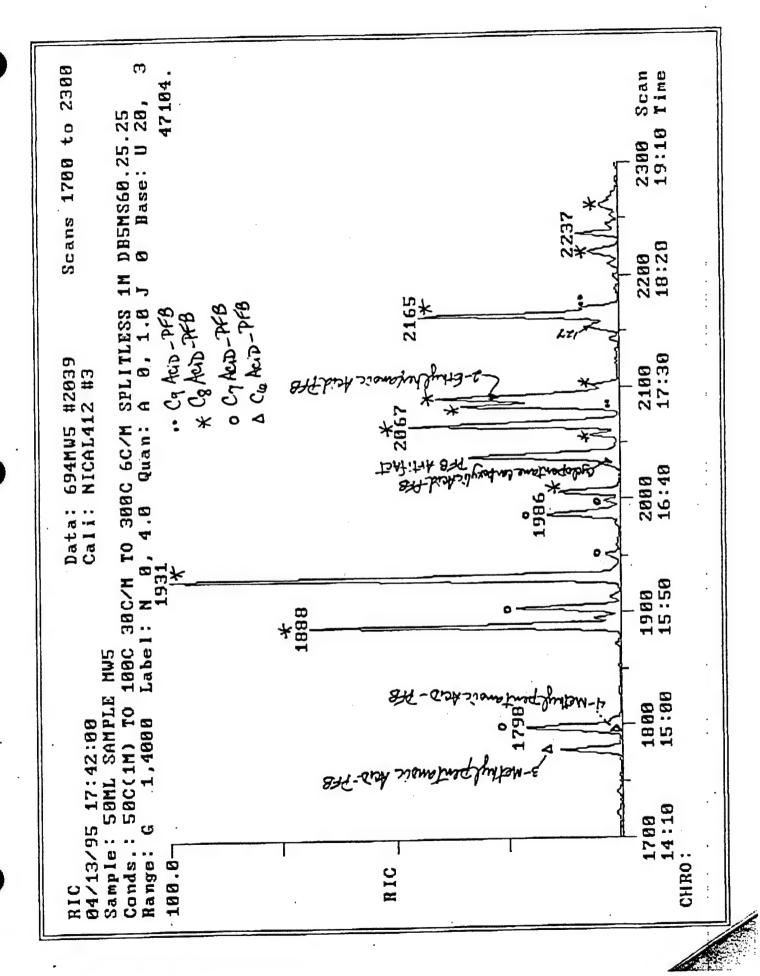
		Extraction	60 ppb Extraction		
		BENK	Recovery	MW45	
1	PROPANOIC ACID - PFB	22	44	1594	
2	Z-METHYLPROPANOIC ACID - PFB	5	14	76	
	TRIMETHYL ACETIC ACID - PFB	5	52	186	
3	BUTYRIC ACID - PFB	11	11	121	
•	2-METHYLBUTYRC ACID - PFB	***	41	187	
5	3-METHYLBUTYRC ACID - PFB	. ***	38	267	
6		N.F.	58	144	
7	3,3-DIMETHY BUTYRIC ACID - PFB	7	41	110	
8	PENTANOIC ACID - PFB	N.F.	58	688	
8	2,3-DIMETHYLBUTYRIC ACID - PFB	N.F.	57	20	
10	2-ETHYLBUTYRIC ACID - PFB		59	44	
13	2-METHYLPENTANOIC ACID - PFB	N.F.	58	248	
12	3-METHYLPENTANOIC ACID - PFB	N.F.		70	
13	4-METHYLPENTANOIC ACID - PFB	N.F.	67		
14	HEXANOIC ACID - PFB	20	61	33	
15	2-METHYLHEXANOIC ACID - PFB	N.F.	63_	14	
16	PHENOL - PFB	-	52		
17	CYCLOPENTANECARBOXYLIC ACID - PFB	N.F.	45	25	
18	S-METHYLHEXANOIC ACID - PFB	N.F.	61	В	
	a-CRESCL - PFB	N.F.	63	N.F.	
19	2-ETHYLHEXANOIC ACID - PFB		64	1938	
20	HEPTANOIC ACID - PFB	8	64	***	
21		N.F.	, 61	N.F.	
22	m-CRESOL - PFB	N.F.	60	***	
23	p-CRESOL - PFB	N.F.	43	N.F.	
24	1-CYCLOPENTENE-1-CARBOXYLIC ACID - PFB	N.F.	63	N.F.	
25	O-ETHYLPHENOL - PFB	N.F.	59	20	
26	CYCLOPENTANEACETIC ACID - PFB	N.F.	50	N.F.	
27	2,6-DIMETHMUPHENOL - PFB	N.F.	58	N.F.	
26	2,5-DIMETHYLPHENOL - PFB			14	
28	CYCLOHEXANECARBOXYUC ACID - PFB	N.F.	61	N.F.	•
30	3-CYCLOHEXENE-1-CARBOXYLIC ACID - PFB	***	55	N.F.	
31	2,4-DIMETHYLPHENOL - PFB	N.F.	46		
32	3,5-DIMETHYLPHENOL & M-ETHYLPHENOL - PFB	N.F.	131	N.F.	
33	OCTANOIC ACID - PFB	***	54	5	
34	2.3-DIMETHYLPHENOL - PFB	N.F.	62	N.F.	
32	P-ETHYLPHENOL - PFB	N.F.	67	N.F.	
36	BENZOIC ACID - PFB	88	71	19	
	3,4-DIMETHYLPHENOL - PFB	N.F.	63	N.F.	
37	m-METHYLBENZOIC ACID - PF8	N.F.	45	•••	
38	1-CYCLOHEXENE-1-CARBOXYUC ACID - PFB	N.F.	59	N.F.	
39		N.F.	67	168	
40	CYCLOHEXANEACETIC ACID - PFB	N.F.		904	
41	2-PHENYUPROPANOIC ACID - PFB	N.F			
42		***		6	
43		N.F.	-	•	•
44		N.F		***	
45		N.F		•••	
46		N.F	•	**	
47			•	7 .	
46		N.F			
44	3-PHENYLPROPANOIC ACID - PFB	N.F			
50	THE PERSON AND THE PE	N.F		***	
51	050	441		***	
5:	THE PENTON ACID DED	N.F			
5	ACUZOC LCID - PER	N.S		•••	
5	THE PERSON ACIO - PER	N.F	. 61	***	
	TOTAL SOLD DED	N.S	. 65		
5	THE PARTY OF THE P	N.F	F. 89	18	
3		N.F	F. 60		
5	THE PARTY OF THE P	N.I	F. 81	***	
5	8 2,4,5 -TRMETHYLEENZOIC ACID -FFS				

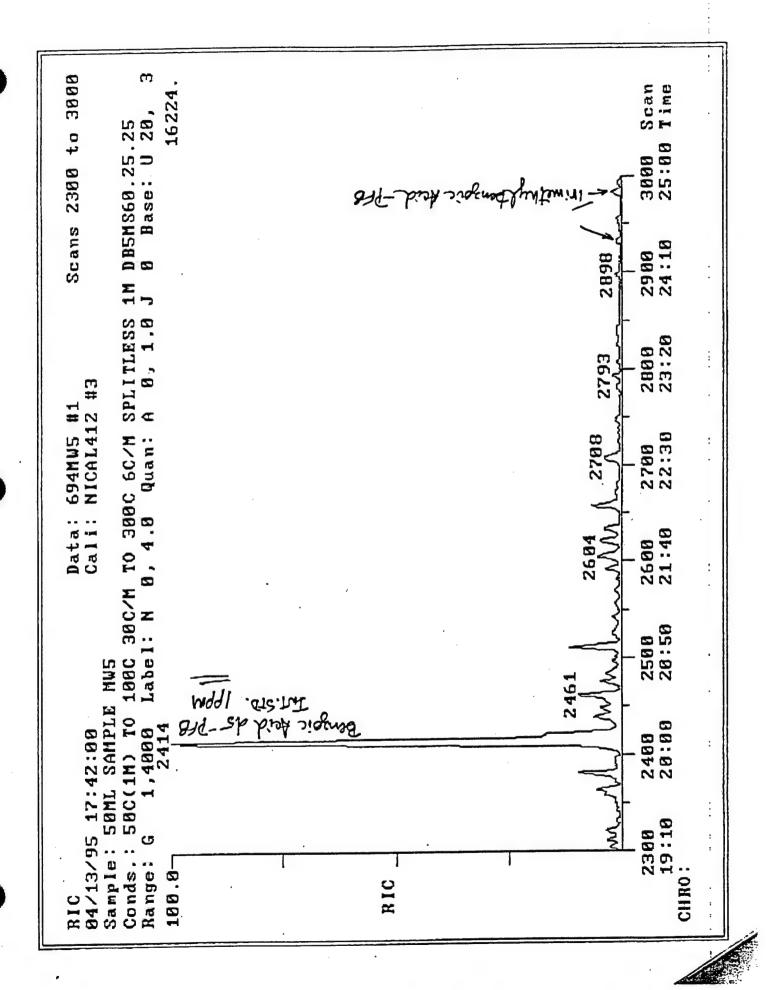
^{***} Indicates concentration of extract was below lowest calibration standard (5 ppts).

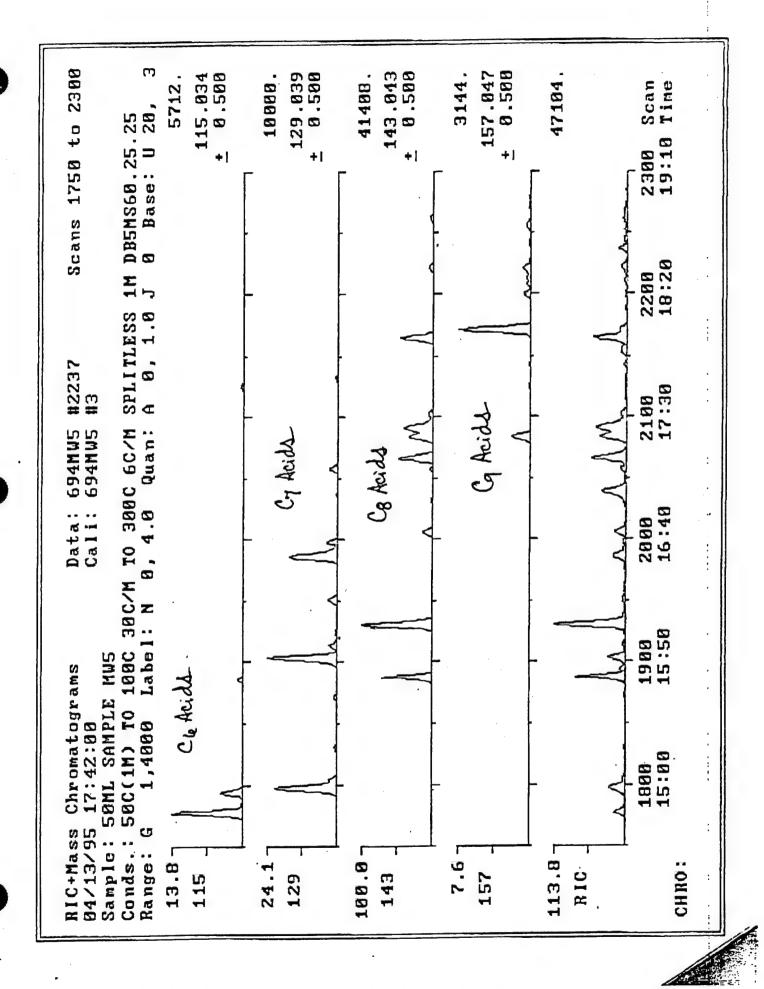
N.F. Indicates not found.

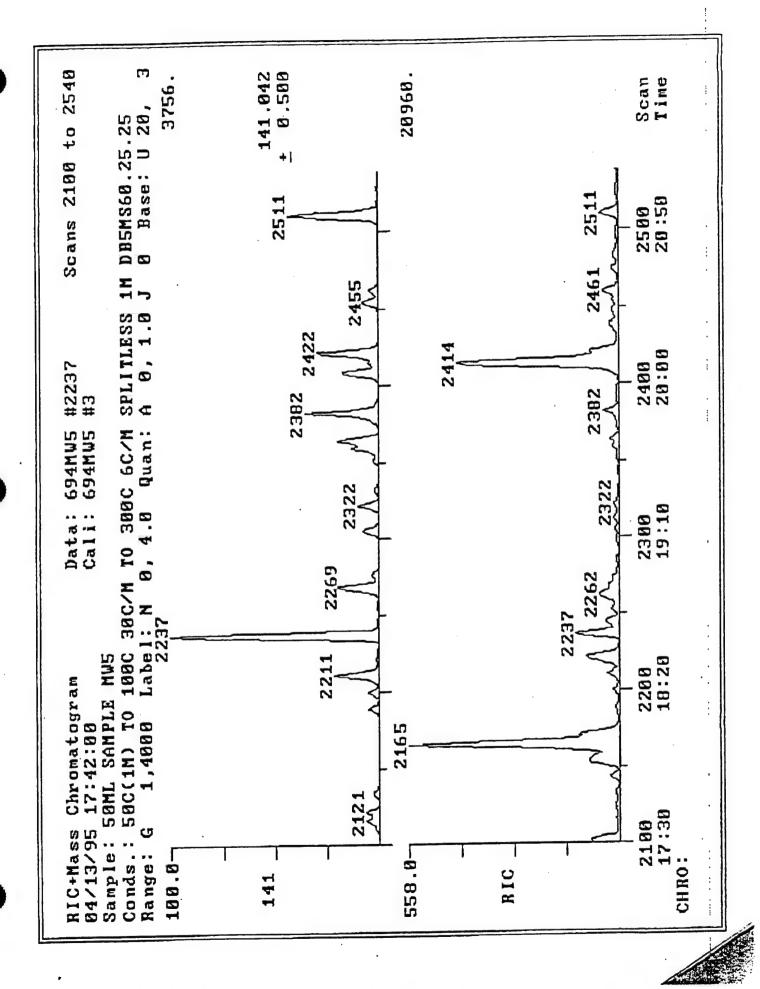












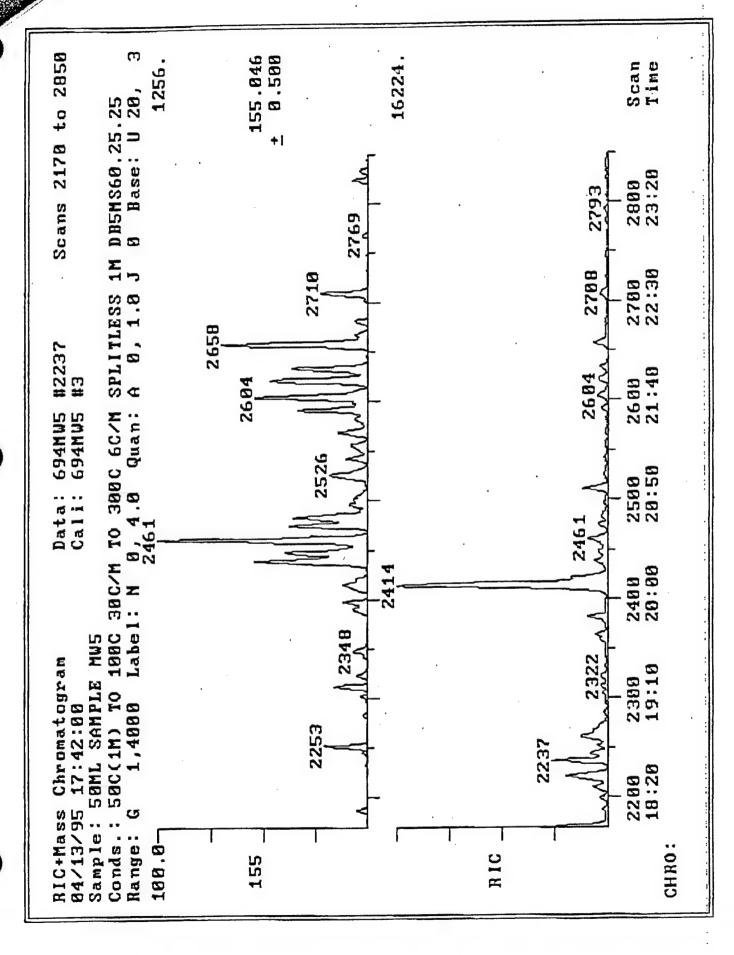


Table I. Quantitative Report and QC Data for PhenoIs and Aliphatic and Aromatic Acids for Samples from Rickenbacker AFB (Service Request SF-1-118).

			С	oncentration	ppb				
		ESMP-13S	ESMP-13D	10 ppb Check Standard	100 ppb Check Standard	Standard Derivative Blank	Standard Derivative Blank	Extraction Method Blank	% Extraction Recovery of 50 ppb Spike
1	PROPANOIC ACID - PFB	840	19	17	116	11	16	14	11
2	2-METHYLPROPANOIC ACID - PFB	154	8	14	116	***	6	6	43
3	TRIMETHYL ACETIC ACID - PFB	•••	12	12	122	***		В	113
4	BUTYRIC ACID - PFB	144	7	15	114	8	1.1	9	33
5	2-METHYLBUTYRIC ACID - PFB	92	***	10	116	***	***		91
6	3-METHYLBUTYRIC ACID - PFB	112	***	10	116	***	***		119
7	3,3-DIMETHYLBUTYRIC ACID - PFB	***	N.F.	11	118	N.F.	N.F.	N.F 7	99
8	PENTANOIC ACID - PFB	43	•••	13	113		N.F.	N.F	118
9	2,3-DIMETHYLBUTYRIC ACID - PFB	5		11	119	N.F.	N.F.	N.F	120
10	2-ETHYLBUTYRIC ACID - PFB	N.F.	N.F.	8	117	N.F. N.F.	N.F.	N.F	123
11	2-METHYLPENTANOIC ACID - PFB	9		8	115	N.F.	N.F.	N.F	120
12	3-METHYLPENTANOIC ACID - PFB	5			112	N.F.	N.F.	N.F	121
13	4-METHYLPENTANOIC ACID - PFB		N.F.	8	112	5	5	***	136
14	HEXANOIC ACID - PFB	29	7	12 8	115 113	N.F.	N.F.	N.F	130
15	2-METHYLHEXANOIC ACID - PFB		N.F.	12	99	14.7	***	***	113
16	PHENOL - PFB	5		7	107	N.F.	N.F.	N.F	106
17	CYCLOPENTANECARBOXYUC ACID - PFB	***	***	11	110	•••	***		144
18	5-METHYLHEXANOIC ACID - PFB	N.F.	N.F.	11	127	N.F.	N.F.	N.F	124
19	o-CRESOL - PFB 2-ETHYLHEXANOIC ACID - PFB	285		12	123	***	***	***	126
20	HEPTANOIC ACID - PFB		***	/ 10	116	***	***	***	131
21	m-CRESOL - PFB	N.F.	N.F.	11	126	N.F.	N.F.	N.F	123
22 23	p-CRESOL - PFB	N.F		11	127	N.F.	N.F.	N.F	123
24	1-CYCLOPENTENE-1-CARBOXYLIC ACID - PFB	N.F.	N.F.	11	113	N.F.	N.F.	N.F	96
25	e-ETHYLPHENOL - PFB	N.F	N.F.	10	126	N.F.	N.F.	N.F	127
26	CYCLOPENTANEACETIC ACID - PFB	N.F.	N.F.	7	113	N.F.	N.F.	N.F	120
27	2,6-DIMETHYLPHENOL - PFB	N.F.	N.F.	9	127	N.F.	N.F.	N.F	
28	2.5-DIMETHYLPHENOL - PFB	N.F	N.F.	10	130	N.F.	N.F.	N.F	
29	CYCLOHEXANECARBOXYLIC ACID - PFB	***	N.F.	7	111	N.F.	N.F.	N,F	
30	3-CYCLOHEXENE-1-CAPBOXYUC ACID - PFB	N.F	N.F.	11	120	•••	N.F.	••	117
31	2.4-DIMETHYLPHENOL - PFB	N.F	N.F.	10	139	N.F.	N.F.	N.F	1
32	3,5-DIMETHYLPHENOL & M-ETHYLPHENOL - PFB	N.F	. N.F.	10	132	N.F.	N.F.	N.F	i .
33	OCTANOIC ACID - PFB	***	***	14	118	***	***	**	127
34	2,3-DIMETHYLPHENOL - PFB	N.F	. N.F.	10	127	N.F.	N.F.	N.F	1
35	P-ETHYLPHENOL - PFB	N.F		10	130	N.F.	N.F.	N.5	
36	BENZOIC ACID - PFB	36		95	161		23 N.F.	N.F	
37	3,4-DIMETHYLPHENOL - PFB	N.F		10	135	N.F.	N.F.	N.I	1
38	m-METHYLBENZOIC ACID - PFB	20		11	110	N.F.	N.F.	N.I	
39	1-CYCLOHEXENE-1-CARBOXYUC ACID - PFB	N.F		11	109	N.F.			
40	CYCLOHEXANEACETIC ACID - PFB	N.F		7 10	111	N.F.			1
41	2-PHENYLPROPANOIC ACID - PFB	N.F		10	118	N.F.	N.F.		1
42	o-METHYLBENZOIC ACID - PFB			10	112	***	***		
43	PHENYLACETIC ACID - PFB	N.F		10	100	N.F.	N.F.	N.I	120
44	m-TOLYLACETIC ACID - PFB	N.F		14	92	N.F.		N.	143
45	O-TOLYLACETIC ACID - PFB	N.F		12		N.F.	N.F.	N.	136
46	2,6-DIMETHYLBENZOIC ACID - PFB p-TOLYLACETIC ACID - PFB	N.F		9	104	N.F.	N.F.	N.	F 155
47	P-METHYLBENZOIC ACID - PFB			10	111	N.F.	N.F.	N.	125
48	3-PHENYLPROPANOIC ACID - PFB	••		11	106	N.F.	N.F.	N.	F 127
49 60	2,5-DIMETHYLBENZOIC ACID - PFB	N.F		11	111	N.F.	N.F.		1
50	DECANOIC ACID - PFB	**		12	112	***	***		
51 52	2,4-DIMETHYLBENZOIC ACID - PFB	**	N.F.	10	116	N.F.	N.F.		1
52 53	3,5-DIMETHYLBENZOIC ACID - PF8	••	· N.F.	9	107	N.F.	N.F.		
53 54	2,3-DIMETHYLBENZOIC ACID - PFB	N.F	N.F.	to	114	N.F.	N.F		
	4-ETHYLBENZOIC ACID - PFB	N.F	N.F.	10	113	N.F.	N.F.		
56	2,4,6—TRIMETHYLBENZOIC ACID — PFB	**	N.F.	10	121	N.F.	N.F	, N.	
57	3,4-DIMETHYLBENZOIC ACID - PFB	••	N.F.	10	105	N.F.	N.F	. ' N.	
58	200	N.S	. N.F.	9	115	N.F.	N.F	, N.	.F 131
30	Piala Humann management								

APPENDIX B-4

INFORMAL TECHNICAL INFORMATION REPORT (IT, 1995B)



11499 Chester Road Cincinnati, Ohio 45246-4012 513-782-4700 Fax: 513-782-4807

18 September 1995

Mr. Alan Friedstrom, P.E. OLR, AFBCA 2146 Club Road, Building 548 Rickenbacker ANGB, Ohio 43217-5548

RE: AFCEE Contract No. 41624-94-D-8047 D.O. 0008, 409878.005.05.001 Sampling of Site 1, Hazardous Waste Storage Area

Dear Mr. Friedstrom:

Enclosed is a summary ITIR for the sampling recently completed at Building 560, RCRA Site 1, including:

- Text summary of field activities
- Tables presenting all field and analytical results.
- Discussion of the geoprobe sampling and analyses near well ESMP-17s.

We are also forwarding at this time, a copy of the data directly to Mr. Jeff Black of Parsons, Denver, for his incorporation into their October deliverable.

Please call either me or Karl Van Keuren if we can provide additional information.

Sincerely,

IT CORPORATION

Bernard J. Franks Project Manager

BJF:srm Enclosure

cc: Jeff Black, Parsons (with enclosure)
Dean Dunn, AFCEE/ERB (letter only)
Richard Haines, OLR, AFBCA (letter only)

Informal Technical Information Report

Groundwater Sampling
Site 1, Hazardous Waste Storage Area
Building 560

Rickenbacker ANGB, Ohio

RANGB Hazardous Waste Storage Area (560) August 1995 Sampling

Field activities conducted by IT personnel from 15 August to 22 August 1995 at the Hazardous Waste Storage Area, in support of Parsons post-closure deliverable due in October 1995, consisted of:

- measuring groundwater levels in 34 well points and monitoring wells
- measuring groundwater level and free product thickness in MW-5
- sampling four monitoring wells for first quarter RCRA samples
- sampling twelve half-inch diameter well points and three monitoring wells for natural attenuation parameters
- geoprobe sampling and analysis of the northern extent of contamination.

Groundwater elevations were measured to the nearest hundredth foot using a SINCO water level probe (Table 1). Free product thickness was also measured in MW-5 using an ORS oil-water interface probe.

All samples collected were analyzed in the field for pH, temperature, conductivity, oxidation reduction potential (ORP), and dissolved oxygen (Table 2). On 8/16/95, an attempt was made to analyze a sample using 9 Hach field kits for a variety of parameters. This process was determined to be ineffective and time consuming. After consulting with the BEC and reviewing notes from the initial USEPA well installation and sampling protocols, the number of field analyses was reduced to ferrous iron and sulfide. After the first day of sampling, it was decided to add alkalinity as a third parameter measured in the field. Results of the field analyses are presented in Table 3. Other natural attenuation parameters were analyzed in a laboratory. A list of all field and laboratory analyses is provided in Table 4. Most laboratory analyses were performed by PDP Analytical in Spring, Texas, with results presented on copies of Form 1 (Attachment A). Methane, ethane, and ethene, were analyzed in IT's Biotechnology Applications Center laboratory in Knoxville, Tennessee. Data are presented in a Certificate of Analysis at the front of Attachment A.

One field variance should be noted. The initial plan indicated sampling from 23 wells, all initially sampled during well installation in 1995. During the August resampling period, several of the shallow well points were unable to be purged cf adequate volumes of water and samples were not collected for field or laboratory analyses. After one hour of purging was attempted at these points, it was decided, in consultation with the BEC, to eliminate these wells from further sampling. These wells are: ESMP-2S, ESMP-3S, ESMP-6S, and ESMP-12S.

In addition to the groundwater monitoring well sampling conducted, four Geoprobe soil borings, 560-01, 560-02, 560-03, and 560-04 were advanced radially upgradient of monitoring point ESMP-17s (Figure 1), on August 15th and 16th. The purpose of these borings was to collect additional shallow groundwater data to delineate the extent of the vinyl chloride (VC), cis-dichloroethene (cis-DCE), and trans-dichloroethene (trans-DCE) detected in monitoring point ESMP-17s.

The Geoprobe drilling program was conducted by using a Geoprobe 8M van-mounted hydraulic sampler. Boreholes were advanced using 1.125-inch inside diameter (ID) (1.375 nominal outside diameter) lead samplers and drive-shoes and 1-inch drill rods through the application of downhole pressure. Each soil boring was continuously sampled by advancing the lead sampler and drive shoe to substrate 24 inches. Following advancement, the lead sampler was removed from the boring. Soil sampling for lithologic description and flame ionization detector (FID) screening was conducted by using a clear acetate liner in the lead sampler. Lithologic descriptions are presented in the attached Soil Boring Logs. No FID readings were recorded above background readings. Each soil borings was advanced to a minimum of two feet below first encountered groundwater. Well points were then installed to collect groundwater samples. Well points consisted of a 1.25-inch ID by 2-foot slotted rods.

Following installation, the well points were developed and purged using a disposable Teflon tube bailer and a peristaltic pump. Each well point was first surged with the tube bailer and then pumped with the peristaltic pump, until either turbidity cleared or the well point was pumped dry.

Each well point was sampled using pre-cleaned, disposable equipment. Groundwater samples were collected using the following sampling technique:

- A disposable Teflon tube bailer was used at each well point. The bailer was lowered and withdrawn slowly from the monitoring well to minimize agitation of the groundwater. The recovered groundwater was then transferred from the bailer to a clean sample bottle. The sample bottle was the labeled as to time and date collected, analyses requested, capped and immediately placed in an ice bath.
- Clean latex gloves were worn while sampling at each monitoring well.

Two equipment rinsate sample were collected from decontaminated slotted rods to evaluate the effectiveness of the decontamination procedures. The rinsate samples were collected by the following method:

- Deionize I water was allowed to pour down through the entire length of the slotted rod.
- The rinsate water was then collected in clean sample bottles and labeled as to time and date collected and analyses requested, capped, and immediately placed on ice.

The groundwater samples collected from the well points and the rinsate samples were analyzed for Volatile Organic Compounds (VOCs). A total of four groundwater samples and two rinsate samples were collected. Analytical results for the samples collected from well points 560-02, 560-03, 560-04 were all nondetect (ND) for VOCs. VOCs were detected in the sample collected from well point 560-01; VC at 26 micrograms per liter (ug/l), cis-DCE at 27 ug/l, and trans-DCE at a laboratory estimated 4 ug/l. Analytical results for the rinsate sample R081695 were nondetect for VOCs. One VOC was detected in rinsate sample R081795; methylene chloride at a laboratory estimated 2 ug/l. The data are presented on Figure 1.

Table 1
Groundwater Elevations 15 August 1995
Referenced from Top of Inside Casing
Site 1, Hazardous Waste Storage Area (560)
Rickenbacker ANGB
Page 1 of 2

Well ID	Time	Depth to Water (feet)
MW-2	1702	12.14
MW-3	1656	10.46
MW-4	1625	11.38
MW-5	1723	13.37 (13.31 to free product)
MW-6	1602	13.75
MW-8	1747	10.53
MW-9	1629	8.05
MW-10	1706	11.23
MW-11	1548	12.19
MW-12	1520	13.10
ESMP-1S	1646	7.43
ESMP-1D	1651	10.17
ESMP-2S	1726	3.57
ESMP-2D	1728	6.18
ESMP-3S	1609	7.68
ESMP-3D	1611	10.69
ESMP-4S	1553	5.94
ESMP-4D	1556	11.15
ESMP-5S	1358	3.83
ESMP-5D	1354	6.20
ESMP-6S	1517	10.79
ESMP-6D	1518	9.69

Table 1 Groundwater Elevations 15 August 1995 Referenced from Top of Inside Casing Site 1, Hazardous Waste Storage Area (560) Rickenbacker ANGB Page 2 of 2

Well ID	Time	Depth to Water (feet)
ESMP-7S	1513	6.02
ESMP-7D	1514	5.96
ESMP-8S	1503	3.76
ESMP-8D	1504	6.17
ESMP-8DD	1505	7.38
ESMP-9S	1456	3.38
ESMP-9D	1458	6.53
ESMP-10S	1450	6.15
ESMP-10D	1451	6.41
ESMP-11S	1508	6.66
ESMP-11D	1509	7.11
ESMP-12S	1621	2.57
ESMP-13S	1711	10.34
ESMP-13D	1717	10.06
ESMP-14S	1738	10.13
ESMP-14D	1736	9.89
ESMP-14DD	1739	9.72
ESMP-15S	1743	9.45
ESMP-15D	1744 -	8.98
ESMP-16S	1634	8.00
ESMP-16D	1635	8.77
ESMP-17S	1750	9.19

Table 2
Water Quality Data
Site 1, Hazardous Waste Storage Area (560)
Rickenbacker ANGB
Page 1 of 3

Well ID	Date	Time	Volume (gallons) Purged	Temperature (Celsius)	pH (S.U.)	Conductivity (millimhos/cm)	ORP (mV)	Dissolved Oxygen (mg/L)
MW-3	8/17/95	1616	0.0	19.0	6.55	0.843	152	1.4
		1618	1.0	18.5	6.61	0.825	147	1.0
		1620	2.0	18.2	6.67	0.830	159	1.6
		1623	3.0	17.0	6.70	0.776	157	1.8
MW-4	8/17/95	1026	0.0	20.6	6.61	0.680	196	1.7
		1028	1.0	19.5	6.57	0.808	207	1.1
		1030	2.0	19.4	6.60	0.827	204	1.2
		1031	2.5	19.2	6.62	0.781	202	1.6
		1033	3.5	18.8	6.64	0.783	202	1.6
		1034	4.5	18.1	6.66	0.765	200	1.6
		1056	-	18.4	6.52	0.800	231	2.0
MW-6	8/17/95	1512	0.0	19.6	6.43	0.929	83	1.0
		1514	1.5	17.3	6.45	0.787	60	0.9
		1516	3.0	17.3	6.48	0.849	47	1.1
		1519	3.5	17.5	6.55	0.841	42	1.1
		1522	5.0	17.5	6.57	0.833	41	1.1
MW-8	8/17/95	1131	0.0	20.5	6.64	0.695	187	1.1
		1132	1.0	19.0	6.69	0.649	186	2.1
		1134	2.5	18.1	6.70	0.650	186	2.5
		1136	3.5	18.0	6.72	0.646	186	3.4
		1139	4.0	18.0	6.92	0.656	183	5.4
		1358	-	18.8	6.69	0.753	233	4.3
MW-11	8/16/95	1710	0.0	20.1	7.00	0.575	183	3.7
. <u> </u>		1716	1.0	18.5	6.86	0.568	189	4.1
		1719	3.0	18.9	7.01	0.560	179	4.6
		1720	3.5	18.6	7.00	0.556	179	4.8

Table 2 Water Quality Data Site 1, Hazardous Waste Storage Area (560) Rickenbacker ANGB Page 2 of 3

Well ID	Date	Time	Volume (gallons) Purged	Temperature (Celsius)	pH (S.U.)	Conductivity (millimhos/cm)	ORP (mV)	Dissolved Oxygen (mg/L)
		1722	4.5	18.2	7.01	0.560	178	4.7
MW-12	r-12 8/16/95 1422		0.0	20.1	6.75	0.832	141	2.0
		1425	0.5	18.9	6.80	0.813	139	2.9
		1427	2.0	18.2	6.81	0.807	136	3.2
		1429	3.5	18.0	6.85	0.789	122	2.9
		1432	4.5	19.4	6.82	0.780	112	2.5
		1433	5.0	17.5	6.83	0.774	111	2.4
		1520	-	17.4	6.81	0.834	143	1.0
ESMP-2D	8/21/95	1515	0.0	17.3	6.74	0.726	-65	0.7
		1517	0.3	16.4	6.65	0.730	-68	0.5
		1519	0.5	16.2	6.79	0.728	-76	0.3
		1530	2.0	17.1	6.76	0.740	-59	0.6
ESMP-3D	8/21/95	1427	0.0	17.9	6.67	0.822	-48	1.3
		1430	0.25	16.2	6.57	0.819	-46	1.4
		1433	0.35	16.2	6.73	0.818	-48	1.3
		1445	1.5	16.4	6.78	0.798	-51	1.3
ESMP-4S	8/22/95	1248	0.15	21.8	6.58	-	67	2.9
		1250	0.25	21.5	6.51	-	73	2.5
		1307	1.0	23.6	6.58	-	83	0.8
ESMP-4D	8/21/95	1208	0.0	21.1	6.57	0.986	20	1.4
		1214	0.5	18.9	6.65	0.930	37	1.3
		1220	0.8	18.9	6.68	0.915	42	1.3
		1226	1.2	17.5	6.66	0.905	57	1.4
ESMP-6D	8/21/95	1040	0.0	20.9	6.73	0.812	10	1.5
		1045	0.2	19.2	6.70	0.831	-11	1.3
		1049	0.4	19.9	6.69	0.815	-14	1.0

Table 2 Water Quality Data Site 1, Hazardous Waste Storage Area (560) Rickenbacker ANGB Page 3 of 3

Well ID	Date	Time	Volume (gallons) Purged	Temperature (Celsius)	pH (S.U.)	Conductivity (millimhos/cm)	ORP (mV)	Dissolved Oxygen (mg/L)
		1115	2.0	19.8	6.70	0.839	-31	1.3
ESMP-8S	8/18/95	1012	0.5	25.3	7.54	0.578	156	6.7
ESMP-10S	8/18/95	1319	0.0	23.4	7.09	0.667	165	4.3
		1424	0.3	22.7	7.31	0.687	159	6.4
ESMP-13S	8/22/95	1104	0.0	18.0	6.95	-	-125	1.0
		1106	0.2	16.4	6.95	-	-122	0.7
		1116	1.0	17.2	6.96	-	-125	0.4
ESMP-14D	8/21/95	1643	0.0	17.0	6.74	0.042	-148	0.6
		1646	0.25	16.0	6.82	0.044	-147	0.1
		1648	0.4	15.4	6.78	0.045	-134	0.7
		1650	0.45	15.5	6.79	0.045	-126	0.6
		1659	2.5	16.1	6.73	0.046	-89	0.7
ESMP-16S	8/22/95	0930	0.0	19.6	6.78	-	-141	0.9
		0934	0.25	18.0	6.79	-	-146	0.5
		0941	1.75	19.1	6.82	-	-134	0.4
ESMP-16D	8/21/95	1602	0.0	19.1	6.69	0.043	-145	0.4
		1605	0.3	16.6	6.72	: 0.044	-153	0.3
		1606	0.4	16.6	6.74	0.044	-153	0.4
		1624	2.0	19.0	6.74	0.044	-126	0.8
ESMP-17S	8/22/95	0855	0.2	17.9	6.93	-	-85	0.4
		0859	0.4	17.6	6.96	-	-102	0.4
		0910	1.7	17.9	6.97	-	-113	0.5

Missing conductivity measurements were the result of a faulty conductivity probe, which was removed from use; no replacement was available.

Table 3 HACH Test Kit Results Site 1, Hazardous Waste Storage Area (560) Rickenbacker ANGB Page 1 of 1

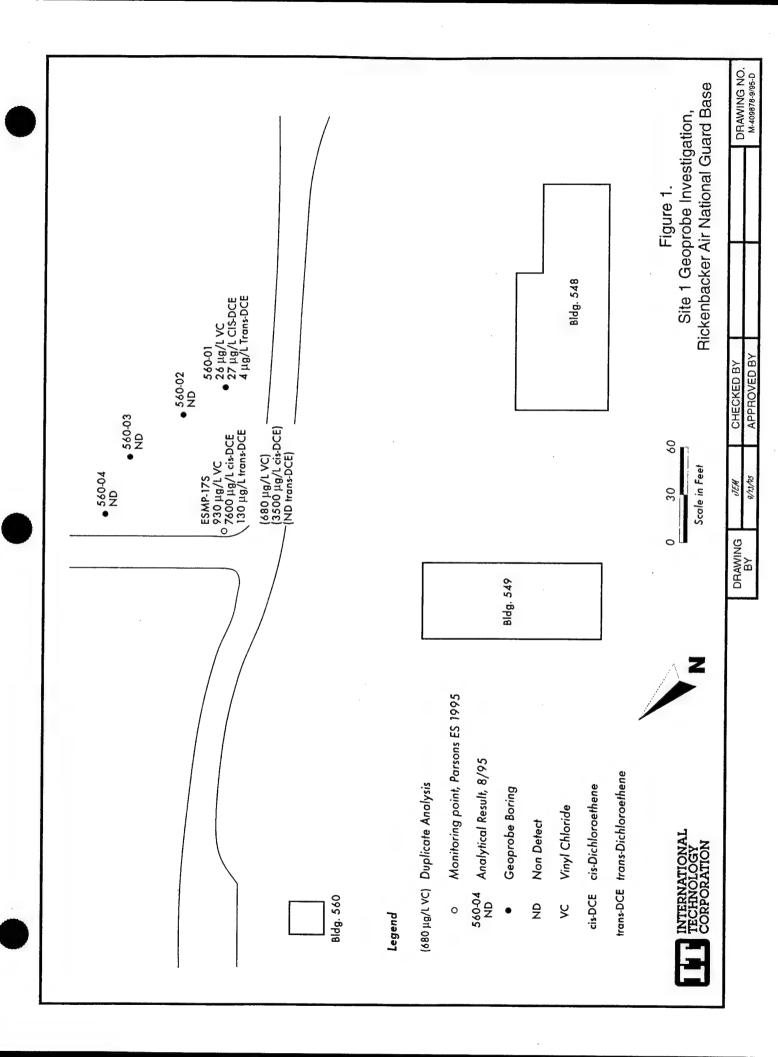
Well ID	Date	Ferrous Iron (mg/L)	Sulfide (mg/L)	Alkalinity (mg/L as CaCO ₃)	Comments
ESMP-2D	8/21/95	1.51	0.007	500	
ESMP-3D	8/21/95	1.26	0.006	480	
ESMP-4S	8/22/95	. -	0.648	520	Sample too turbid for accurate reading for ferrous iron
ESMP-4D	8/21/95	0.65	0.113	520/500	Original/Duplicate
ESMP-6D	8/21/95	1.32/1.33	0.006/0.005	540/480	Original/Duplicate
ESMP-8S	8/18/95	0.37	0.283	380	
ESMP-10S	8/18/95	-	0.421/0.386	400/400	Sample too turbid for ferrous iron. Original/Duplicate
ESMP-13S	8/22/95	2.72/2.84	0.011/0.009	460/460	Original/Duplicate
ESMP-14D	8/21/95	- 1.58	0.044	420	
ESMP-16S	8/22/95	2.94	0.012	560	2:1 dilution for ferrous iron
ESMP-16D	8/21/95	2.34	0.028	500	2:1 dilution for ferrous iron
ESMP-17S	8/22/95	2.69	0.017	460	
MW-3	8/17/95	0.37	0.078	-	
MW-4	8/17/95	0.07	0.051	-	
MW-5	8/17/95	2.62	0.365	-	
MW-6	8/17/95	0.43	0.270	-	·
MW-8	8/17/95	0.42	0.261	-	
MW-11	8/17/95	0.24	0.118		
MW-12	8/16/95	1.10	0.091	480	

Table 4 Analytical Parameters Site 1, Hazardous Waste Storage Area (560) Rickenbacker ANGB Page 1 of 1

Well	voc	svoc	Metals (unfiltered)	Metals (filtered)	Natural Attenuation Parameters ¹	Field Parameters ²
MW-4	Х	Х	Х	· x	Х	х
MW-8	х	х	Х	х	Х	х
MW-11	х	х	х	х	X	х
MW-12	х	х	Х	х	х	х
MW-3	х				X	х
MW-5	х				Х	х
MW-6	х				Х	х
ESMP-2D	х				. X	Х
ESMP-3D	Х				Х	Х
ESMP-4S	х				Х	х
ESMP-4D	х				х	X
ESMP-6D	х				Х	Х
ESMP-8S	х				Х	Х
ESMP-10S	х				Х	х
ESMP-13S	х				х	х
ESMP-14D	х				х	Х
ESMP-16S	х				х	Х
ESMP-16D	х				х	Х
ESMP-17S	х				X	Х

¹ Natural attenuation parameters included methane, ethene, ammonia, nitrogen, nitrite, nitrate, sulfate, alkalinity, and chloride.

² Field Parameters included pH, temperature, conductivity and the following natural attenuation parameters: ferrous iron, sulfide, alkalinity, dissolved oxygen, and oxidation reduction potential.





PROJECT NUMBER: 4098	PROJECT NAME: RANGB					
BORING NUMBER: 560 - 6			DATE: 08 - 15 - 95			
ELEVATION:	GWL: Depth Date/T	ime	DATE STARTED: 08- 15 95			
ENGINEER/GEOLOGIST: P. Y		ime	DATE COMPLETED: 08-15			
DRILLING METHODS: GEO PE	THODS: GEOPEOSE PAGE /					
	•		.			
SAMPLE TYPE & NO. BLOWS ON SAMPLER PER (NA) RECOVERY	DESCRIPTION	USCS SYMBOL MEASURED CONSISTENCY (TSF)	REMARKS REMARKS			
	0.0-6.0 10 yr 4/4 dack yellowight brown sicry clay, trace gravel, and simplicity. Very inff, damp 6.0-12.0 10 yr 4/, dark grzy claycy silt, trace medin hand, stiff damp		0.8	4 - 1 - 1 - 1 - 1 - 1 - 1 - 1 - 1 - 1 -		
	12.0 - 14.0 1048 4/1 carl gry sand i grand will graded, link fines, wise, with	 	0.8 SCREEN SE 12'-14' B6	S		
- 15 - - 15 - 	BOHON of Body 14.0					

NOTES:

GEODEOSE COMPANY: PH.LA ENVIRONMENTAL

OPERATOR: BARRY SOWER



PROJECT NUMBER: 409878	PROJECT NAME:	RANGB		
BORING NUMBER: 560-02	COORDINATES:		DATE: 08 -	
ELEVATION:	GWL: Depth	Date/Time		ED: 08-15-95
ENGINEER/GEOLOGIST: P. MCAZZEN	Depth	Date/Time	DATE COMPL	LETED: 08-15 -95
DRILLING METHODS: GEO PROSE			PAGE /	OF /
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F C B O N S C E C C C C C C C C C C C C C C C C C		MBO	- 25	DEMARKS

DHILLI	NG ME	HOUS.	GEOPE	205				
ОЕРТН (<i>A</i> 7 .)	SAMPLE TYPE & NO.	BLOWS ON SAMPLER PER (N/A)	RECOVERY 63	DESCRIPTION	USCS SYMBOL	MEASURED CONSISTENCY (TSF)	READWG	REMARKS
0				0.0 - 4.0 1048.414 darky ellans H bean			0.8	
-				bilty clay, reacc gravel, medium plasming				·
-				very wift, amp				
-								
-								
-2-				a so while her are classed				_
-6-				6.0-11.5 10 yr 4/1 dael gray clayey on Ly reace midul sand, soit. comp			0.8	
-				of the second said, street, of				-
-								-
-10-								-
L _				-11-5 - 14.0 104R4/, carl grey madigra-1			-	
11.5				well graded, losse, wet			0.8	SCREEN SET -
L -								12-14-865
-14 -								
-15-				BOTTOM of BORING 14.0		1		1
-								1
-		1						
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-20-								
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30	<u> </u>				<u> </u>		1	<u> </u>
NOTE	: C·							

NOTES:

GEODROSE COMPANY: PHILD ENVIRONMENTAL

OPERATOR: BARRY SOWER



PROJECT NUMBER: 409878	PROJECT NAME: RANGB			·
BORING NUMBER: 560-03	COORDINATES:			3-16-95
ELEVATION:	GWL: Depth Date/Tir	me		ARTED: 08-16 95
ENGINEER/GEOLOGIST: P. MCCAZ	Depth Date/Tit	me	DATE CO	OMPLETED: 03-16 -95
DRILLING METHODS: GEO PEOSE			PAGE	/ OF /
SAMPLE TYPE & NO. BLOWS ON SAMPLER PER (NA) RECOVERY	DESCRIPTION	USCS SYMBOL MEASURED CONSISTENCY (TSF)	READING (pom)	REMARKS
Acou	- 8.0 love 4ly darkyellows H w =. love clay, react grand, anden uny, very anff, canj		0.6	- - - -
54md Coos	10.0 1048 41, deck grey ¿ gravel, well graded, traces, et		0.4	SCREEN SET - 8'-10' B63 -
	, Haed, out		ن. به	- -
	Bonn of Bong 14.0			- - - -
				-
				- - -

JO NOTES:

GEODZOSE COMPANY: PH.LA ENVIRONMENTAL

OPERATOR: BARRY SOWER



PROJECT NUMBER: 409%	78	PROJECT NAME:	RANGB						
BORING NUMBER: 540-		COORDINATES:				DATE:	08-1	4 -95	
ELEVATION:	GWL: Depth	Date/Tin	ne		DATE	DATE STARTED: 08-16 -95			
ENGINEER/GEOLOGIST: P. 1	MCAZZEN	Depth	Date/Tin	ne		DATE	COMPLE	TED: 08-	16-95
DRILLING METHODS: GEO PE		. <u> </u>				PAGE	/	OF	1
Dilleting in Charles									
DEPTH (A-) SAMPLE TYPE & NO. BLOWS ON SAMPLER PER (AA) RECOVERY		DESCRIPTION			MEASURED CONSISTENCY (TSF)	READWG:		REMARKS	
	Made play 12:0-14.0 And Geneliu	10yR 4/1 dack gr 10yR 4/1 dack gr 10yR 4/1 dack gr 10yR 4/1 dack gr 11 spaced, less	of med			ර ග		ceen c'-14	sev -

GEODZOSE COMPANY: PHILA ENVIRONMENTAL

OPEZATOR: BARRY SOWER